



PROGRAM MANAGER FOR ROCKY MOUNTAIN ARSENAL

U.S. ARMY
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— COMMITTED TO PROTECTION OF THE ENVIRONMENT —

FINAL
COMPREHENSIVE AIR QUALITY AND
METEOROLOGICAL MONITORING PROGRAM
AIR QUALITY DATA ASSESSMENT REPORT
FOR FY 1993
VOLUME II of IV

VERSION 2.0

OCTOBER, 1994
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Prepared by:

EBASCO SERVICES INCORPORATED
AGEISS ENVIRONMENTAL INCORPORATED

Prepared for:

U.S. Army Program Manager's Office for
Rocky Mountain Arsenal

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LIST OF ACRONYMS AND ABBREVIATIONS

| | |
|---------------------------------|--|
| AA | Atomic Absorption |
| ACGIH | American Council of Governmental Industrial Hygienists |
| ADI | Acceptable Daily Intake |
| Army | U.S. Army |
| As | Arsenic |
| Atrazine | 2-chloro-4-ethylamino-6-isopropylamino-s-trianine |
| BCHPD | Bicycloheptadiene |
| BLDG | Building |
| °C | Degrees Celsius |
| C ₆ H ₆ | Benzene |
| CAQMMP | Comprehensive Air Quality and Meteorological Monitoring Program |
| CCl ₄ | Carbon Tetrachloride |
| ccm | cubic centimeters per minute |
| CDH | Colorado Department of Health |
| cfm | cubic feet per minute |
| CFR | Code of Federal Regulation |
| CH ₂ Cl ₂ | Methylene Chloride |
| CHCl ₃ | Chloroform |
| Chlordane | 1,2,4,5,6,7,8,8-Octachloro-2,3,3a,4,7,7a-hexahydro-4,7-methano-1H-indene |
| CiC ₆ H ₅ | Chlorobenzene |
| CMP | Comprehensive Monitoring Program |
| CMP FY90 | Comprehensive Monitoring Program Fiscal Year 1990 |
| CO | Carbon Monoxide |
| CRL | Certified Reporting Limit |
| CVAAS | Cold Vapor Atomic Absorption Spectroscopy |
| DBCP | Dibromochloropropane |
| 11DCLE | 1,1-Dichloroethane |
| 12DCLE | 1,2-Dichloroethane |
| 12DCE | trans-1,2-Dichloroethene |
| DCPD | Dicyclopentadiene |
| DIMP | Diisopropylmethyl phosphonate |
| 12DMB | 1,2-Dimethylbenzene |
| DMDS | Dimethyldisulfide |
| DMMP | Dimethylmethylphosphonate |
| DPS | Denver Public Schools |
| EA | Endangerment Assessment |
| EBASCO | Ebasco Services Incorporated |
| EPA | Environmental Protection Agency |
| ESH | Effective Stack Height |
| ETC ₆ H ₅ | Ethylbenzene |

LIST OF ACRONYMS AND ABBREVIATIONS (Cont.)

| | |
|---------------------------------|--|
| °F | Degrees Fahrenheit |
| FID | Flame Ionization Detector |
| FS | Feasibility Study |
| FY | Fiscal Year |
| GC | Gas Chromatograph |
| GC/MS | Gas Chromatography/Mass Spectrometry |
| GC/ECD | Gas Chromatography/Electron Capture Detection |
| GT | Greater Than |
| H ₂ S | Hydrogen Sulfide |
| HEAST | Health Effects Assessment Summary Table |
| Hg | Mercury |
| ICAP | Inductively Coupled Argon Plasma |
| IRA | Interim Reponse Action |
| IRA-F | Interim Response Action at Basin F |
| IRDMIS | Installation Restoration Data Management Information System |
| IRIS | Integrated Risk Information System |
| ISC | Industrial Source Complex Dispersion Model |
| lpm | liters per minute |
| m | meter |
| Malathion | S-[1,2-bis(ethoxycarbonyl)ethyl]0,0-dimethyl-phosphorodithioate |
| MEC ₆ H ₅ | Toluene |
| µg | micrograms |
| µg/std m ³ | micrograms per standardized cubic meter of air (760 mm Hg, 25°C) |
| µg/m ³ | micrograms per cubic meter |
| µm | micrometer (micron) |
| mg/m ³ | milligrams per cubic meter |
| MIBK | Methyl Isobutyl Ketone |
| mph | miles per hour |
| MRI | Midwest Research Institute |
| MST | Mountain Standard Time |
| NAAQS | National Ambient Air Quality Standards |
| NAD | North American Datym |
| NATICH | National Air Toxics Information Clearinghouse |
| ng/m ³ | nanograms per cubic meter |
| NH ₃ | Ammonia |
| NIOSH | National Institute of Occupational Safety and Health |
| NIST | Nation Institute of Standards and Technology |
| NMOC | Non Methane Organic Compound |
| NNDMEA | N-Nitrosodimethylamine |
| NO | Nitric Oxide |
| NO ₂ | Nitrogen Dioxide |
| NO _x | Nitrogen Oxides |

LIST OF ACRONYMS AND ABBREVIATIONS (Cont.)

| | |
|-----------------|---|
| O ₃ | Ozone |
| OCP | Organochlorine Pesticides |
| OVA | Organic Vapor Analyzer |
| OVM | Organic Vapor Meter |
| Parathion | 0,0-Diethyl-0(p-nitrophenyl)phosphorothioate |
| PID | Photoionization Detector |
| PM-10 | Particulates less than 10 micrometers |
| PMRMA | Program Manager Rocky Mountain Arsenal |
| ppb | parts per billion |
| ppbr | parts per billion volume |
| PPDDE | 2,2-Bis(4-chlorophenyl)-1,1-dichloroethylene |
| PPDDT | 1,1-Bis(4-chlorophenyl)-2,2,2-trichloroethane |
| ppm | parts per million |
| PSD | Prevention of Significant Deterioration |
| PUF | Polyurethane Foam |
| QA | Quality Assurance |
| QC | Quality Control |
| RAP | Remedial Action Program |
| RBACs | Risk-Based Air Concentrations |
| RfCs | Reference Concentrations |
| RI | Remedial Investigation |
| RI/FS | Remedial Investigation/Feasibility Study |
| RMA | Rocky Mountain Arsenal |
| SARA | Superfund Amendments and Reauthorization Act |
| sccm | standard cubic centimeters per minute (760 mm Hg, 25°C) |
| scfm | standard cubic feet per minute (760 mm Hg, 25°C) |
| SO ₂ | Sulfur Dioxide |
| SOP | Standard Operating Procedures |
| SQI | Submerged Quench Incinerator |
| Supona | 2-chloro-1-(2,4-dichlorophenyl)vinyl diethyl phosphate |
| SVE | Soil Vapor Extraction |
| SVOC | Semivolatile Organic Compounds |
| 111TCE | 1,1,1-Trichloroethane |
| 112TCE | 1,1,2-Trichloroethane |
| TCLEE | Tetrachloroethene |
| TIC | Tentatively Identified Compound |
| TLV | Threshold Limit Value |
| tpy | tons per year |
| TRCLE | Trichloroethene |
| TSP | Total Suspended Particulates |
| UATMP | Urban Air Toxic Monitoring Program |
| UNK | Unknown Number |

LIST OF ACRONYMS AND ABBREVIATIONS (Cont.)

| | |
|----------|--|
| USAEHA | U.S. Army Environmental Hygiene Agency |
| USATHAMA | U.S. Army Toxic and Hazardous Materials Agency |
| UTM | Universal Transverse Mercator |
| VOC | Volatile Organic Compounds |
| VOTA | Volatile Organic Toxic Air collection system (Graseby-Andersen, Inc., trade name) |
| XAD | Adsorbent resin for semivolatile compounds (Supelco, Inc., trade name) |
| XYLENE | Total Xylenes |
| % | Percent |

4.0 RESULTS OF FY93 PROGRAM

4.1 BASIS OF AIR QUALITY DATA EVALUATION

The purpose of the CAQMMP is to maintain a baseline database to verify ambient air quality at RMA and to evaluate progress made in current and future remedial actions. The program has several related key objectives:

- Evaluate potential air quality health hazards that may exist within the RMA boundaries
- Confirm progress made to date in removing potential air contaminants resulting from previous activities
- Provide standardized procedures of data collection to measure impacts of ongoing remedial activities
- Describe potential impacts of other sources of air pollutants

The data monitoring, verification, and evaluation programs have been designed to achieve these objectives. Additionally, the data analyses have been related to the source of contaminants detected. Source attribution requires consideration of a number of variables unique to air quality assessment:

- Source- and receptor-specific meteorological conditions including wind direction, wind speed, atmospheric stability, mixing height, temperature, and precipitation
- RMA land disturbances, in particular, remedial construction and excavation activities
- Source environment, including topography, soil types, and vegetation
- Ambient air quality concentration levels directly upwind from the RMA area, including metropolitan Denver

In summary, measured background levels of TSP, PM-10, VOCs, SVOCs, OCPs, metals, arsenic, mercury, and asbestos within RMA cannot be evaluated in terms of the stated objectives without identifying causative and contributing factors including off-post sources. Thus, a computer database of ambient air quality concentrations, meteorological data, and significant influencing parameters (e.g., local air contaminant sources) has been established and maintained during the CAQMMP. As the database continues to expand over several years, the statistical significance and additional applications of the database may increase. For example, the progressive effect of previous remedial activities should indicate decreasing concentration levels in the vicinity of

RMA. If certain contaminant levels are above normal background values, a persistence of a given element and/or an indication of the ineffectiveness of remedial actions would be confirmed. Additional or alternate mitigating actions may be required.

During FY93, a full year of data were collected in the vicinity of Basin F under the CAQMMP. In this report, these data are compared with data collected at identical monitoring sites under similar meteorological and climatological conditions during the Basin F remediation period and subsequent post-remedial periods. The basis of this evaluation is discussed further in Section 4.1.2.

4.1.1 Computerized Documentation

Computerized documentation and analysis have provided the following information for this Air Quality Data Assessment Report:

- For each sampling station, a list is provided that includes the compounds detected at that station, the range of concentrations reported, the maximum concentration, pertinent weather conditions, and other factors such as mitigation through remedial activity. High-event monitoring activity was similarly documented and reported.
- Comparison of measured TSP, PM-10, VOCs, SVOCs, OCPs, metals, arsenic, mercury, and asbestos is provided with available regional data, regulatory guidelines, and other toxic guidelines as appropriate. Comparisons with standards and guidelines are discussed further in this report for each specific group of analytes.
- Comparison is provided of significant (high or low) contaminant levels under similar conditions to indicate remedial progress. For example, VOC, SVOC, and OCP values at an RMA site over a particular season could result from remedial activity, transport from off post or specific meteorological conditions. Similarly, high metals levels could result from strong wind speeds from a persistent direction, transport from the metropolitan area, significant RMA excavation activities, or all of the above concurrently. The database

compiled over the life of the CAQMMP describes these relationships for this and future reports.

- Identification is provided of meteorological conditions, excavation activities, or discrete sources and influences that may trigger high levels of contaminant activity and require special precautions and mitigating actions.

In addition, this CAQMMP Air Quality Data Assessment Report provides, for the purpose of remediation assessment, the results of the Interim Action Basin F Cleanup Program from the initiation of remedial activities at Basin F in 1988 through September 30, 1993. Evaluation of RMA data includes the use of the following strategies:

- Incorporate all remedial activity monitoring data into the CAQMMP and assess these data as a subset of the overall database
- Use specific standardized guidelines and criteria for air monitoring support of excavation and remedial activities at RMA. Criteria have been based on available literature, existing state-of-the-art techniques, and direct FY88, FY89, FY90, FY91, FY92 and FY93 experience with air quality monitoring and contaminants of concern at RMA. Guidelines include specific instrumentation, monitoring procedures and techniques, real-time prediction and alert procedures, and models relating to Health and Safety. For example, during FY93 the CAQMMP staff and PMRMA drew upon this experience for planning an air quality program in support of the forthcoming South Plants Pilot Demolition Program.

4.1.2 Basin F Remediation Evaluation

During the CMP FY88 and FY89 periods, Basin F remediation proceeded from March 22, 1988 to May 4, 1989. The post-remediation period, starting May 5, 1989, has continued through the conclusion of CAQMMP FY93. The Basin F Remediation Monitoring Program, the follow-on IRA-F data, and the CMP/CAQMMP programs have provided results of potential air contaminant levels prior to, during, and subsequent to the remediation program. These data reflect

remediation progress. The results of this assessment are provided in this report under the phases and stages shown in Table 4.1-1. All data shown in tables and figures for FY93 cover the period from October 1, 1992, through September 30, 1993, and have been designated Phase 7.

4.1.3 Dispersion Model Application

The CMP/Basin F Air Quality Modeling Programs included the development and/or application of several standard and special purpose models for assessing potential contaminant impacts, including the Industrial Source Complex (ISC) Model and a PUFF advection model (USEPA 1986b). Dispersion modeling assisted appreciably in addressing impacts from potential emission sources, both on and off post. During FY93, dispersion modeling was used to evaluate CAQMMP-monitored data at Basin F, the South Plants, Basin A, the SQI, and other special contingency monitoring events. Both the ISC and PUFF Model were also used effectively in the Basin F cleanup program to assess real-time contamination levels during remedial activities and to provide a forewarning of potential hazardous conditions.

The FY93 Report provides a joint analysis of CAQMMP and Basin F data over a 67-month period to evaluate remedial impacts and post-remedial progress. Dispersion models were used to identify potential source impacts and overriding meteorological influences. One objective was to confirm that concentrations measured under the remedial and post-remedial phase periods occurred during typical meteorological conditions. Figures 4.1-1 and 4.1-2 show mean dispersion patterns for the Phase 1 remediation period and the most recent Phase 7 (FY93) post-remediation period for an area source centered at Basin F. These patterns are closely similar and suggest that anomalous meteorological conditions did not influence results. Similar annual dispersion comparisons have been made in previous post-remedial annual assessments (WCC 1993).

The dispersion patterns are depicted by relating ambient concentration (X) to source strength (Q), incorporating site-specific meteorological factors without source measurements. The X/Q contours were obtained from the EPA ISC Model (using Basin F or other potential emissions sources as a hypothetical area source); they reflect the influences of wind speed, wind direction,

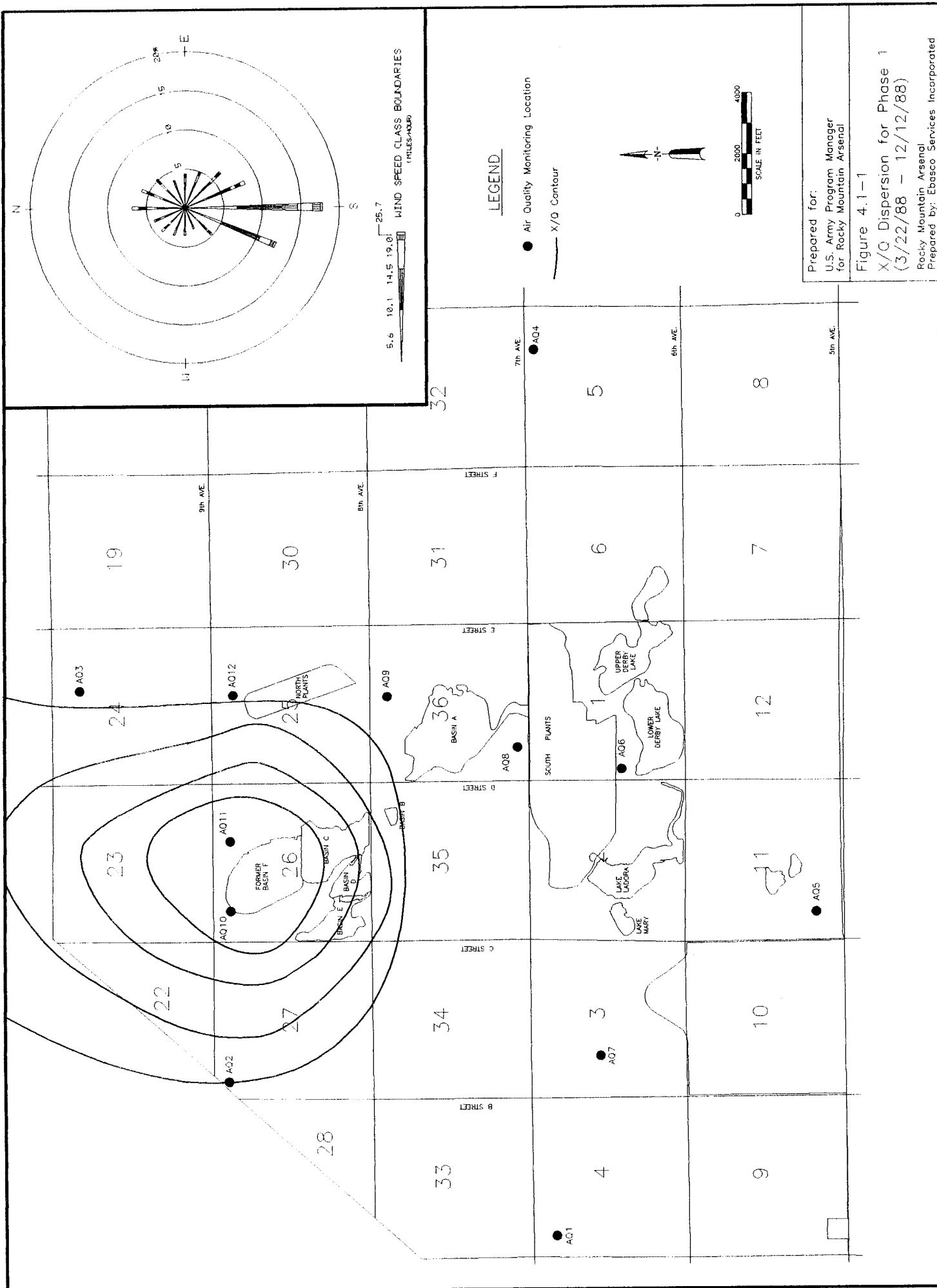
Table 4.1-1 Basin F Remediation Phases

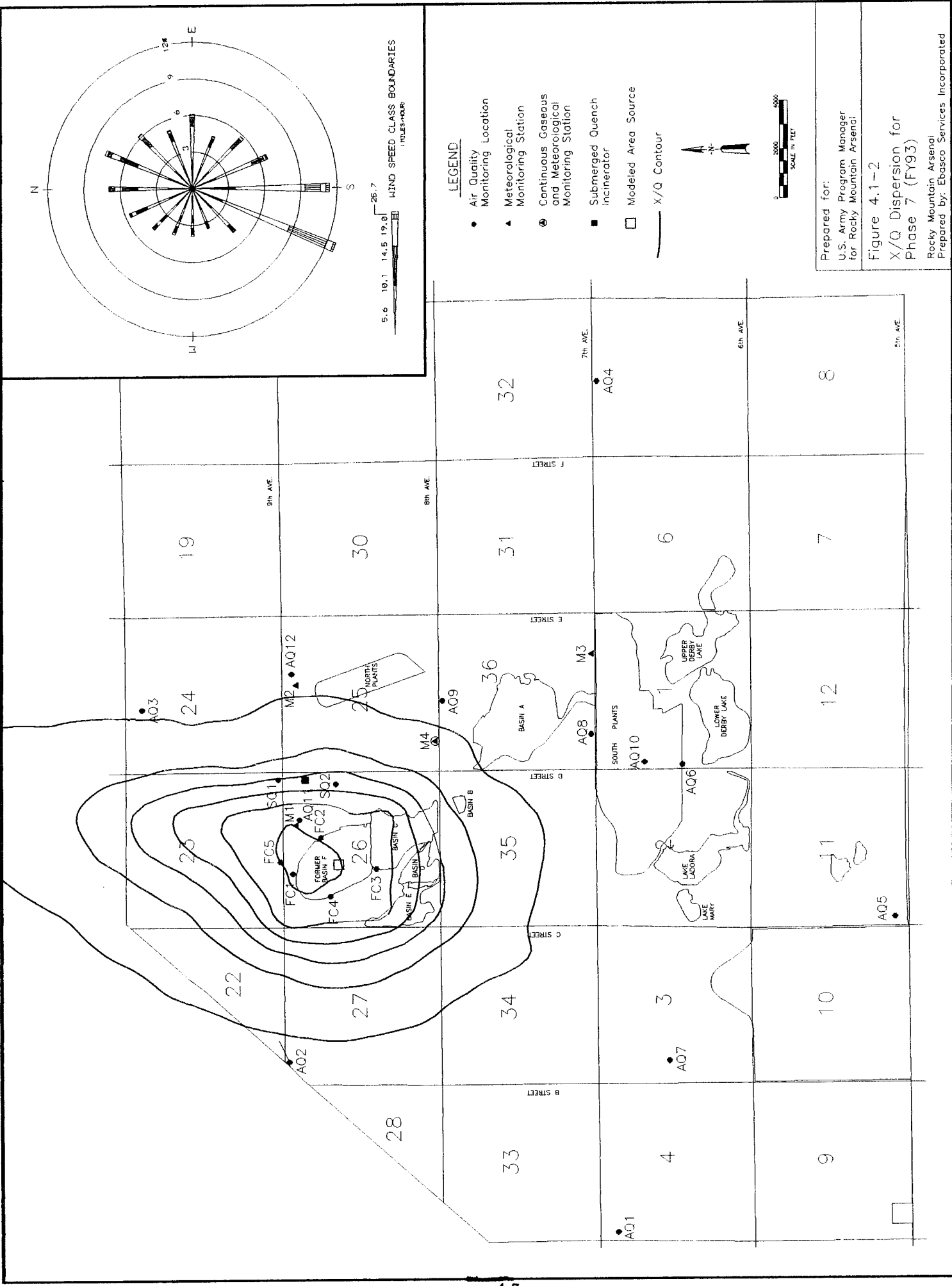
Page 1 of 1

| Phase | Dates | Activity | Monitoring Data Available |
|-------|---------------------------------|---|---|
| 1 | 3/22/88 - 12/12/88 | Basin F clean-up. Stockpiling of clay for future capping of basin. | CMP FY88, CMP FY89, Basin F Remediation Monitoring Program, RIFS Odor Program |
| 2 | 12/13/88 - 2/15/89 (Stage 1) | Capping of basin with clay. | CMP FY89, Basin F Remediation Monitoring Program, RIFS Odor Program |
| | 2/16/89 - 6/6/89 (Stage 2) | Capping of basin with topsoil. Grading and reseedling of basin and surrounding area. Liner installed over holding pond. | |
| 3 | 5/6/89 - 9/30/89 | Post-remedial period: Pumping of accumulated waste pile liquids to holding pond. | CMP FY89, IRA-F FY89 |
| 4 | 10/1/89 - 9/30/90 | Post-remedial period | CMP FY90, IRA-F FY90 |
| 5 | 10/1/90 - 9/30/91 | Post-remedial period | CMP* FY91 |
| 6 | 10/1/91 - 9/30/92 | Post-remedial period | CMP* FY92 |
| 7 | 10/1/92 - 9/30/93 | Post-remedial period | CAQMMP* FY93 |

* CMP and CAQMMP are integrations of the CMP and IRA-F monitoring programs designated during the previous monitoring fiscal years.

CMP - Comprehensive Monitoring Program
 CAQMMP - Comprehensive Air Quality and Meteorological Monitoring Program
 IRA-F - Interim Response Action at Basin F
 RIFS - Remedial Investigative/Feasibility Study





Prepared for:
 U.S. Army Program Manager
 for Rocky Mountain Arsenal

Figure 4.1-2
 X/Q Dispersion for
 Phase 7 (FY93)

Rocky Mountain Arsenal
 Prepared by: Ebasco Services Incorporated

atmospheric stability, and inversion conditions on the spread of pollutants for the monitoring period. The X/Q values do not indicate actual ambient concentration, but indicate relative strength, or potential for concentration levels, based on meteorological conditions and a unitized (1 gram per second) source strength (see Section 6.5 for a further description of CAQMMP-employed dispersion models). During remediation, higher measured ambient concentration levels corresponding to high X/Q values were found close to the Basin F source and directly downwind from prevailing winds. The dispersion patterns were also skewed to the north of Basin F, reflecting the prevailing air flow during the monitoring periods, as shown in the wind rose insert in Figures 4.1-1 and 4.1-2. Thus, higher ambient concentrations as a result of potential Basin F source contaminants during remediation activities were confirmed by modeling and suggested the model's further application for prediction and extrapolation.

As stated, the dispersion patterns were closely similar for each phase of the remediation and post-remediation periods. The implication, therefore, is that the dispersion characteristics defined by the measured meteorology were similar during the 7 phases. Therefore, any significant variations in measuring air quality data are a function of source emissions. For example, high pesticide levels were noted during the Phase 1 period. During Phase 2, Phase 3, Phase 4, Phase 5, Phase 6 and Phase 7, these contaminant levels decreased significantly (under similar dispersion conditions), implying that the potential Basin F emissions sources for these compounds were effectively contained at the conclusion of the Phase 1 period.

The X/Q approach, which has been effective in evaluating Basin F remediation progress, will be employed for the assessment of other potential RMA sources such as the South Plants, Basin A, and the SQI facility.

4.1.4 Source Emission Factors

One objective of the air quality assessment is to identify emission sources that contribute to ambient air levels measured by the RMA monitoring program. The CMP demonstrated that Basin F was a potential source of several VOCs, SVOCs, and metals compounds associated with

the remedial programs during FY88. High TSP levels were also evident at Basin F as a result of intense remedial construction activity. Significant decreases in ambient levels for most of these compounds occurred during Phases 3, 4, 5, 6 and 7 post-remediation periods, reflecting decreased emissions from Basin F and the associated remedial program.

As Basin F emissions decrease, other RMA sources could come into focus, such as Basin A, the South Plants, the SQI facility, and localized construction activities. These sources, as of the FY93 data collection period, appear to be at lower levels of intensity than metropolitan Denver, and in most cases their impacts are comparable to or less than VOC, OCP, and metals levels measured there.

It is important to note that CAQMMP monitoring results are difficult to evaluate for most pollutants measured during FY93 without examining the potential emission sources across the Denver urban area. A number of mobile and stationary sources in metropolitan Denver may impact CAQMMP air quality monitoring sites, depending upon wind direction, inversion conditions, temperature, and other meteorological factors influencing the Denver area. This report further identifies these external conditions and sources. Table 4.1-2 and Figure 4.1-3 show selected criteria pollutant sources in metropolitan Denver that may influence RMA air quality under certain meteorological conditions. TSP point sources are further identified in Table 4.2-8 and Figure 4.2-4; VOC point sources are further identified in Tables 4.6-8, and 4.6-9 and Figure 4.6-13 in Section 4.6.6 of this report.

4.2 TOTAL SUSPENDED PARTICULATES

4.2.1 CAQMMP FY93 TSP Results

As discussed in Section 2.1.1.1, TSP regulatory standards are in the process of review and revision by CDH. Nevertheless, TSP data collected at RMA over the past 6 years have been a strong indicator of remedial activity and progress; as such, TSP data continue to be a significant tool for evaluating important assessment criteria pertaining to air quality. A summary of data sampling frequency and recovery for the FY93 monitoring program for TSP at each of the

Table 4.1-2 Emission Inventory Summary for Regulated
Pollutants (tons per year)

Page 1 of 2

| Facility | Map # | Total Emissions | | | | |
|---|-------|------------------|------------------|------------------------------|------------------------------|------------------|
| | | VOC ¹ | TSP ² | SO ₂ ³ | NO ₂ ⁴ | CO ⁵ |
| Chemical Systems Technology, Inc | 1 | 6655 | | | | |
| Colorado Refining Co. / Total Petroleum | 2 | 989 | 174 | 632 | 288 | 141 ⁶ |
| Conoco Inc. - Denver Refinery | 3 | 740 | 465 | 2336 | 626 | 272 ⁶ |
| Denver Industrial Sales & Service | 4 | 257 | | | | |
| WYCO Pipe Line Co. DuPont Terminal | 5 | 236 | | | | |
| Pillow Kingdom MFG | 6 | 213 | | | | |
| Swedish Medical Center | 7 | 155 | | | | 2 |
| Union Chemicals Division | 8 | 143 | | | | |
| Chase Terminal Co. | 9 | 121 | | | | |
| Mastercraft | 10 | 120 | | | | |
| Kiewit Western | 11 | | 1830 | 16 | | |
| Denver Metals | 12 | | 336 | | | |
| Public Service Co. - Cherokee | 13 | 72 | 254 | 13362 | 14295 | 550 ⁶ |
| Public Service Co. - Arapahoe | 14 | 17 | 94 | 4567 | 5143 | 160 ⁶ |
| Waste Management of CO - Arapahoe | 15 | | 72 | | | |
| Cooley Gravel Co. 88th & Riverdale | 16 | | 70 | | | |
| Purina Mills Inc. - Henderson Mill | 17 | | 66 | | | |
| Centennial Materials - Thornburg | 18 | | 65 | | | |
| Cooley Gravel | 19 | | 64 | | | |
| BFI - Tower Landfill | 20 | | 58 | 1 | | 3 |
| Metro Wastewater Reclamation | 21 | 25 | | 134 | | 84 |
| Buckley Air National Guard Base | 22 | 17 | 27 | 37 | 263 | 56 |
| Owens Corning Fiberglass - Trumbull | 23 | | 27 | 29 | | 42 |
| General Chemical Corporation | 24 | | | 28 | | |
| Gates Rubber Company | 25 | 118 | | 28 | 245 | |
| Bar S Foods Company | 26 | | | 25 | | |
| Fast Construction Company | 27 | | 32 | 25 | | |
| Amoco Production Co. - Wattenberg | 28 | 116 | | 24 | | |
| Asarco Inc. - Globe Plant | 29 | | | 18 | | |
| Brannan S & G | 30 | 17 | 29 | 16 | | |

Table 4.1-2 Emission Inventory Summary for Regulated
Pollutants (tons per year)

Page 2 of 2

| Facility | Map # | Total Emissions | | | | |
|------------------------------------|-------|------------------|------------------|------------------------------|------------------------------|-----------------|
| | | VOC ¹ | TSP ² | SO ₂ ³ | NO ₂ ⁴ | CO ⁵ |
| Colo Interstate Gas Co. - Watkins | 31 | 55 | | | 866 | 335 |
| Amoco Production - Wattenberg | 32 | 116 | | 24 | 549 | |
| Koch Hydrocarbon Co. - Third Creek | 33 | 18 | | | 504 | |
| Public Service Co. - Delganey | 34 | | | | 313 | 23 |
| Koch Hydrocarbon Co. - Mitchell | | | | | 176 | 22 |
| Koch Hydrocarbon Co. - Radar Plant | | | | | 153 | 19 |
| Koch Hydrocarbon Co. - Dagoon | | | | | 146 | 19 |
| Public Service Co. - Zuni | 35 | | | 2 | 146 | 11 |
| Colo. Interstate Gas Co. - Latigo | | | | | 126 | 16 |
| Koch Hydrocarbon Co. - Boxelder | 36 | | | | 115 | 15 |
| Explosive Fabrics | | | | | | 17 |
| Presbyterian/St. Lukes Healthcare | 37 | | | | | 74 |
| Littleton/Englewood WWTP | 38 | | | | | 11 |
| Koch Hydrocarbon Co. - Rattlesnake | | | | | | 10 |
| ITT Continental Baking Co. | 39 | | | | | 9 |
| Schafer Commercial Seating Inc. | 40 | | | | | 9 |
| Panhandle Easter Pipeline Co. | 41 | | | | | 9 |
| Lowry Air Force Base | 42 | 45 | | | | 9 |
| Koch Hydrocarbon Co. - Antelope | | | | | | 9 |
| Irondale Gas Processing Co. | | | | | | 8 |
| Rocky Mountain Arsenal | 42 | 1.1 | 1.1 | 6 | 32.6 | 7 |

Source: Colorado Department of Health EISPS Inventory, 1993.

¹ Blank entry indicates total annual VOC emissions of less than 15 tons per year or did not report VOC emissions.

² Blank entry indicates total annual TSP emissions of less than 25 tons per year or did not report TSP emissions.

³ Blank entry indicates total annual SO₂ emissions of less than 1 ton per year or did not report SO₂ emissions.

⁴ Blank entry indicates total annual NO₂ emissions of less than 5 tons per year or did not report NO₂ emissions.

⁵ Blank entry indicates total annual CO emissions of less than 1 ton per year or did not report CO emissions.

⁶ CO emissions based on FY92 data report.

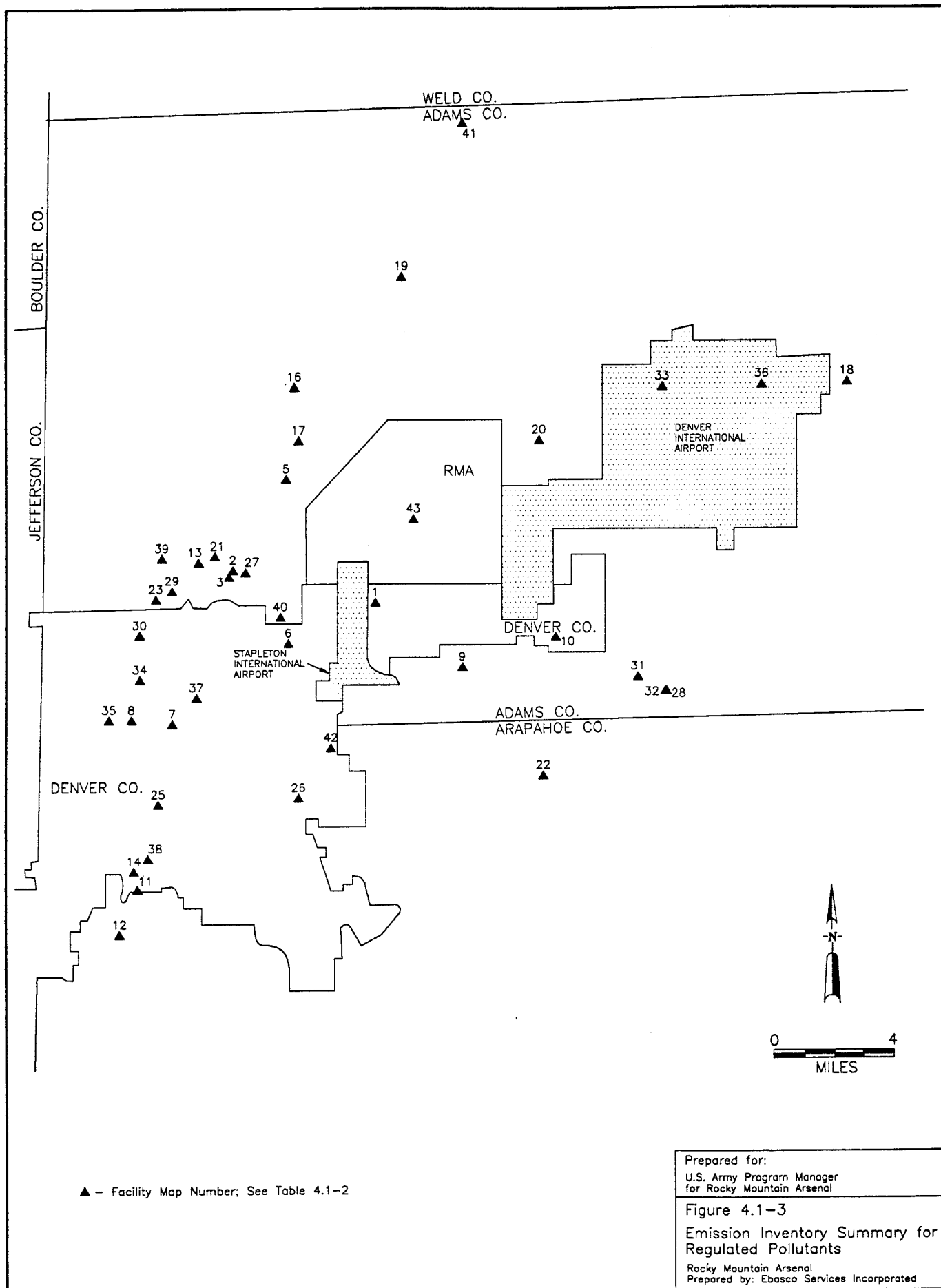
VOC - Volatile Organic Compounds

TSP - Total Suspended Particulates

SO₂ - Sulfur Dioxide

NO₂ - Nitrogen Dioxide

CO - Carbon Monoxide



monitoring locations is presented in Table 4.2-1. Recoveries are based on the total number of scheduled days during the monitoring year. Samples were considered not valid if there were equipment malfunctions or the sample filter was damaged. According to PSD guidelines, a minimum of 23 hours of sampling was required for a valid sample.

At the beginning of the FY93 CAQMMP, problems not previously experienced were observed with the TSP gravimetric analyses including filter blanks showing negative weights or significant high values, PM-10 values greater than TSP values at the same sampling location, and duplicate (collocated) samples showing significant variations. These difficulties were principally attributed to excessive and uncontrolled static electricity on the filters, so more stringent controls and quality assurance procedures were implemented, including use of a controlled weighing environment, use of anti-static devices, and increased replication of pre- and post-sampling filters. Significant improvement was achieved after January 31, 1993. Consequently, only the TSP and PM-10 results subsequent to this date are included in the summary analyses. All TSP sample analysis results obtained in FY93 (October 1992 to September 30, 1993) are provided in Appendix A. Much of the data prior to February 1, 1993 may be reliable and can be considered qualitatively on a case-by-case basis; however, because of the general concerns discussed above they are excluded from the fiscal year summaries. The last column in Table 4.2-1 reflects the percentage of FY93 TSP data considered in the TSP summary analyses.

TSP data are often reported by using the annual geometric mean values. This approach is based on early EPA guidance and is applied because the TSP monitoring data can be expected to fit a log-normal distribution. Under a log-normal distribution, there are a relatively large number of low concentrations and a small number of high concentrations. An arithmetic mean value would be greatly affected by the few very large values, while a geometric mean is much less affected by these extremes.

An arithmetic mean can be characterized as follows:

Table 4.2-1 Summary of Total Suspended Particulate Monitoring for FY93

Page 1 of 1

| Station | Samples Scheduled | Samples Collected | Valid Samples | Field % Recovery | Overall % Recovery |
|-----------|----------------------|----------------------|------------------|---------------------|-----------------------|
| AQ1 | 61 | 59 | 37 | 97 | 61 |
| AQ2 | 61 | 58 | 35 | 95 | 57 |
| AQ3 | 61 | 59 | 37 | 97 | 61 |
| AQ4 | 61 | 52 | 32 | 85 | 52 |
| AQ5 | 61 | 60 | 35 | 98 | 57 |
| AQ6 | 61 | 60 | 34 | 98 | 56 |
| AQ7 | 61 | 60 | 39 | 98 | 64 |
| AQ8 | 61 | 59 | 38 | 97 | 62 |
| AQ9 | 61 | 59 | 35 | 97 | 57 |
| AQ10 | 61 | 60 | 40 | 98 | 66 |
| AQ11 | 61 | 60 | 39 | 98 | 64 |
| AQ12 | 61 | 60 | 40 | 98 | 66 |
| FC1 | 31 | 31 | 22 | 100 | 71 |
| FC2 | 31 | 30 | 21 | 97 | 68 |
| FC3 | 12 | 13 | 9 | 108 | 75 |
| FC4 | 12 | 13 | 9 | 108 | 75 |
| FC5 | 31 | 30 | 21 | 97 | 68 |
| SQ1 | 61 | 60 | 36 | 98 | 59 |
| SQ2 | 61 | 59 | 38 | 97 | 62 |
| Duplicate | 61 | 59 | 35 | 97 | 57 |
| Overall | 1,032 | 1,001 | 632 | 97 | 61 |

% Percent

$$A = \frac{(a_1 + a_2 + a_3 + \dots + a_n)}{n}$$

While a geometric mean is the nth root of the product of the n observations:

$$G = (a_1 * a_2 * a_3 \dots a_n)^{1/n}$$

Where

a_n is the nth observation (of TSP data);

n is the total number of valid observations;

A is the arithmetic mean of the sample; and

G is the geometric mean of the sample.

Monthly and annual TSP results for FY93 for each monitoring station are summarized in Tables 4.2-2, 4.2-3, and 4.2-4. Table 4.2-2 provides the annual geometric mean values; Table 4.2-3 provides the annual arithmetic mean values; and Table 4.2-4 provides 24-hour maximum concentrations. The arithmetic mean is provided for continuity purposes and for comparison with previous historical data. The 24-hour sequential data are also provided in Appendix A. Figures 4.2-1 and 4.2-2 (for Basin F monitoring sites) provide graphical depictions of these data and comparisons with the annual standards.

Annual arithmetic mean values for TSP during FY93 ranged from a high of $56 \mu\text{g}/\text{m}^3$ at AQ2 on the northwest perimeter of RMA, to a low of $34 \mu\text{g}/\text{m}^3$ at AQ12 just north of the North Plants. The location of the maximum TSP concentration indicated influences from the Denver Metropolitan area. Annual geometric mean values ranged from a high of $51 \mu\text{g}/\text{m}^3$ at perimeter site AQ2 to a low of $28 \mu\text{g}/\text{m}^3$ at interior site AQ12. Table 2.1-1 shows all ambient air quality standards.

Table 4.2-2 Total Suspended Particulate (TSP) Sampling Results For FY93, Geometric Mean Concentrations ($\mu\text{g}/\text{m}^3$) Page 1 of 1

| | AQ1 | AQ2 | AQ3 | AQ4 | AQ5 | AQ6 | AQ7 | AQ8 | AQ9 | AQ10 | AQ11 | AQ12 | SQ1 | SQ2 | FC1 | FC2 | FC3 | FC4 | FC5 |
|-----------|-----|-----|-----|-----|-----|-----|-----|-----|-----|------|------|------|-----|-----|-----|-----|-----|-----|-----|
| October* | | | | | | | | | | | | | | | | | | | |
| November* | | | | | | | | | | | | | | | | | | | |
| December* | | | | | | | | | | | | | | | | | | | |
| January* | | | | | | | | | | | | | | | | | | | |
| February | 78 | 78 | 45 | 40 | 71 | 47 | 56 | 45 | 46 | 46 | 46 | 35 | 45 | 45 | 45 | 47 | 58 | 59 | 53 |
| March | 58 | 48 | 28 | 36 | 36 | 30 | 37 | 42 | 30 | 33 | 31 | 24 | 34 | 28 | 36 | 24 | 31 | 30 | 27 |
| April | 16 | 29 | 20 | 12 | 13 | 19 | 14 | 25 | 12 | 24 | 9 | 9 | 19 | 17 | 12 | 15 | 9 | 7 | 19 |
| May | 37 | 43 | 29 | 37 | 30 | 30 | 31 | 27 | 27 | 29 | 31 | 30 | 31 | 32 | 27 | 26 | 20 | 14 | 29 |
| June | 41 | 43 | 38 | 38 | 35 | 31 | 35 | 30 | 32 | 31 | 35 | 32 | 41 | 34 | 30 | 30 | 23 | 7 | 39 |
| July | 42 | 68 | 44 | 50 | 41 | 36 | 24 | 38 | 40 | 37 | 36 | 36 | 41 | 42 | 33 | 34 | 28 | 32 | 38 |
| August | 45 | 63 | 38 | 57 | 41 | 35 | 40 | 37 | 39 | 40 | 40 | 44 | 43 | 42 | 49 | 37 | 47 | 43 | 39 |
| September | 47 | 62 | 48 | 52 | 38 | 40 | 41 | 41 | 40 | 39 | 39 | 39 | 47 | 42 | 45 | 38 | 36 | 32 | 47 |
| Annual | 41 | 51 | 34 | 42 | 36 | 31 | 32 | 34 | 31 | 34 | 31 | 28 | 36 | 32 | 33 | 30 | 29 | 24 | 36 |

* - As noted in text, data for these months were not considered sufficiently reliable to include in the summaries.
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

Table 4.2-3 Total Suspended Particulate (TSP) Sampling Results For FY93, Arithmetic Mean Concentrations ($\mu\text{g}/\text{m}^3$) Page 1 of 1

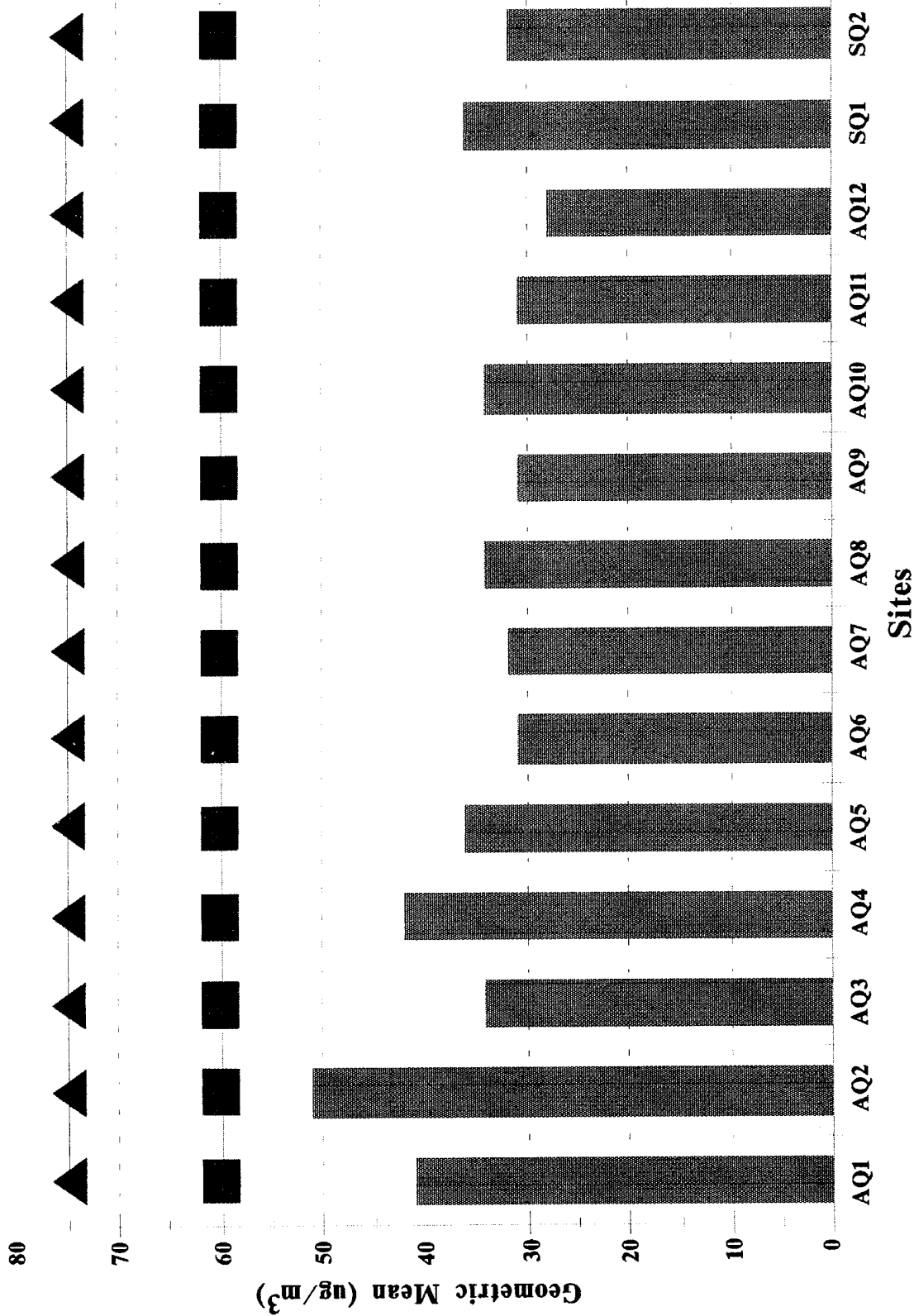
| | AQ1 | AQ2 | AQ3 | AQ4 | AQ5 | AQ6 | AQ7 | AQ8 | AQ9 | AQ10 | AQ11 | AQ12 | SQ1 | SQ2 | FC1 | FC2 | FC3 | FC4 | FC5 |
|-----------|-----|-----|-----|-----|-----|-----|-----|-----|-----|------|------|------|-----|-----|-----|-----|-----|-----|-----|
| October* | | | | | | | | | | | | | | | | | | | |
| November* | | | | | | | | | | | | | | | | | | | |
| December* | | | | | | | | | | | | | | | | | | | |
| January* | | | | | | | | | | | | | | | | | | | |
| February | 82 | 81 | 47 | 41 | 80 | 48 | 60 | 46 | 47 | 47 | 48 | 37 | 48 | 47 | 47 | 48 | 58 | 59 | 54 |
| March | 66 | 51 | 33 | 38 | 44 | 33 | 46 | 50 | 34 | 39 | 41 | 32 | 43 | 33 | 42 | 30 | 31 | 30 | 28 |
| April | 18 | 34 | 23 | 12 | 14 | 22 | 15 | 45 | 14 | 33 | 13 | 12 | 22 | 20 | 14 | 17 | 9 | 7 | 26 |
| May | 38 | 44 | 30 | 39 | 31 | 31 | 32 | 28 | 28 | 30 | 32 | 31 | 33 | 33 | 27 | 26 | 20 | 14 | 29 |
| June | 47 | 48 | 42 | 47 | 42 | 38 | 43 | 38 | 39 | 41 | 44 | 40 | 49 | 41 | 45 | 39 | 40 | 7 | 58 |
| July | 44 | 70 | 45 | 56 | 43 | 37 | 30 | 40 | 42 | 39 | 37 | 38 | 43 | 42 | 34 | 35 | 28 | 32 | 39 |
| August | 46 | 64 | 41 | 60 | 42 | 37 | 41 | 38 | 39 | 40 | 41 | 45 | 45 | 43 | 49 | 37 | 47 | 43 | 50 |
| September | 49 | 64 | 50 | 54 | 39 | 41 | 43 | 42 | 41 | 40 | 40 | 40 | 48 | 43 | 46 | 39 | 37 | 32 | 48 |
| Annual | 47 | 56 | 38 | 47 | 42 | 35 | 38 | 40 | 35 | 38 | 37 | 34 | 41 | 37 | 38 | 34 | 35 | 29 | 42 |

* - As noted in the text, data for these months were not considered sufficiently reliable to include in the summaries.
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

Table 4.2-4 Total Suspended Particulate (TSP) Sampling Results For FY93, 24-Hour Maximum Concentration ($\mu\text{g}/\text{m}^3$) Page 1 of 1

| | AQ1 | AQ2 | AQ3 | AQ4 | AQ5 | AQ6 | AQ7 | AQ8 | AQ9 | AQ10 | AQ11 | AQ12 | SQ1 | SQ2 | FC1 | FC2 | FC3 | FC4 | FC5 |
|-----------|-----|-----|-----|-----|-----|-----|-----|-----|-----|------|------|------|-----|-----|-----|-----|-----|-----|-----|
| October* | | | | | | | | | | | | | | | | | | | |
| November* | | | | | | | | | | | | | | | | | | | |
| December* | | | | | | | | | | | | | | | | | | | |
| January* | | | | | | | | | | | | | | | | | | | |
| February | 105 | 109 | 63 | 46 | 126 | 57 | 91 | 61 | 58 | 62 | 63 | 50 | 65 | 60 | 60 | 58 | 58 | 59 | 61 |
| March | 128 | 90 | 68 | 69 | 121 | 65 | 103 | 87 | 71 | 73 | 110 | 80 | 97 | 80 | 69 | 64 | 31 | 30 | 34 |
| April | 26 | 66 | 48 | 12 | 21 | 42 | 25 | 149 | 19 | 67 | 23 | 23 | 42 | 42 | 19 | 25 | 9 | 7 | 42 |
| May | 58 | 61 | 51 | 58 | 46 | 52 | 44 | 45 | 39 | 47 | 53 | 41 | 59 | 48 | 34 | 28 | 20 | 14 | 35 |
| June | 72 | 63 | 71 | 90 | 73 | 82 | 76 | 70 | 73 | 83 | 77 | 70 | 92 | 79 | 78 | 67 | 73 | 7 | 94 |
| July | 59 | 88 | 67 | 91 | 55 | 50 | 51 | 50 | 57 | 53 | 48 | 50 | 59 | 56 | 40 | 43 | 28 | 32 | 47 |
| August | 56 | 78 | 56 | 86 | 58 | 50 | 46 | 49 | 48 | 53 | 50 | 59 | 60 | 54 | 55 | 47 | 47 | 43 | 59 |
| September | 74 | 85 | 67 | 71 | 51 | 58 | 66 | 62 | 54 | 55 | 56 | 54 | 67 | 59 | 58 | 54 | 46 | 37 | 58 |
| Annual | 128 | 109 | 71 | 91 | 126 | 82 | 103 | 149 | 73 | 83 | 110 | 80 | 97 | 80 | 78 | 67 | 73 | 59 | 94 |

* - As noted in the text, data for these months were not considered sufficiently reliable to include in the summaries.
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

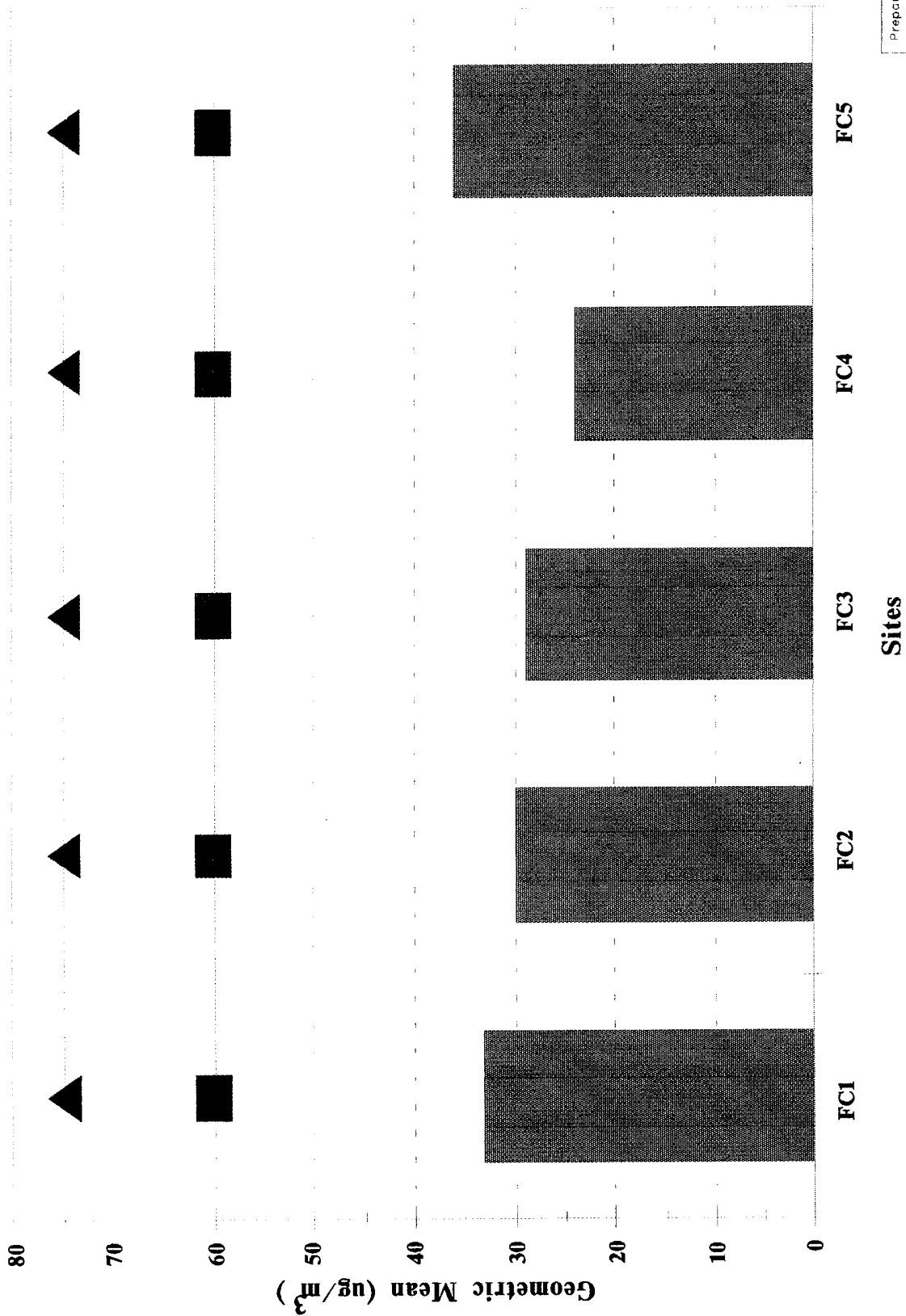


Secondary Annual Standard (60 $\mu\text{g}/\text{m}^3$)
 Primary Annual Standard (75 $\mu\text{g}/\text{m}^3$)

Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.2-1

CMP FY93 Total Suspended
Particulate Results
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



Prepared for:

U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.2-2

FY93 FC Total Suspended
Particulate Results

Rocky Mountain Arsenal

Prepared by: Ebasco Services Incorporated

There were no exceedances of the 24-hour primary ambient air quality standard of $260 \mu\text{g}/\text{m}^3$ for TSP and one exceedance of the secondary 24-hour average ambient air quality standard of $150 \mu\text{g}/\text{m}^3$ during FY93. By contrast, there were 27 exceedances of the 24-hour TSP secondary standard during FY89; all of these exceedances occurred during the Basin F remediation period and almost all were at sites close to Basin F cleanup activities.

The maximum validated 24-hour concentration during FY93 was $154 \mu\text{g}/\text{m}^3$ at AQ6. In addition, on December 8, 1993, a day when the downtown Denver CAMP Station measured $610 \mu\text{g}/\text{m}^3$ and all other Denver TSP stations recorded maximum values, 16 of the 18 RMA stations measured TSP values in excess of $100 \mu\text{g}/\text{m}^3$ with a maximum 24-hour value of $179 \mu\text{g}/\text{m}^3$ measured at AQ1 at the Arsenal western boundary. As this occurred during the period when static electricity on the filters resulted in unreliable data (the blank filter on this date measured $53 \mu\text{g}$), these results have been excluded from the summary tables. Nevertheless, these data are consistent with metropolitan Denver data and appear qualitatively useful. The December 8th sampling results including the RMA and TSP data for this date are reviewed further in Section 4.2.4.

4.2.2 Assessment of Basin F TSP Post-Remedial Impacts

One of the principal tasks of the CAQMMP has been the assessment of Basin F post-remedial impacts and progress. Since source impacts from Basin F have varied during the remedial and post-remedial monitoring periods, results and comparisons with all monitored data were considered with respect to the different phases and stages of the cleanup operations. Tables 4.2-5 and 4.2-6 (for Basin F sites) provide listings of annual geometric mean, arithmetic average and 24-hour maximum TSP monitoring results at air monitoring sites from the start of FY89 to the conclusion of FY93. Phase 1 shows results of FY88 and FY89 data, which were concurrent with remediation activity; Phase 2 (Stage 1) shows results of the FY89 program after the initial cap was placed on Basin F and extensive earth moving activities were in progress; and Phase 2 (Stage 2) shows results during the final Basin F remedial and landscaping activity. Phases 3 through 7 provide post-remedial TSP monitoring data during the CAQMMP periods of subsequent monitoring through FY93.

Table 4.2-5 Total Suspended Particulate (TSP) Sampling Results for CMP Phases 1 through 7 ($\mu\text{g}/\text{m}^3$)

Page 1 of 2

| Phase | AQ1 | AQ2 | AQ3 | AQ4 | AQ5 | AQ6 | AQ7 | AQ8 | AQ9 | AQ10 | AQ11 | AQ12 | SQ1 | SQ2 |
|---|-----|-----|-----|-----|-----|-----|-----|-----|-----|------|------|------|-----|-----|
| Summary of Geometric Mean Concentrations | | | | | | | | | | | | | | |
| Phase 1 | 50 | 68 | 40 | 40 | 39 | 35 | 40 | 40 | 38 | 81 | 74 | 55 | | |
| Phase 2-1 | 58 | 76 | 37 | 40 | 48 | 33 | 38 | 39 | 31 | 68 | 145 | 49 | | |
| Phase 2-2 | 44 | 58 | 26 | 27 | 37 | 29 | 30 | 29 | 28 | 44 | 72 | 41 | | |
| Phase 3 | 42 | 52 | 37 | 37 | 36 | 36 | 38 | 38 | 32 | 41 | 37 | 36 | | |
| Phase 4 | 40 | 52 | 32 | 27 | 33 | 36 | 32 | 31 | 28 | 30 | 30 | 28 | | |
| Phase 5 | 40 | 52 | 35 | 33 | 39 | 34 | 34 | 34 | 30 | 34 | 41 | 47 | 40 | 40 |
| Phase 6 | 48 | 54 | 37 | 39 | 39 | 36 | 39 | 42 | 37 | 34 | 42 | 37 | 46 | 44 |
| Phase 7 | 41 | 51 | 34 | 42 | 36 | 31 | 32 | 34 | 31 | 34 | 31 | 28 | 36 | 32 |
| Summary of Arithmetic Mean Concentrations | | | | | | | | | | | | | | |
| Pre-Rem | 55 | 52 | 35 | 42 | 42 | 38 | 39 | 38 | 36 | 44 | 38 | 35 | | |
| Phase 1 | 56 | 77 | 47 | 47 | 43 | 39 | 45 | 44 | 43 | 99 | 103 | 79 | | |
| Phase 2-1 | 73 | 88 | 44 | 37 | 62 | 45 | 50 | 47 | 41 | 84 | 214 | 81 | | |
| Phase 2-2 | 49 | 64 | 30 | 31 | 42 | 32 | 34 | 34 | 31 | 51 | 96 | 77 | | |
| Phase 3 | 46 | 61 | 41 | 42 | 40 | 40 | 42 | 43 | 36 | 46 | 41 | 40 | | |
| Phase 4 | 47 | 63 | 39 | 33 | 40 | 55 | 38 | 37 | 34 | 35 | 37 | 34 | | |
| Phase 5 | 46 | 61 | 41 | 39 | 46 | 39 | 40 | 41 | 34 | 40 | 55 | 42 | 47 | 48 |
| Phase 6 | 51 | 59 | 40 | 44 | 43 | 39 | 42 | 46 | 39 | 43 | 47 | 41 | 51 | 48 |
| Phase 7 | 47 | 56 | 38 | 47 | 42 | 35 | 38 | 40 | 35 | 38 | 37 | 34 | 41 | 37 |

Table 4.2-5 Total Suspended Particulate (TSP) Sampling Results for CMP Phases 1 through 7 ($\mu\text{g}/\text{m}^3$)

Page 2 of 2

| Phase | AQ1 | AQ2 | AQ3 | AQ4 | AQ5 | AQ6 | AQ7 | AQ8 | AQ9 | AQ10 | AQ11 | AQ12 | SQ1 | SQ2 |
|---|-----|-----|-----|-----|-----|-----|-----|-----|-----|------|------|------|-----|-----|
| Summary of 24-hour Maximum Concentrations | | | | | | | | | | | | | | |
| Pre-Rem | 143 | 112 | 80 | 47 | 109 | 151 | 101 | 95 | 82 | 71 | 91 | 77 | | |
| Phase 1 | 134 | 196 | 96 | 120 | 87 | 86 | 91 | 87 | 93 | 279 | 542 | 590 | | |
| Phase 2-1 | 179 | 198 | 113 | 102 | 183 | 143 | 156 | 130 | 127 | 175 | 738 | 425 | | |
| Phase 2-2 | 86 | 115 | 53 | 58 | 75 | 57 | 59 | 78 | 65 | 122 | 294 | 467 | | |
| Phase 3 | 82 | 130 | 82 | 92 | 79 | 76 | 78 | 84 | 76 | 90 | 84 | 86 | | |
| Phase 4 | 184 | 208 | 194 | 171 | 170 | 396 | 180 | 188 | 164 | 100 | 214 | 167 | | |
| Phase 5 | 113 | 202 | 119 | 116 | 100 | 93 | 127 | 103 | 96 | 102 | 286 | 95 | 94 | 116 |
| Phase 6 | 115 | 140 | 83 | 130 | 84 | 83 | 88 | 138 | 90 | 105 | 120 | 117 | 129 | 120 |
| Phase 7 | 128 | 109 | 71 | 91 | 126 | 82 | 103 | 149 | 73 | 83 | 110 | 80 | 97 | 80 |

Note: Pre-Rem refers to data collected during the Air Remedial Investigation, June 1986 to June 1987, where only arithmetic mean and maximum data were available

Phase 1 is from March 22 to December 12, 1988.

Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989.

Phase 2, Stage 2 is from February 15 to May 5, 1989.

Phase 3 is from May 6 to September 30, 1989

Phase 4 is from October 1, 1989 to September 30, 1990.

Phase 5 is from October 1, 1990 to September 30, 1991.

Phase 6 is from October 1, 1991 to September 30, 1992.

Phase 7 is from October 1, 1992 to September 30, 1993.

$\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

Table 4.2-6 Total Suspended Particulate (TSP) Sampling Results
for Basin F/IRAF Phases 1 through 7 ($\mu\text{g}/\text{m}^3$)

Page 1 of 2

| Phase | BF1/FC1 | BF2/FC2 | BF3/FC3 | BF4/FC4 | FC5 |
|---|---------|---------|---------|---------|-----|
| Summary of Geometric Mean Concentration | | | | | |
| Phase 1 | 105 | 122 | 68 | 117 | |
| Phase 2-1 | 78 | 119 | 53 | 72 | |
| Phase 2-2 | 62 | 81 | 55 | 80 | |
| Phase 3 | 49 | 39 | 44 | 42 | 54 |
| Phase 4 | 39 | 36 | 32 | 35 | 44 |
| Phase 5-1 | 44 | 32 | 31 | 36 | 40 |
| Phase 5-2 | 32 | 36 | 31 | 33 | 31 |
| Phase 6 | 45 | 38 | 37 | 38 | 48 |
| Phase 7 | 33 | 31 | 27 | 23 | 36 |
| Summary of Arithmetic Mean Concentration | | | | | |
| Phase 1 | 123 | 173 | 83 | 137 | |
| Phase 2-1 | 96 | 173 | 60 | 79 | |
| Phase 2-2 | 66 | 110 | 60 | 88 | |
| Phase 3 | 53 | 43 | 47 | 46 | 57 |
| Phase 4 | 49 | 46 | 40 | 43 | 62 |
| Phase 5-1 | 91 | 34 | 36 | 43 | 48 |
| Phase 5-2 | 37 | 41 | 34 | 35 | 45 |
| Phase 6 | 49 | 44 | 41 | 43 | 54 |
| Phase 7 | 38 | 35 | 34 | 29 | 42 |

Table 4.2-6 Total Suspended Particulate (TSP) Sampling Results
for Basin F/IRAF Phases 1 through 7 ($\mu\text{g}/\text{m}^3$)

Page 2 of 2

| Phase | BF1/FC1 | BF2/FC2 | BF3/FC3 | BF4/FC4 | FC5 |
|--|---------|---------|---------|---------|-----|
| Summary of 24-Hour Maximum Concentration | | | | | |
| Phase 1 | 399 | 902 | 286 | 324 | |
| Phase 2-1 | 276 | 687 | 143 | 162 | |
| Phase 2-2 | 107 | 514 | 107 | 160 | |
| Phase 3 | 105 | 89 | 94 | 103 | 87 |
| Phase 4 | 227 | 220 | 173 | 181 | 471 |
| Phase 5-1 | 410 | 54 | 67 | 104 | 115 |
| Phase 5-2 | 87 | 98 | 60 | 51 | 87 |
| Phase 6 | 82 | 78 | 73 | 71 | 96 |
| Phase 7 | 78 | 67 | 73 | 59 | 94 |

Note: Phase 1 is from March 22 to December 12, 1988.
Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989.
Phase 2, Stage 2 is from February 16, 1989 to May 5, 1989.
Phase 3 is from May 5 to September 30, 1989.
Phase 4 is from October 1, 1989 to September 30, 1990.
Phase 5, Stage 1 is from October 1, 1990 to January 22, 1991.
Phase 5, Stage 2 is from January 23, 1991 to September 30, 1991.
Phase 6 is from October 1, 1991 to September 30, 1992.
Phase 7 is from October 1, 1992 to September 30, 1993.

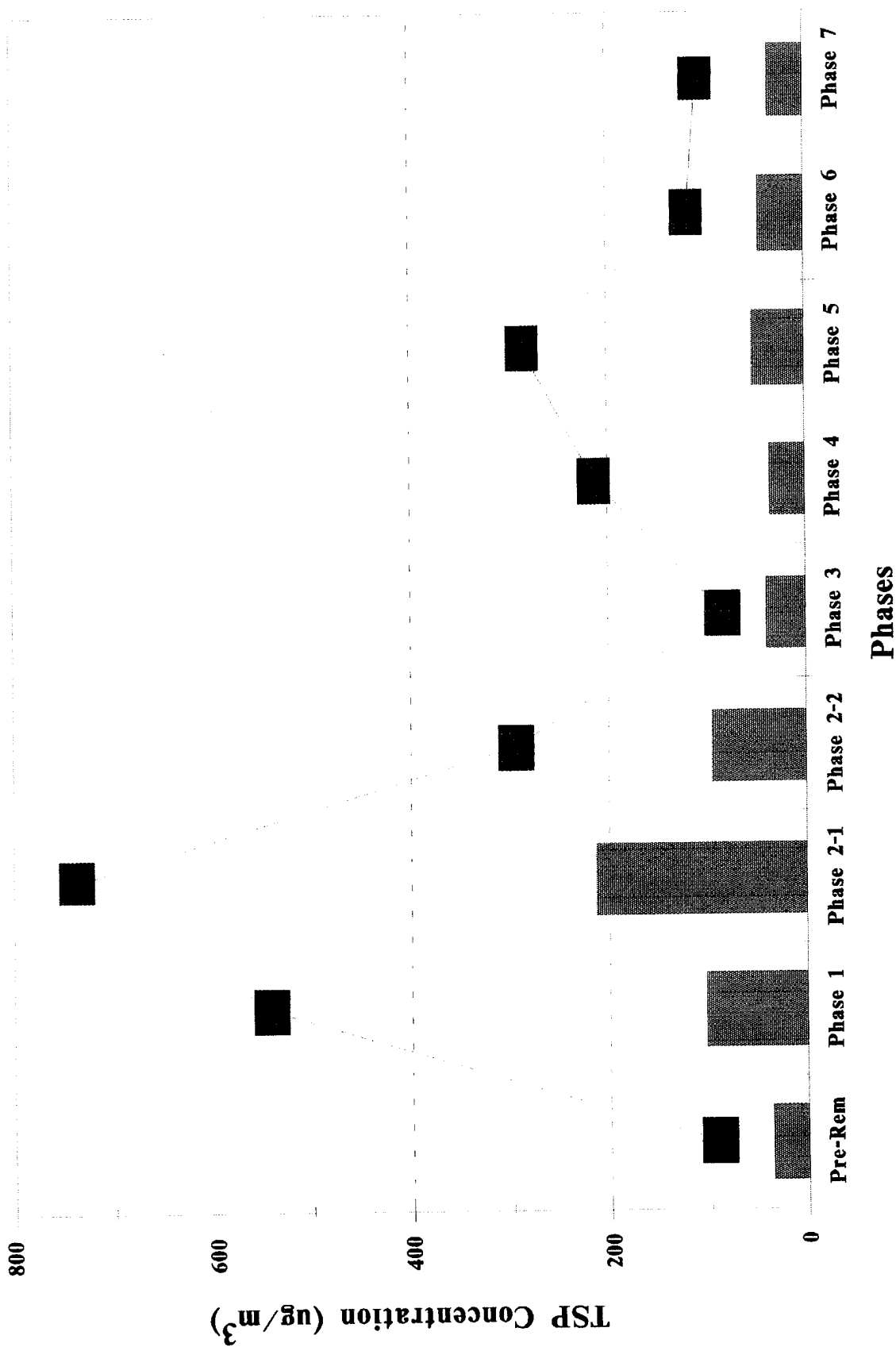
The Basin F sites were designated as "FC" sites upon commencement of the IRA-F program. Sites BF3 and BF4 remained in operation until June 1989, when they were moved slightly within the same area designated as "FC" sites under the IRA-F program.

$\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

Note that the AQ sites (Table 4.2-5) reflect little variation throughout the seven-phase remediation and post-remediation periods, with the exception of AQ10 and AQ11, directly downwind from Basin F during the remediation activities. On the other hand, Basin F remedial activity (as reflected primarily in Table 4.2-6) evidently had an impact on those stations that were located immediately adjacent to Basin F and also downwind from the basin. The highest monthly average and maximum 24-hour concentrations occurred during Phase 1, and in Phase 2 (Stage 1) when earth moving activities were most intense. A maximum 24-hour TSP value of 902 $\mu\text{g}/\text{m}^3$ was measured at BF2/FC2, and a maximum of 738 $\mu\text{g}/\text{m}^3$ was measured at AQ11. At all stations, TSP levels decreased to pre-remedial baseline values during the post-remedial periods (Phases 3, 4, 5, 6, and 7). Figure 4.2-3 provides a graphical depiction of the arithmetic means and 24-hour maximum values for each phase for CMP station AQ11 immediately downwind and north of Basin F. The transitory impacts of remediation activity are evident based on the TSP concentration differences between the remedial and post-remedial phases. Mean TSP levels reached their peaks during the intense remediation of Phases 1 and 2, then decreased to pre-remedial levels during the post-remediation periods. While maximum 24-hour values may have varied extensively on a day-to-day basis, with higher 24-hour levels occurring in Phases 4 and 5 due to local construction activities, the same general trend is evident for the average values.

It is noted that at the conclusion of the CAQMMP FY93 period, the Basin F (FC) stations were comparable to or lower than other interior stations. Station FC5 showed slightly higher geometric mean and maximum 24-hour levels than other FC stations; however, this can be accounted for by its close proximity to Ninth Avenue (within RMA), which is a gravel surfaced road, and also to the nearby tank farm which has generated considerable traffic. More detailed discussions of Basin F TSP impacts at individual stations on a monthly and seasonal basis have been provided in previous CMP/CAQMMP annual reports.

Five years of intensive post-remedial monitoring at sampling sites surrounding Basin F have indicated that TSP impacts from the restored basin are minimal and comparable to other Arsenal



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.2-3

AQ11TSP Results during all
Phases of Basin F Remediation
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

interior levels. It is anticipated that future TSP monitoring at Basin F will be conducted at one downwind site and one upwind site.

4.2.3 RMA TSP Causal Effects

Several studies were conducted (CMP FY88 and FY89) relating the impacts of wind gusts and source locations to TSP concentrations measured at the various monitoring locations. This section provides a brief summary and update on the status of these investigations. The causal effects of TSP concentrations at RMA monitoring sites are varied and cannot be reduced to two or three sources or unique meteorological events. The impacts of Basin F remediation activity were described in the previous section. The impacts from metropolitan Denver are also significant and will be discussed further in Section 4.2.4. Other factors such as wind speed, wind direction, precipitation, inversion conditions, seasonal and climatological effects, and various construction and remediation activities at RMA all influence the short-term and long-term monitoring results.

In the CMP FY88 Data Assessment Report, a direct correlation was established between 24-hour TSP levels and strong winds blowing downwind from a remediation source. During the FY88 remediation period, AQ11, downwind from Basin F, showed a correlation of 0.4 between TSP levels and winds blowing from the Basin toward the station ($\pm 30^\circ$). Further downwind at other RMA sites, the relationship was no longer evident, indicating that these effects were not only dependent on wind speed and wind direction, but also on distance. The TSP concentration data suggest that at a distance of 1 mile, remediation impacts became minimal. During FY89 when the Basin F remediation work was completed, the correlation between high TSP levels at AQ11 and winds blowing off Basin F was no longer observed (the correlation was -0.03). The observed data indicated that Basin F was a major TSP source during the remediation activity, but was largely eliminated during the post-remedial period except for minor residual impacts associated with continuing reseeding and construction activities.

A long-term statistical analysis of RMA data indicates other unique remediation and construction activities associated with the data collection period. For example, during FY88 and FY89, the Borrow Pit, which was used as fill dirt for Basin F remediation, contributed significantly to TSP concentrations measured at AQ12 (reaching a peak 24-hour concentration of $590 \mu\text{g}/\text{m}^3$); however, these levels decreased in FY90. During FY91, TSP concentrations at AQ11 and AQ12 increased, apparently as a result of regional influences. During FY92, activity from the Lower Derby Lake Spillway area strongly influenced TSP levels as evidenced by higher TSP concentrations measured at AQ6 (maximum 24-hour concentrations, $396 \mu\text{g}/\text{m}^3$).

During the past several years, TSP measurements at AQ4, which is located in a rural area at the southeastern boundary of RMA, have shown a slight but steady increase in ambient concentration levels. The annual geometric mean level at this site has increased from $27 \mu\text{g}/\text{m}^3$ in 1990, to $33 \mu\text{g}/\text{m}^3$ in 1991, to $39 \mu\text{g}/\text{m}^3$ in 1992, and to $42 \mu\text{g}/\text{m}^3$ in 1993 (excluding winter data in 1993). Although annual meteorological influences may be a partial factor, it would also appear that increased traffic on the eastern perimeter associated with construction activities at the new Denver International Airport is contributing to higher particulate levels at the RMA eastern boundary. The CAQMMP will continue to evaluate these TSP impacts, as well as the impacts of other potential air pollutants from the new airport.

In addition to local, off-post, and remediation source influences, seasonal and climatological conditions also influenced the data reported. Although some of these effects can be identified in the long term database, local and unique meteorological events also have overriding influences. Table 4.2-7 shows seasonal TSP results (arithmetic average and 24-hour maximum concentrations) for each of the RMA monitoring sites. The highest average TSP concentrations occurred in the summer and fall, the lowest in spring, and moderate levels occurred in winter. At AQ1, AQ2, and AQ5 on the west, northwest, and southern boundaries (which are closest to metropolitan Denver influences), the highest concentrations occurred in the summer or winter, while the interior monitoring sites experienced markedly higher average TSP concentrations during the summer period only (excluding Basin F impacts at sites AQ10 and AQ11). Highest

Table 4.2-7 Combined Seasonal Total Suspended Particulate (TSP) Concentrations for
FY88 to FY92 ($\mu\text{g}/\text{m}^3$)

Page 1 of 1

| | Fall | Winter | Spring | Summer |
|------------------|------|--------|--------|--------|
| Arithmetic Mean | | | | |
| AQ1 | 60 | 55 | 42 | 53 |
| AQ2 | 78 | 60 | 52 | 73 |
| AQ3 | 43 | 32 | 34 | 54 |
| AQ4 | 45 | 34 | 38 | 56 |
| AQ5 | 45 | 53 | 37 | 44 |
| AQ5B | 43 | 43 | 35 | 45 |
| AQ6 | 39 | 36 | 46 | 57 |
| AQ7 | 46 | 38 | 35 | 45 |
| AQ8 | 42 | 36 | 36 | 50 |
| AQ9 | 40 | 31 | 34 | 44 |
| AQ10 | 66 | 45 | 37 | 60 |
| AQ11 | 81 | 76 | 49 | 63 |
| AQ12 | 49 | 47 | 38 | 52 |
| SQ1 | 56 | 37 | 46 | 54 |
| SQ2 | 49 | 31 | 44 | 55 |
| Seasonal Average | 52 | 44 | 40 | 54 |
| 24-Hour Maximum | | | | |
| AQ1 | 134 | 179 | 88 | 184 |
| AQ2 | 162 | 198 | 118 | 207 |
| AQ3 | 98 | 113 | 95 | 194 |
| AQ4 | 79 | 102 | 80 | 171 |
| AQ5 | 96 | 183 | 94 | 170 |
| AQ5B | 93 | 172 | 86 | 181 |
| AQ6 | 84 | 143 | 396 | 321 |
| AQ7 | 91 | 156 | 86 | 180 |
| AQ8 | 116 | 130 | 99 | 188 |
| AQ9 | 85 | 127 | 96 | 164 |
| AQ10 | 279 | 175 | 122 | 252 |
| AQ11 | 542 | 738 | 247 | 388 |
| AQ12 | 165 | 467 | 238 | 590 |
| SQ1 | 64 | 42 | 59 | 94 |
| SQ2 | 104 | 44 | 120 | 116 |
| Seasonal Maximum | 542 | 738 | 396 | 590 |

$\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

seasonal average concentrations occurred during the fall at the perimeter sites, AQ1 and AQ2. Metropolitan Denver influences upon RMA were greatest in the fall and winter during intense inversion periods.

Seasonal summary data also highlight the unique influences of remediation activity at sites AQ10, AQ11, and AQ12. For example, AQ11 experienced especially high concentrations during Basin F cleanup work on several winter days, including a maximum report concentration of 738 $\mu\text{g}/\text{m}^3$. Station AQ12 was directly impacted by Borrow Pit activity during the summer of FY88 (maximum of 590 $\mu\text{g}/\text{m}^3$), resulting in a high overall summer seasonal TSP concentration at this station. During FY90, construction activities at the Lower Derby Lake spillway resulted in very high concentrations during spring and summer months at AQ6, downwind from the activity. The highest seasonal concentrations at this site occurred during spring and summer of FY90. The anomalously high value of 388 $\mu\text{g}/\text{m}^3$ at AQ11 on September 9, 1991 contributed to higher summer seasonal concentrations during FY91 as compared to previous post-remedial phases. No anomalously high values occurred at any of the stations during FY93.

Rather than identifying strong seasonal trends, these data confirm the overriding influence of individual source impacts and meteorological events. As the database continues to expand, real trends in seasonal impacts may become more apparent. Also, as the effect of individual events are further minimized, higher TSP concentrations are anticipated at the perimeter stations close to metropolitan Denver, especially at AQ1 and AQ2, during the fall and winter periods. Also, increasing traffic and industrial activity resulting from Denver International Airport operations may have a significant influence at AQ4 in the future.

4.2.4 Denver Metropolitan Area TSP Influences

4.2.4.1 Denver FY93 Results

RMA activities were not the only contribution to TSP concentrations that influenced RMA and adjacent areas as evident from previous assessments. The 1986-1987 Remedial Investigation study clearly established that prior to remediation, TSP values were highest at the perimeter

stations and lowest in the interior, suggesting that the principal long-term major sources in the area were the industrial or transportation activities of metropolitan Denver. The Basin F Remediation Program temporarily altered this situation at several sites in the interior of RMA; however, metropolitan Denver sources were still a major contributor.

Table 4.2-8 shows the major stationary sources and updated emission inventory with TSP emission rates of 25 tpy or more surrounding RMA. Several of the sources are located within 2 miles of RMA to the west and southwest as shown in Figure 4.2-4, and contribute a large portion of the total TSP emissions for Adams, Arapahoe, and Denver counties (CDH 1993). Compared to these major TSP sources, RMA accounts for a small fraction of the total TSP emissions. In addition to these external stationary sources, dust from vehicle traffic and off-road sources contributes greatly to the TSP concentrations.

Table 4.2-9 is a summary of Denver metropolitan area TSP data measured concurrently with the FY93 program, as well as with previous remediation and post-remediation phases; because of the decrease in CDH emphasis on TSP, several Denver area TSP monitoring sites have been discontinued during this period. Figure 4.2-5 shows the geographic distribution of TSP geometric mean levels across the metropolitan area, including RMA, for FY93. The highest levels were in the downtown area at the CAMP station (2105 Broadway), where TSP concentrations have exceeded the ambient air quality primary and secondary standards for at least the past 15 years. As shown in Table 4.2-9, the annual geometric mean frequently exceeded $100 \mu\text{g}/\text{m}^3$ and the 24-hour maximum concentration has reached levels of $500 \mu\text{g}/\text{m}^3$ to $600 \mu\text{g}/\text{m}^3$ depending upon the intensity of specific inversion episodes. The highest level in the last 6 years occurred on December 8, 1992, when $610 \mu\text{g}/\text{m}^3$ was measured at the CAMP station. This episode was heaviest in the downtown section of the city but was distributed across the metropolitan area as shown in Figure 4.2-6. Although RMA TSP data are suspect on this date, the mean concentration of 18 nonvalidated observations was $115 \mu\text{g}/\text{m}^3$ with a maximum measurement of $179 \mu\text{g}/\text{m}^3$ at the western perimeter.

Table 4.2-8 Sources of Total Suspended Particulates in the Vicinity of Rocky Mountain Arsenal (25 tons per year [tpy] or more) Page 1 of 1

| Facility | City (Source) | Update Year | Map No. | Estimate (tons per year) |
|--|------------------|----------------|------------|--------------------------------|
| Kiewit Western | Portable Source | 90 | 1 | 1830.2 |
| Conoco Inc. - Denver Refinery | Commerce City | 91 | 2 | 465.0 |
| Denver Metals | Denver | 90 | 3 | 336.1 |
| Public Service Company - Cherokee | Denver | 91 | 4 | 254.1 |
| Colorado Refining Company/Total Petroleum | Commerce City | 91 | 5 | 174.2 |
| Public Service Company - Arapahoe | Denver | 91 | 6 | 94.4 |
| Waste Management of Colo-Denver Arapahoe | Aurora | 92 | 7 | 72.4 |
| Cooley Gravel Co., 88th & Riverdale Rd. | No city name | 89 | 8 | 69.9 |
| Purina Mills, Inc. - Henderson Mill | Henderson | 91 | 9 | 66.4 |
| Centennial Materials - Thornburg | Bennett | 91 | 10 | 65.4 |
| Cooley Gravel | Brighton | 92 | 11 | 63.5 |
| BFI - Tower Landfill | Commerce City | 92 | 12 | 57.5 |
| Bennett Sand and Gravel Inc. | Bennett | 92 | 13 | 50.2 |
| Ralston Purina Company Pet Food Plant | Denver | 92 | 14 | 40.8 |
| Fast Construction Company, 64th and Steel | Commerce City | 90 | 15 | 32.4 |
| Conagra | Adams County | 91 | 16 | 31.2 |
| Mountain Cement Co., 1630 35th Street | Denver | 89 | 17 | 29.0 |
| Brannan S&G | Denver | 90 | 18 | 28.8 |
| Bituminous Roadways of Colorado, Inc. | Commerce City | 92 | 19 | 28.7 |
| Buckley Air Force Base - Colorado Air National Guard | Aurora | 91 | 20 | 27.3 |
| Owens Corning | Denver | 92 | 21 | 26.5 |
| Lightweight Materials, Inc. | Denver | 92 | 22 | 26.5 |
| Highway 36 Land Dev-BFI Sec 25 & 36 T3S R15 | Broomfield | 90 | 23 | 26.2 |
| Rocky Mountain Arsenal | Commerce City | 92 | 24 | 1.1 |

Source: EISPS inventory from CDH, 1993.

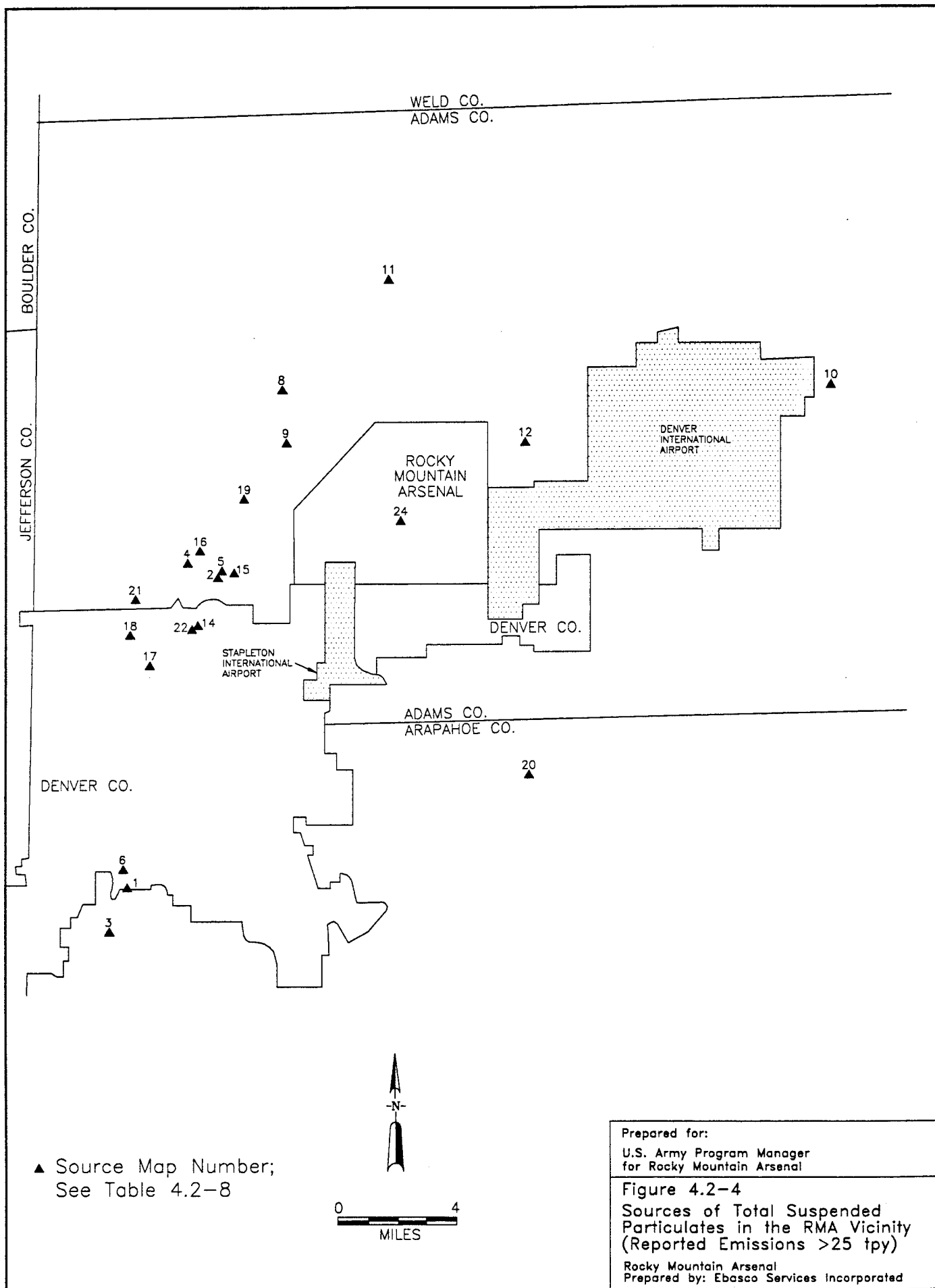
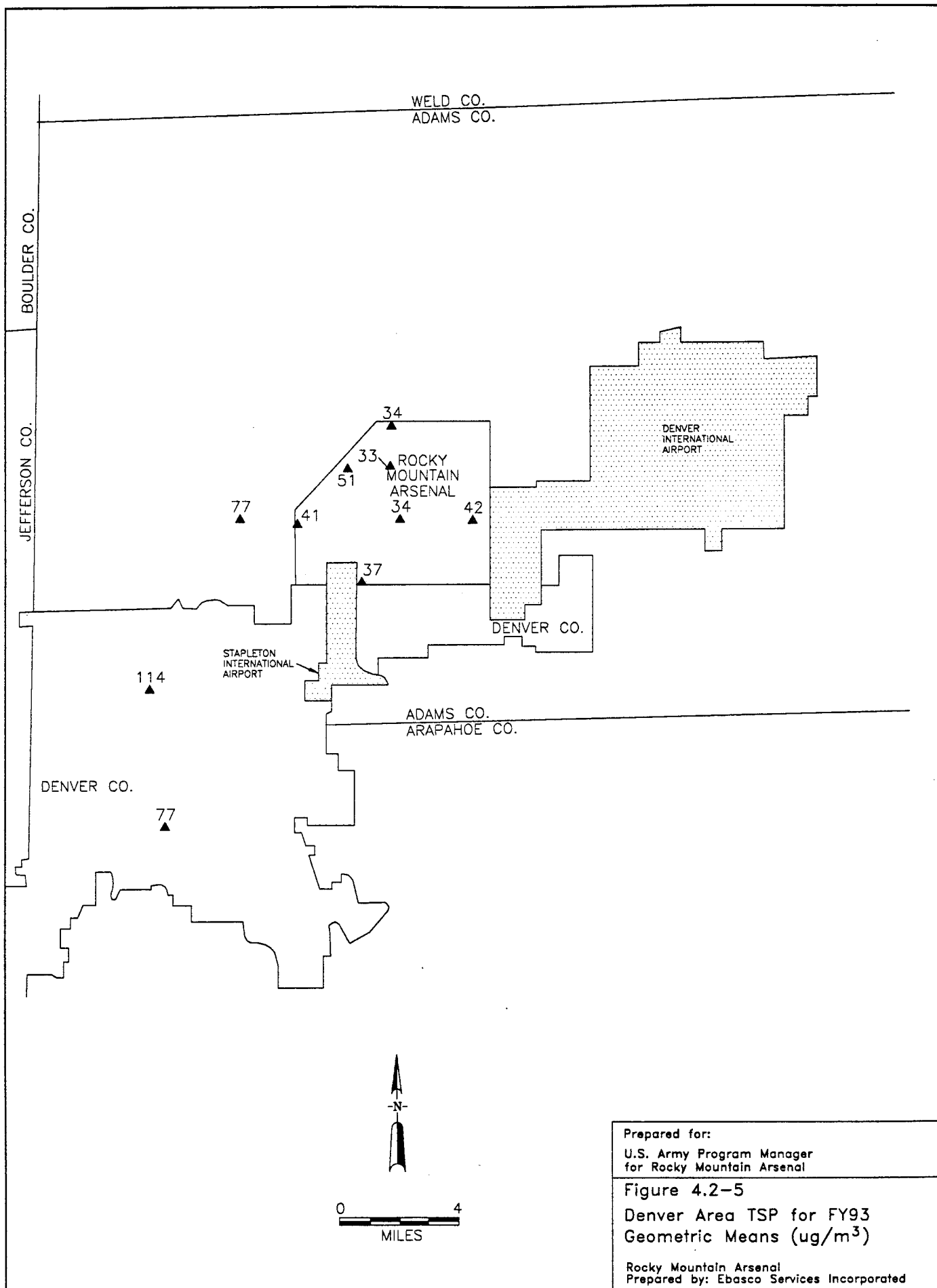


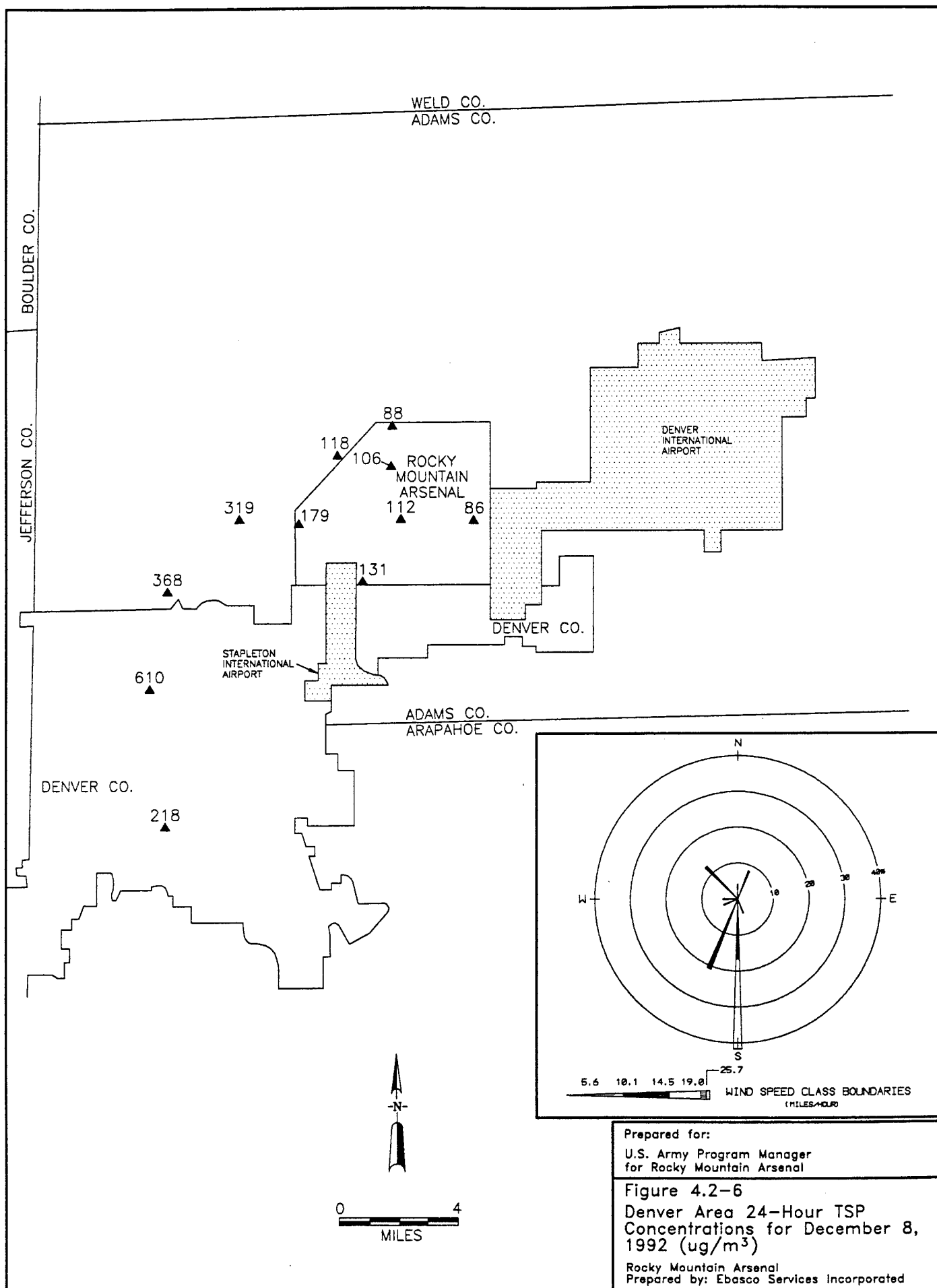
Table 4.2-9 Denver Metropolitan Area Total Suspended Particulates ($\mu\text{g}/\text{m}^3$) Page 1 of 1

| Station | Phase 1 | Phase 2 Stage 1 | Phase 2 Stage 2 | Phase 3 | Phase 4 (FY90) | Phase 5 (FY91) | Phase 6* (FY92) | Phase 7 (FY93) |
|-------------------------------|---------|--------------------|--------------------|---------|-------------------|-------------------|--------------------|-------------------|
| <u>Geometric Means</u> | | | | | | | | |
| Adams City | 87.3 | 103.0 | 61.2 | 71.0 | 66.8 | 73.0 | - | 77.0 |
| 414 14th Street | 83.7 | 137.4 | - | - | - | - | - | - |
| CAMP/2105 Broadway | 107.9 | 192.2 | 132.2 | 98.9 | 100.8 | 109.6 | 120.0 | 113.7 |
| 1050 S. Broadway | 74.1 | 135.2 | 93.7 | 63.4 | 66.6 | 52.4 | - | 76.5 |
| 208 3rd Street | 62.9 | 62.7 | 57.0 | 90.2 | - | - | - | - |
| 4857 S. Broadway | 64.4 | 84.9 | - | - | - | - | - | - |
| <u>24-Hour Maximum Values</u> | | | | | | | | |
| Adams City | 283.0 | 244.0 | 85.0 | 131.0 | 206.0 | 228.0 | 268.0 | 319.0 |
| 414 14th Street | 242.0 | 183.0 | - | - | - | - | - | - |
| CAMP/2105 Broadway | 260.0 | 472.0 | 256.0 | 149.0 | 363.0 | 563.0 | 379.0 | 610.0 |
| 1050 S. Broadway | 162.0 | 278.0 | 162.0 | 100.0 | 287.0 | 177.0 | - | 212.0 |
| 208 3rd Street | 153.0 | 179.0 | 144.0 | 385.0 | - | - | - | - |
| 4857 S. Broadway | 152.0 | 94.0 | - | - | - | - | - | - |

* Phase 6 data taken from incomplete record
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter
 - Samples not collected or data not available

Phase 1 is from March 22 to December 12, 1988
 Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989
 Phase 2, Stage 2 is from February 16 to May 5, 1989
 Phase 3 is from May 6 to September 30, 1989
 Primary 24-hour standard for Colorado is $260 \mu\text{g}/\text{m}^3$ (suspended in August 1993 for one year)
 Primary annual (geometric mean) standard for Colorado is $75 \mu\text{g}/\text{m}^3$ (suspended in August 1993 for one year)



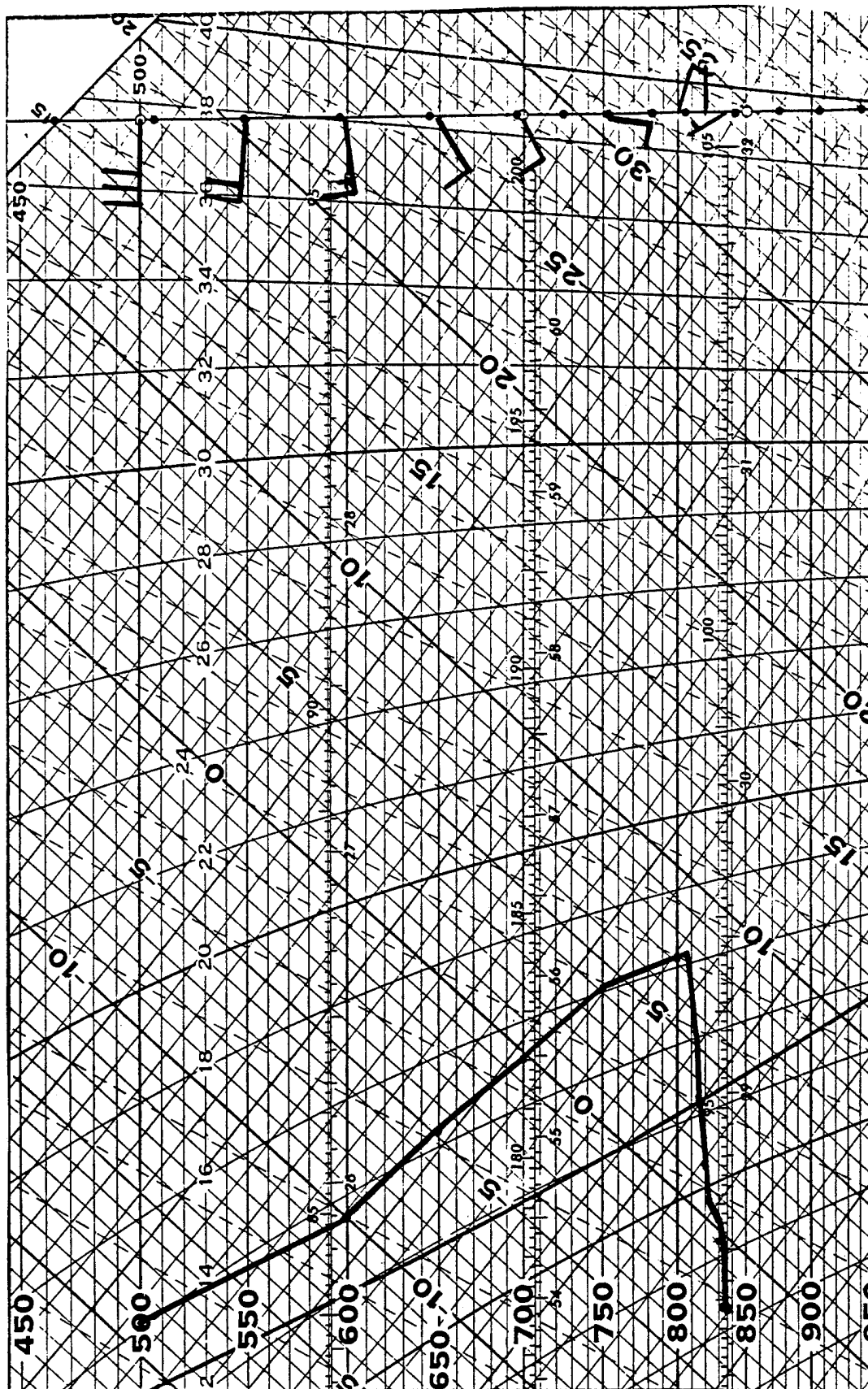


The upper air sounding on this date (0000Z, 8 December 1993) at Stapleton Airport is shown in Figure 4.2-7. This is the afternoon sounding (1700MST) and indicates that the inversion did not "break" during the day. The inversion extended to 1,500 feet above ground level and, in fact, did not break for several days. Winds were light (under 5 mph) and primarily from the south (see wind rose in Figure 4.6-6). Temperatures remained below freezing for several days. This episode was associated with high levels of PM-10, metals, VOCs, and gaseous criteria pollutants at RMA and will be discussed further in appropriate sections of this report.

The December 8, 1992 episode is typical of occasional fall and winter occurrences when, under intense inversion conditions, high TSP concentrations from the general metropolitan area have been transported to RMA. This episode, however, was especially severe.

As a general rule, mean concentrations decrease from the center of Denver and are typically about 50 percent lower in the outlying suburbs, including RMA. Prior to Basin F remediation activity in 1986-1987, the average concentration at RMA for all sites was $40 \mu\text{g}/\text{m}^3$. The boundary sites, AQ1 and AQ2, averaged $54 \mu\text{g}/\text{m}^3$ and were comparable to other suburban measurements, whereas the interior RMA sites averaged $38 \mu\text{g}/\text{m}^3$ (see Table 4.2-5). This trend was temporarily altered during Basin F remediation activities. However, during the FY93 post-remedial period, the average TSP concentration for all CMP sites was $34 \mu\text{g}/\text{m}^3$, with the boundary sites averaging $42 \mu\text{g}/\text{m}^3$ and the interior sites averaging $32 \mu\text{g}/\text{m}^3$. These results are similar to the pre-remedial monitoring results at RMA.

One important difference between high TSP values measured in metropolitan Denver and those at RMA is that those experienced in metropolitan Denver were frequently associated with high pollution episodes in which industrial and vehicle traffic emissions were trapped under a blanketing inversion that spread across the area. In these cases, winds were generally light to moderate. Occasionally, these metropolitan Denver impacts extended to the vicinity of RMA; however, during remedial operations at RMA, higher concentrations were invariably associated with strong wind gusts that lifted dirt that had been loosened by handling.



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.2-7

Denver Upper-Air Sounding
for 12/7/92 1700 MST

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

4.2.5 Analysis Implications for Mitigation and Controls

The general results of these investigations indicate that many of the RMA sources were localized and that although very high concentrations were frequently measured adjacent to remedial activities, concentrations decreased significantly with distance from the source. Consequently, those sources located in the RMA interior were less likely to cause impacts off post, while those sources closer to the boundaries were more likely to cause temporary impacts at the boundary when strong or gusty winds were directed from the source to the boundary.

Also, certain atmospheric conditions and other factors that vary diurnally can spread pollutants greater distances from the disturbing source. The ongoing CAQMMP and summarized data results, with analyses of probable causative factors, form the basis for deciding appropriate mitigating actions. For example, when the monitoring data show excessive TSP concentrations, chemical stabilizers, covers, reseeded, and landscaping can be used over sensitive areas, as in the case of Basins F and A. Dirt excavation, hauling, and unloading can be minimized during periods of high winds and poor dispersion potential, particularly when transport of the resultant airborne dust would be towards nearby boundaries. Surface watering during these operations would also help minimize the generation of dust. This practice has been applied periodically and as appropriate at RMA.

4.2.6 Summary

TSP concentrations observed at RMA can be attributed to two principal sources: (1) the influx of industrial or urban traffic emissions from metropolitan Denver, and (2) remedial activity producing wind-blown dust, usually during high wind episodes. A noticeable increase in TSP concentrations at sites adjacent to or directly downwind from construction work, primarily at Basin F, occurred during the FY88 and FY89 CMP monitoring periods because of increased remedial activity at RMA. However, the impacts from sources originating within RMA were localized and decreased rapidly with distance from the source. From FY90 through FY93, after

termination of the Basin F cleanup program, TSP concentrations downwind from Basin F decreased significantly.

4.3 PARTICULATE MATTER LESS THAN 10 MICROMETERS

4.3.1 PM-10 Monitoring Program

Respirable particulate matter, designated PM-10, are those particles in air with aerodynamic diameters less than 10 micrometers (μm). These particles are typically small enough to be inhaled deep within the lungs. As a continuation of the ongoing CAQMMP, PM-10 was monitored at twelve individual sites (AQ1, AQ2, AQ3, AQ4, AQ5, AQ6, AQ8, AQ9, AQ10, FC1, SQ1, and SQ2) during the FY93 monitoring period. A collocated monitor at Site AQ5 (designated as AQ5C) was also a continuation from the previous monitoring years. Because of the increasing PM-10 emphasis, three additional monitoring sites AQ4, AQ6, and AQ8 were established in FY93. A summary of sampling frequency and data recovery for FY93 PM-10 monitoring is displayed in Table 4.3-1. The discussion in Section 4.2.1 pertaining to TSP filters also applies to PM-10. Samples collected prior to February 1, 1993 are considered suspect and are not included in the annual summaries. Again, all PM-10 FY93 measurements taken during the year are provided in Appendix B, including the data prior to February 1; some of these data may be reliable and may be considered for analysis on a case-by-case basis.

The average (arithmetic mean) and 24-hour maximum concentrations for the CAQMMP FY93 PM-10 program are presented in Tables 4.3-2 and 4.3-3, respectively. As shown in Section 2, Table 2.1-1, the PM-10 federal and state standards are arithmetic mean values of $50 \mu\text{g}/\text{m}^3$ on an annual basis, and $150 \mu\text{g}/\text{m}^3$ for a 24-hour period. For FY93, PM-10 did not exceed either the annual or 24-hour ambient air quality standards at any of the monitored sites at RMA.

Highest PM-10 concentrations were measured at the RMA boundaries. AQ1 measured the highest annual average concentration of $25 \mu\text{g}/\text{m}^3$ during FY93. AQ2, at the northwestern boundary and closest to Commerce City traffic, also measured an annual average of $25 \mu\text{g}/\text{m}^3$. The highest 24-hour levels were $86 \mu\text{g}/\text{m}^3$ at AQ1 and $71 \mu\text{g}/\text{m}^3$ at AQ2 during FY93. AQ4,

Table 4.3-1 Summary of Particulate Matter less than 10 Micrometers
Monitoring for FY93

Page 1 of 1

| Station | Samples Scheduled | Samples Collected | Valid Samples | Field % Recovery | Overall % Recovery |
|-----------|----------------------|----------------------|------------------|---------------------|-----------------------|
| AQ1 | 61 | 59 | 38 | 97 | 62 |
| AQ2 | 61 | 59 | 36 | 97 | 59 |
| AQ3 | 61 | 60 | 38 | 98 | 62 |
| AQ4 | 61 | 58 | 38 | 95 | 62 |
| AQ5 | 61 | 60 | 35 | 98 | 57 |
| AQ6 | 61 | 60 | 36 | 98 | 59 |
| AQ8 | 61 | 51 | 36 | 84 | 59 |
| AQ9 | 61 | 60 | 36 | 98 | 59 |
| AQ10 | 61 | 59 | 39 | 97 | 64 |
| FC1 | 31 | 30 | 21 | 97 | 68 |
| SQ1 | 61 | 61 | 38 | 100 | 62 |
| SQ2 | 61 | 59 | 38 | 97 | 62 |
| Duplicate | 61 | 60 | 35 | 98 | 57 |
| Overall | 763 | 736 | 464 | 96 | 61 |

% Percent

Table 4.3-2 Particulate Matter less than 10 Micrometers (PM-10) Sampling Results for FY93
Arithmetic Mean Concentration ($\mu\text{g}/\text{m}^3$)

Page 1 of 1

| | AQ1 | AQ2 | AQ3 | AQ4 | AQ5 | AQ6 | AQ8 | AQ9 | AQ10 | SQ1 | SQ2 | FC1 |
|-----------|-----|-----|-----|-----|-----|-----|-----|-----|------|-----|-----|-----|
| October* | | | | | | | | | | | | |
| November* | | | | | | | | | | | | |
| December* | | | | | | | | | | | | |
| January* | | | | | | | | | | | | |
| February | 56 | 46 | 32 | 29 | 44 | 31 | 33 | 33 | 30 | 35 | 34 | 32 |
| March | 35 | 22 | 20 | 19 | 29 | 16 | 31 | 16 | 21 | 25 | 16 | 23 |
| April | 10 | 11 | 11 | 14 | 8 | 11 | 26 | 5 | 11 | 12 | 10 | 6 |
| May | 25 | 22 | 18 | 23 | 20 | 14 | 18 | 16 | 18 | 18 | 17 | 14 |
| June | 21 | 27 | 19 | 22 | 17 | 16 | 17 | 17 | 17 | 20 | 18 | 16 |
| July | 20 | 31 | 19 | 36 | 17 | 17 | 18 | 25 | 16 | 20 | 20 | 16 |
| August | 24 | 23 | 19 | 31 | 18 | 18 | 19 | 21 | 19 | 20 | 19 | 20 |
| September | 22 | 26 | 20 | 24 | 18 | 19 | 19 | 18 | 18 | 20 | 19 | 17 |
| Annual | 25 | 25 | 19 | 24 | 21 | 16 | 22 | 18 | 18 | 20 | 18 | 18 |

* - As noted in text, data for these months were not considered sufficiently reliable to include in the summaries.
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

Table 4.3-3 Particulate Matter less than 10 Micrometers (PM-10) Sampling Results for FY93
24-Hour Maximum Concentration ($\mu\text{g}/\text{m}^3$)

Page 1 of 1

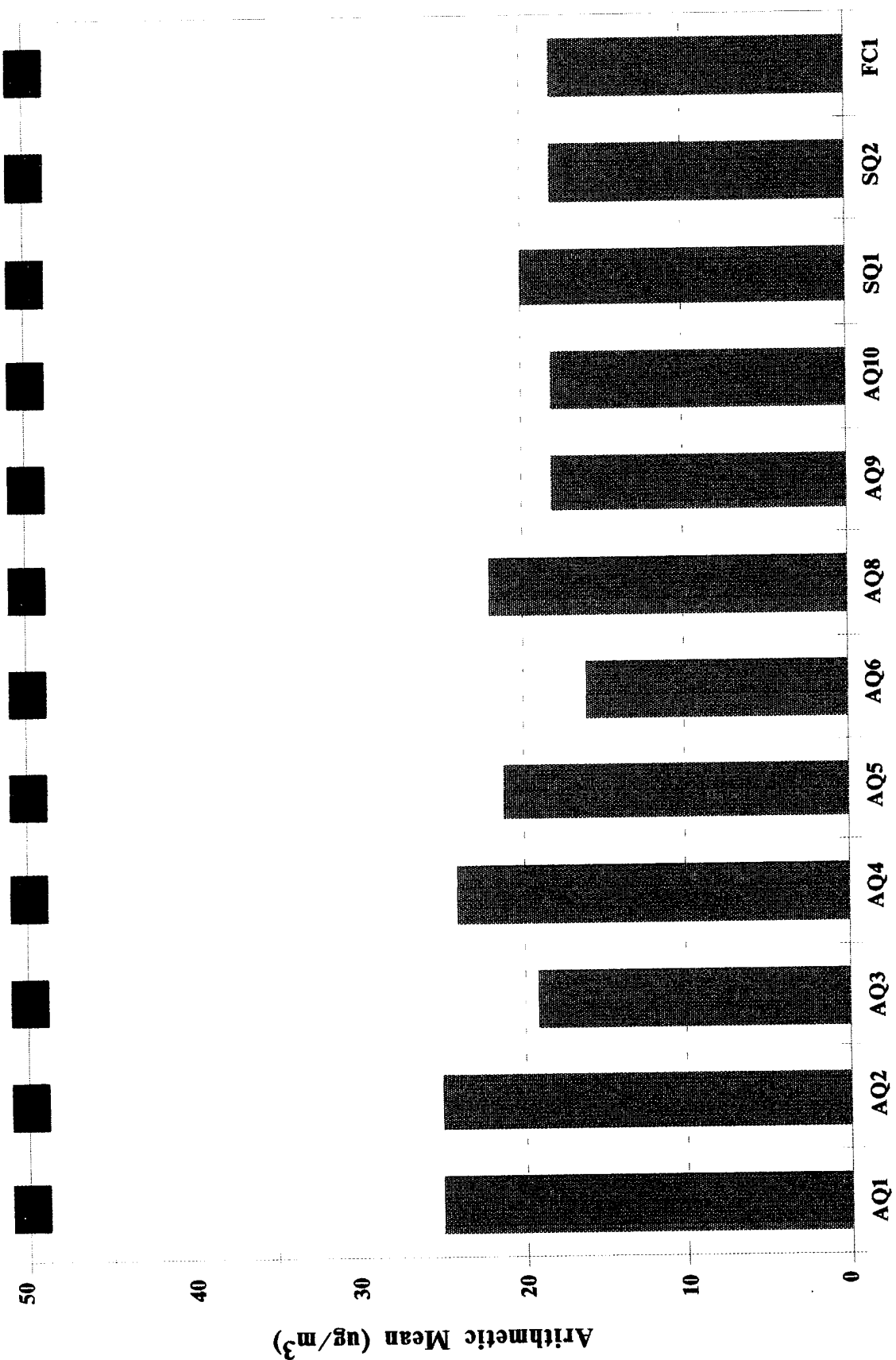
| | AQ1 | AQ2 | AQ3 | AQ4 | AQ5 | AQ6 | AQ8 | AQ9 | AQ10 | SQ1 | SQ2 | FCI |
|-----------|-----|-----|-----|-----|-----|-----|-----|-----|------|-----|-----|-----|
| October* | | | | | | | | | | | | |
| November* | | | | | | | | | | | | |
| December* | | | | | | | | | | | | |
| January* | | | | | | | | | | | | |
| February | 74 | 71 | 41 | 34 | 66 | 39 | 42 | 40 | 44 | 50 | 46 | 41 |
| March | 86 | 29 | 29 | 35 | 66 | 26 | 64 | 36 | 39 | 50 | 36 | 32 |
| April | 15 | 20 | 22 | 40 | 14 | 19 | 77 | 7 | 22 | 24 | 24 | 9 |
| May | 38 | 30 | 26 | 35 | 25 | 23 | 25 | 23 | 25 | 29 | 25 | 19 |
| June | 30 | 51 | 30 | 41 | 27 | 27 | 28 | 28 | 31 | 40 | 31 | 27 |
| July | 30 | 36 | 25 | 55 | 26 | 23 | 25 | 44 | 25 | 25 | 27 | 21 |
| August | 30 | 37 | 27 | 48 | 22 | 21 | 23 | 27 | 24 | 25 | 24 | 23 |
| September | 30 | 40 | 28 | 32 | 22 | 27 | 29 | 26 | 25 | 25 | 25 | 22 |
| Annual | 86 | 71 | 41 | 55 | 66 | 39 | 77 | 44 | 44 | 50 | 46 | 41 |

* - As noted in text, data for these months were not considered sufficiently reliable to include in the summaries.
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

which is a new PM-10 monitoring site at the eastern boundary, and is closest to increasing traffic from the Denver International Airport development activities, measured an arithmetic mean value of $24 \mu\text{g}/\text{m}^3$ (the third highest) and a maximum 24-hour value of $55 \mu\text{g}/\text{m}^3$. It is interesting that all of the other PM-10 monitoring stations measured arithmetic mean concentrations ranging from 16 to $21 \mu\text{g}/\text{m}^3$, suggesting little variation in PM-10 levels except at the boundary locations. A summary of annual arithmetic mean PM-10 concentrations for FY93 is presented graphically in Figure 4.3-1. The FY93 PM-10 annual arithmetic mean values were lower than those obtained during the FY92 monitoring period. A likely explanation for this is that the FY93 data do not include several fall and winter months when static electricity problems with the PM-10 filters made these data suspect. A qualitative analysis of the data results shown in Appendix B indicate occasional moderate to high PM-10 levels that also correspond to high PM-10 levels in metropolitan Denver and that were associated with winter inversion episodes. This will be discussed further in Section 4.3.3 in connection with the December 8, 1992 intense inversion episode.

4.3.2 Basin F PM-10 Impacts

PM-10 was not monitored in close proximity to Basin F during the Basin F remediation program or the follow-on IRA-F program. With increasing emphasis on respirable particles less than 10 micrometers, PM-10 monitoring at RMA has increased from three stations in 1987 to fifteen stations in FY93, including FC1 which is directly downwind from the remediated Basin F. Table 4.3-4 shows annual arithmetic mean and 24-hour ambient PM-10 concentrations for the various phases of the Basin F assessment. As noted, during the early remediation activity, there were no close-in PM-10 monitors, and it is difficult to discern significant increases at the several (more distant) stations that were sampled. Earlier CMP reports have suggested that the impact on PM-10 levels from remediation activity was less than from the larger total suspended particulates (RLSA 1991). This was based on an assessment of the more distant PM-10 monitoring stations (AQ1, AQ2, AQ3, AQ5 and AQ6). Under recent revisions to the CAQMMP, it is anticipated that PM-10 monitoring will be an integral part of all future remediation monitoring programs.



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.3-1
FY93 CAQMMP Particulate Matter
less than 10 Micrometers
(PM-10) Results
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

Sites

Primary Annual Standard (50 ug/m³)

Table 4.3-4 Particulate Matter less than 10 Micrometers (PM-10) Results for Phases 1 through 7
($\mu\text{g}/\text{m}^3$) Page 1 of 1

| Phase | AQ1 | AQ2 | AQ3 | AQ5 | AQ9 | AQ10 ^c | SQ1 | SQ2 | FC1 |
|-------------------------------------|-----|-----|-----|-----|-----|-------------------|-----|-----|-----|
| Arithmetic Mean^a | | | | | | | | | |
| Pre-Rem | | 36 | | 33 | 18 | | | | |
| Phase 1 | 39 | 30 | 30 | 22 | 22 | | | | |
| Phase 2 Stage 1 | 52 | 36 | 30 | 35 | 27 | | | | |
| Phase 2 Stage 2 | 31 | 22 | 19 | 21 | 17 | | | | |
| Phase 3 | 24 | 26 | 21 | 21 | 20 | | | | 23 |
| Phase 4 | 23 | 25 | 20 | 21 | 19 | 36 | | | 24 |
| Phase 5 Stage 1 | * | * | * | * | * | * | * | * | 26 |
| Phase 5 Stage 2 | 26 | 25 | 18 | 20 | 20 | 21 | 23 | 23 | 18 |
| Phase 6 | 30 | 30 | 22 | 24 | 24 | 26 | 26 | 26 | 27 |
| Phase 7 | 25 | 25 | 19 | 21 | 18 | 18 | 20 | 18 | 18 |
| 24- Hour Maximum^b | | | | | | | | | |
| Pre-Rem | | 94 | | 90 | 36 | | | | |
| Phase 1 | 77 | 72 | 60 | 47 | 51 | | | | |
| Phase 2 Stage 1 | 168 | 105 | 94 | 116 | 94 | | | | |
| Phase 2 Stage 2 | 56 | 33 | 30 | 37 | 27 | | | | |
| Phase 3 | 39 | 46 | 37 | 33 | 36 | | | | 35 |
| Phase 4 | 72 | 81 | 72 | 66 | 71 | 90 | | | 102 |
| Phase 5 Stage 1 | * | * | * | * | * | * | * | * | 84 |
| Phase 5 Stage 2 | 70 | 77 | 48 | 43 | 92 | 48 | 43 | 47 | 47 |
| Phase 6 | 92 | 76 | 41 | 52 | 53 | 43 | 56 | 50 | 52 |
| Phase 7 | 86 | 71 | 41 | 66 | 44 | 44 | 50 | 46 | 41 |

- a - Annual arithmetic mean standard is $50 \mu\text{g}/\text{m}^3$.
b - 24-Hour Maximum standard is $150 \mu\text{g}/\text{m}^3$.
c - AQ10 was installed in the northwest quarter of Section 1 on September 5, 1990.
* - No samples collected, contract lapse.
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter

Note: Pre-Rem refers to data collected during the Air Remedial Investigation (Environmental Science & Engineering, Inc., 1988).

Phase 1 is from March 22 to December 12, 1988.
Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989.
Phase 2, Stage 2 is from February 16, 1989 to May 5, 1989.
Phase 3 is from May 5 to September 30, 1989.
Phase 4 is from October 1, 1989 to September 30, 1990 (FY90).
Phase 5, Stage 1 is from October 1, 1990 to January 23, 1991.
Phase 5, Stage 2 is from January 24, 1991 to September 30, 1991 (FY91).
Phase 6 is from October 1, 1991 to September 30, 1992 (FY92).
Phase 7 is from October 1, 1992 to September 30, 1993 (FY93).

With respect to post-remediation PM-10 impacts at Basin F, FY93 data from FC1, directly downwind from the remediated basin, indicate an annual arithmetic mean of $18 \mu\text{g}/\text{m}^3$ and a 24-hour maximum value of $41 \mu\text{g}/\text{m}^3$, among the lowest levels at RMA.

4.3.3 Metropolitan Denver PM-10 Data

As in the case of TSP, PM-10 concentrations were higher in metropolitan Denver than they were at RMA. The distribution of PM-10 data across metropolitan Denver and the prevailing south and south-southwest wind flow across the area suggests that there is an influx of PM-10 into RMA from the Denver area. PM-10 annual arithmetic mean and 24-hour maximum concentrations for Denver throughout the CMP/CAQMMP history, including FY93, are shown in Table 4.3-5. Note the number of additional PM-10 monitoring stations that have been added by CDH in 1993, reflecting the increasing emphasis on this parameter. Figure 4.3-2 shows the distribution of PM-10 annual average values across metropolitan Denver and RMA for the FY93 period. The highest annual average was at the "Gates Rubber" site at 1050 S. Broadway, $40.7 \mu\text{g}/\text{m}^3$; however, comparable levels were also measured at the CAMP station at 21st Street and Broadway, and also at the Adams City 72nd Street site. RMA averages were considerably lower, but did not include data from November 1992 through January 1993. It is also noted that the highest PM-10 levels occurred at RMA boundaries.

Maximum 24-hour PM-10 concentrations were $183 \mu\text{g}/\text{m}^3$ at Adams City and $175 \mu\text{g}/\text{m}^3$ at the Gates site. These occurred on December 8, 1992 as did all 24-hour maximum concentrations at other individual sampling sites both in Denver and at RMA. This was also the day that the highest concentrations of TSP, most of the metals, and most of the VOCs occurred. Figure 4.3-3 shows the distribution of PM-10 over Denver and RMA on December 8 (note, however, that RMA PM-10 data could not be validated on this date). See Section 4.2.4 in the TSP section for a further discussion of this event.

Table 4.3-5 Denver Metropolitan Area Particulate Matter of less than 10 Micrometers (PM-10) ($\mu\text{g}/\text{m}^3$) Page 1 of 1

| Station | Phase 1 | Phase 2 Stage 1 | Phase 2 Stage 2 | Phase 3 | Phase 4 (FY90) | Phase 5 (FY91) | Phase 6 (FY92) | Phase 7 (FY93) |
|---|---------|--------------------|--------------------|---------|-------------------|-------------------|-------------------|-------------------|
| <u>Arithmetic Means^a</u> | | | | | | | | |
| Adams City | 40.5 | 72.8 | 27.2 | 31.2 | 32.1 | 38.7 | 39.8 | 39.5 |
| 8101 Ralston | 27.2 | 36.0 | 27.2 | 34.8 | * | * | * | * |
| 1633 Florence | 26.7 | 31.2 | 26.3 | 26.0 | 24.1 | 23.1 | * | * |
| 414 14th Street | 30.6 | 71.7* | 39.3* | 32.2 | 31.1 | 32.9 | * | * |
| CAMP/2105 Broadway | 28.7 | 45.1 | 37.2 | 27.4 | 27.3 | 31.6 | 29.3 | 38.4 |
| 1050 S. Broadway | 31.2 | 49.7* | 35.2 | 29.7 | 32.1 | 30.8 | 32.8 | 40.7 |
| 4857 S. Broadway | 23.1 | 39.9 | 27.4 | 23.6 | 22.0 | 13.1 | 21.0 | 29.8 |
| <u>24-Hour Maximum Values^b</u> | | | | | | | | |
| Adams City | 103.0 | 145.0 | 36.0 | 50.0 | 94.0 | 106.0 | 183 | 183 |
| 8101 Ralston | 56.0 | 57.0 | 59.0 | 78.0 | * | * | * | * |
| 1633 Florence | 94.0 | 77.0 | 56.0 | 49.0 | 84.0 | 64.0 | * | * |
| 414 14th Street | 66.0 | 123.0 | 59.0 | 54.0 | 118.0 | 67.0 | * | * |
| CAMP/2105 Broadway | 123.0 | 154.0 | 62.0 | 45.0 | 82.0 | 95.0 | 160 | 160 |
| 1050 S. Broadway | 68.0 | 64.0 | 57.0 | 51.0 | 118.0 | 70.0 | 176 | 176 |
| 4857 S. Broadway | 48.0 | 103.0 | 44.0 | 39.0 | 49.0 | 29.0 | 105 | 105 |

* Incomplete data set or samples no longer collected at this location

^a Annual arithmetic mean standard is $50 \mu\text{g}/\text{m}^3$

^b Maximum 24-hour standard is $150 \mu\text{g}/\text{m}^3$
 $\mu\text{g}/\text{m}^3$ Micrograms per cubic meter

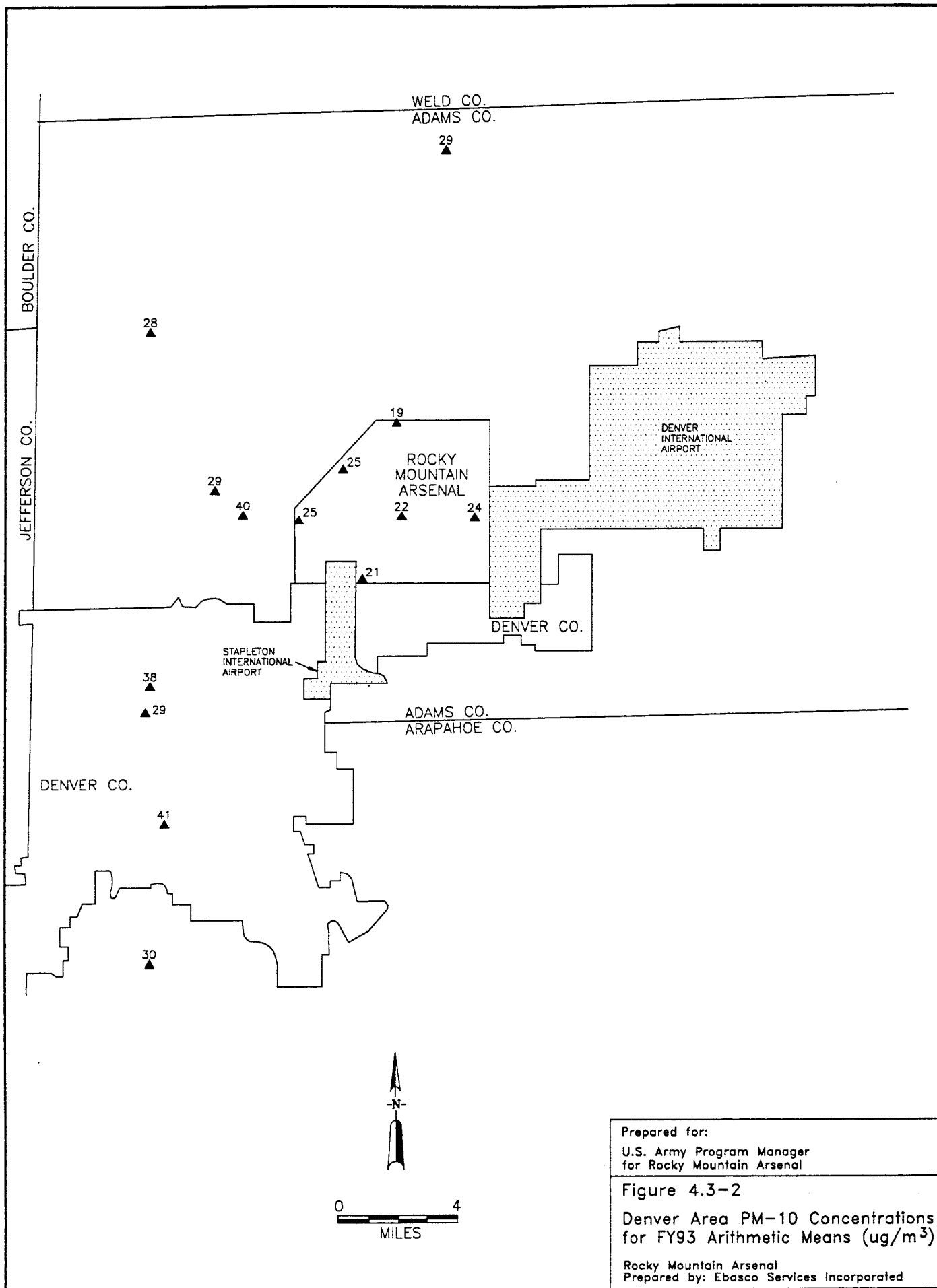
Phase 1 is from March 22 to December 12, 1988

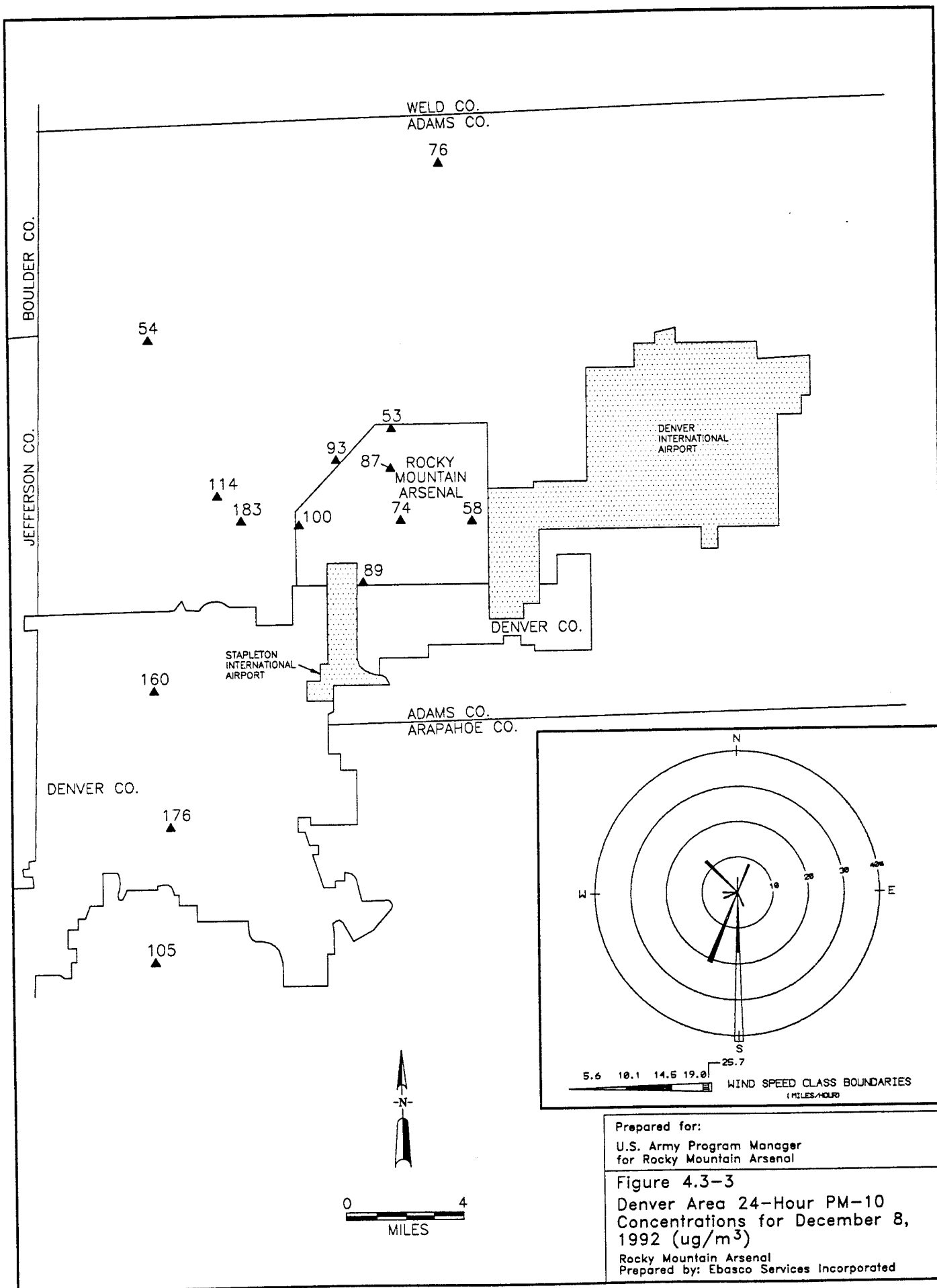
Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989

Phase 2, Stage 2 is from February 16 to May 5, 1989

Phase 3 is from May 6 to September 30, 1989

RMA/1009 10/19/94 2:34 pm ap





4.3.4 Summary of PM-10 Analysis

Strong gusty winds appeared to contribute to higher PM-10 values at RMA, but to a much lesser extent than for TSP. The highest PM-10 readings were observed during strong atmospheric inversion conditions over metropolitan Denver. There were no violations of the annual or 24-hour PM-10 standards during FY93 at RMA. As in the case of TSP, there were potentially two principal sources of PM-10 impacting RMA and its surroundings: (1) the influx of PM-10 emissions from metropolitan Denver, and (2) remediation activity sources at RMA, especially during high wind episodes. Unlike the trends observed in TSP concentrations, PM-10 concentration data do not indicate a detectable increase in average PM-10 levels during the remedial activity. This observation is likely the result of the lack of PM-10 monitoring stations immediately adjacent to remedial activities; nevertheless, most PM-10 remediation-generated emissions appeared to be localized and fell off significantly with distance from the activity. PM-10 sampling at RMA was increased during FY93 and will be included in the vicinity of future remedial activities. It is anticipated that these data will provide a more definitive assessment of PM-10 remediation impacts.

4.4 METALS

A total of six metallic ions were analyzed from TSP filters after the filters were weighed, including arsenic, cadmium, chromium, copper, lead, and zinc. Mercury was measured using Hydrar tubes and analyzed using the Rathje and Marcero method (Rathje and Marcero 1976).

4.4.1 Metals Monitoring Strategies

There were basically two components of the CAQMMP FY93 metals monitoring program. The first component was routine sampling every 6th day at AQ2, AQ3, AQ5, AQ5D, AQ6, AQ8, SQ1, and SQ2; every 12th day at FC1, FC2, and FC5; and monthly at FC3 and FC4. During the SQI special monitoring program, metals were also sampled every third day at SQ1, SQ2, AQ3 and AQ6. A summary of sampling frequency and data recovery for the routine monitoring program is provided in Tables 4.4-1a and 4.4-1b. It is noted that although metals analyses were obtained from TSP filters, the static electricity problems associated with TSP (and PM-10)

Table 4.4-1a Summary of Metals and Arsenic Monitoring for FY93

Page 1 of 1

| Station | Samples Scheduled | Samples Collected | Valid Samples | Field % Recovery | Overall % Recovery |
|-----------|----------------------|----------------------|------------------|---------------------|-----------------------|
| AQ2 | 61 | 57 | 55 | 93 | 90 |
| AQ3 | 61 | 58 | 56 | 95 | 92 |
| AQ5 | 61 | 60 | 58 | 98 | 95 |
| AQ6 | 61 | 58 | 56 | 95 | 92 |
| AQ8 | 61 | 59 | 57 | 97 | 93 |
| FC1 | 31 | 30 | 29 | 97 | 94 |
| FC2 | 31 | 29 | 27 | 94 | 87 |
| FC3 | 12 | 13 | 12 | 108 | 100 |
| FC4 | 12 | 12 | 11 | 100 | 92 |
| FC5 | 31 | 29 | 28 | 94 | 90 |
| SQ1 | 61 | 58 | 56 | 95 | 92 |
| SQ2 | 61 | 57 | 55 | 93 | 90 |
| Duplicate | 61 | 63 | 61 | 103 | 100 |
| Overall | 605 | 583 | 561 | 96 | 93 |

% - Percent

Table 4.4-1b Summary of Mercury Monitoring for FY93

Page 1 of 1

| Station | Samples Scheduled | Samples Collected | Valid Samples | Field % Recovery | Overall % Recovery |
|-----------|-------------------|-------------------|---------------|------------------|--------------------|
| AQ2 | 61 | 60 | 59 | 98 | 97 |
| AQ3 | 61 | 59 | 58 | 97 | 95 |
| AQ5 | 61 | 61 | 60 | 100 | 98 |
| AQ6 | 61 | 58 | 57 | 95 | 93 |
| AQ8 | 61 | 55 | 54 | 90 | 89 |
| FC1 | 31 | 32 | 32 | 103 | 103 |
| FC2 | 31 | 30 | 30 | 97 | 97 |
| FC3 | 12 | 13 | 13 | 108 | 108 |
| FC4 | 12 | 13 | 13 | 108 | 108 |
| FC5 | 31 | 32 | 32 | 103 | 103 |
| SQ1 | 61 | 59 | 58 | 98 | 95 |
| SQ2 | 61 | 58 | 58 | 95 | 95 |
| Duplicate | 61 | 61 | 58 | 100 | 95 |
| Overall | 605 | 591 | 582 | 98 | 96 |

% - Percent

* - Values over 100% reflect more samples taken than scheduled

analyses had no significant impact on the metals results. The second component involved high-event monitoring during specified high wind episodes (i.e., winds generally in excess of 10 mph and gusts in excess of 25 mph), or during specified contingency events. Twelve high-event episodes were conducted for the various metals components. Monitoring was conducted at various locations across RMA depending upon the meteorological conditions and selected source activities of special interest. Table 4.4-2 provides a detailed account of all metals high-event episodes including sampling dates, location, and potential source target areas; these included previously identified potential metals source areas such as Basins A and F, as well as special FY93 areas of interest such as the SQI facilities, Basin A excavation activities, Basin A soil-sifting activities, and brine loading activities at the railroad yard. In all cases, sampling was conducted for 24 hours with the exception of one 8-hour period during the soil-sifting operations. In most cases, sampling was augmented with mobile monitors to better surround the source of interest.

4.4.2 CMP FY93 Metals Monitoring Results

Average and short-term maximum concentrations of the full CAQMMP monitoring program for arsenic, cadmium, chromium, copper, lead, mercury and zinc are shown in Tables 4.4-3 and 4.4-4 with detailed sequential data shown in Appendix C. For those metals that contained both laboratory detectable and non-detectable values, average concentrations for the monitoring year were estimated by assigning a value of one-half of the detection limit to samples below the detection limit (this is shown in parentheses below the lower certified reporting limits provided in Table 4.4-3).

Results of the CAQMMP FY93 metals monitoring program were generally comparable to past CMP results. With several exceptions, metals levels were fairly uniform across RMA and for the most part were representative of their proximity to the Denver urban environment. For example, the highest annual average and 24-hour maximum concentrations for many of the metals occurred at AQ2 and AQ5 on the northwestern and southern boundaries, respectively. All maximum 24-hour concentrations were still well below concentrations measured during Basin F

Table 4.4-2 FY93 Metals High-Event Sampling Locations

| Date | Target Area | Duration (Hours) | Sampling Locations |
|---------|--|--------------------------|---|
| 1/22/93 | Basin F | 1300 - 1300 | AQ2, AQ3, AQ5, AQ6, FC2, FC4 |
| 3/4/93 | Submerged Quench Incinerator (Baseline) | 1200 - 1200 | AQ2, AQ3, AQ5, AQ6, FC2, M1E, SQ1, SQ2 |
| 4/10/93 | Basin A | 1500 - 1500 | AQ2, AQ3, AQ5, AQ6, M136E, M236W |
| 6/12/93 | Submerged Quench Incinerator (Trial Burn) | 0700 - 0700 | AQ3, AQ6, SQ1, SQ2, FC2, M125W |
| 6/24/93 | Basin A Excavation Activities | 0900 - 0900 | AQ5, AQ8, AQ9, M336E, M436W |
| 6/30/93 | Basin F | 1100 - 1100 | FC1, FC2, FC3, AQ11, AQ5, M226E |
| 7/15/93 | Basin A Soil Sifting | 0700 - 1500 (8 hours) | AQ1, AQ4, AQ5, AQ8, AQ9, AQ10, M601N, M701N |
| 8/29/93 | Submerged Quench Incinerator | 1300 - 1300 | AQ3, AQ6, SQ1, SQ2, FC2, M225W |
| 9/12/93 | Basin F Waste Pile | 1200 - 1200 | FC2, FC3, FC4, M326E, M426S, M526SE |
| 9/25/93 | Brine Loading at Railway | 1400 - 1400 | AQ1, AQ5, AQ7, M104NE, M204SE, M102W |
| 9/28/93 | Pond A Liner Cleaning | 1200 - 1200 | AQ2, AQ11, AQ12, FC1, FC4, FC5 |
| 9/30/93 | Section 35 Excavation Activities | 1200 - 1200 | AQ2, AQ7, AQ9, FC3, M135N, M134E |

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Table 4.4-2 FY93 Mercury High-Event Sampling Locations

| Date | Target Area | Duration (Hours) | Sampling Locations |
|---------|--|--------------------------|---|
| 1/22/93 | Basin F | 1300 - 1300 | AQ2, AQ3, AQ5, AQ6, FC2, FC4 |
| 3/4/93 | Submerged Quench Incinerator (Baseline) | 1200 - 1200 | AQ2, AQ3, AQ5, AQ6, FC2, SQ1, SQ2, M1E |
| 4/10/93 | Basin A | 1500 - 1500 | AQ2, AQ3, AQ5, AQ6, M136E, M236W |
| 6/12/93 | Submerged Quench Incinerator (Trial Burn) | 0700 - 0700 | AQ3, AQ6, SQ1, SQ2, FC2, M125W |
| 6/24/93 | Basin A Excavation Activities | 1300 - 1300 | AQ5, AQ8, AQ9, M336E, M436W |
| 6/30/93 | Basin F | 1100 - 1100 | FC1, FC2, FC3, FC4, AQ5, AQ11, M226E |
| 7/15/93 | Basin A Soil Sifting | 0700 - 1500 (8 hours) | AQ1, AQ4, AQ5, AQ8, AQ9, AQ10, M601N, M701N |
| 8/29/93 | Submerged Quench Incinerator | 1300 - 1300 | AQ3, AQ6, FC2, SQ1, SQ2, M225W |
| 9/12/93 | Basin F Waste Pile | 1200 - 1200 | FC2, FC3, FC4, M326E, M426S, M526SE |
| 9/25/93 | Brine Loading at Railyard | 1400 - 1400 | AQ1, AQ5, AQ7, M104NE, M204SE, M102W |
| 9/28/93 | Pond A Liner Cleaning | 1200 - 1200 | AQ2, AQ11, AQ12, FC1, FC4, FC5 |
| 9/30/93 | Section 35 Excavation Activities | 1200 - 1200 | AQ2, AQ7, AQ9, FC3, M135N, M134E |

Table 4.4-3 Summary of FY93 Average Metals, Arsenic, and Mercury Concentrations (µg/m³)

Page 1 of 1

| Station | Cadmium | Chromium | Copper | Lead | Zinc | Arsenic | Mercury |
|---------|---------------------|---------------------|--------|---------------------|--------|----------------------|---------------------|
| AQ2 | <0.0024 (0.0015) | <0.0123 (0.0062) | 0.1017 | <0.0245 (0.0152) | 0.0434 | <0.0008 (0.00056) | <0.2310 (0.1217) |
| AQ3 | <0.0024 (0.0013) | <0.0123 (0.0062) | 0.0816 | <0.0245 (0.0126) | 0.0244 | <0.0008 (0.00050) | <0.2310 (0.1198) |
| AQ5 | <0.0024 (0.0013) | <0.0123 (0.0062) | 0.1035 | <0.0245 (0.0145) | 0.0335 | <0.0008 (0.00054) | <0.2310 (0.1194) |
| AQ6 | <0.0024 (0.0013) | <0.0123 (0.0062) | 0.0799 | <0.0245 (0.0130) | 0.0245 | <0.0008 (0.00067) | <0.2310 (0.1245) |
| AQ8 | <0.0024 (0.0013) | <0.0123 (0.0062) | 0.0702 | <0.0245 (0.0132) | 0.0269 | <0.0008 (0.00055) | <0.2310 (0.1307) |
| FC1 | <0.0024 (0.0015) | <0.0123 (0.0062) | 0.0974 | <0.0245 (0.0148) | 0.0307 | <0.0008 (0.00066) | <0.2310 (0.1201) |
| FC2 | <0.0024 (0.0015) | <0.0123 (0.0062) | 0.0936 | <0.0245 (0.0149) | 0.0281 | <0.0008 (0.00059) | <0.2310 (0.1223) |
| FC3 | <0.0024 (0.0012) | <0.0123 (0.0062) | 0.0967 | <0.0245 (0.0122) | 0.0262 | <0.0008 (0.00054) | <0.2310 (0.1227) |
| FC4 | <0.0024 (0.0012) | <0.0123 (0.0062) | 0.1037 | <0.0245 (0.0122) | 0.0270 | <0.0008 (0.00044) | <0.2310 (0.1155) |
| FC5 | <0.0024 (0.0015) | <0.0123 (0.0062) | 0.1305 | <0.0245 (0.0147) | 0.0309 | <0.0008 (0.00063) | <0.2310 (0.1155) |
| SQ1 | <0.0024 (0.0014) | <0.0123 (0.0062) | 0.0605 | <0.0245 (0.0131) | 0.0252 | <0.0008 (0.00061) | <0.2310 (0.1174) |
| SQ2 | <0.0024 (0.0015) | <0.0123 (0.0062) | 0.0755 | <0.0245 (0.0137) | 0.0244 | <0.0008 (0.00054) | <0.2310 (0.1169) |

Note: < ### indicates concentration was less than the lower certified reporting limit

indicates computed average, when [Average] is less than the lower certified reporting limit.

µg/m³ Micrograms per cubic meter

RMA/0926 10/19/94 2:36 pm ap

Table 4.4-4 Summary of FY93 24-Hour Maximum Metals, Arsenic, and Mercury Concentrations (µg/m³) Page 1 of 1

| Station | Cadmium | Chromium | Copper | Lead | Zinc | Arsenic | Mercury |
|---------|---------|----------|--------|---------|--------|---------|---------|
| AQ2 | 0.0089 | <0.0123 | 0.2390 | 0.0692 | 0.1190 | 0.0018 | 0.3810 |
| AQ3 | 0.0062 | <0.0123 | 0.1400 | 0.0379 | 0.0853 | 0.0016 | 0.2940 |
| AQ5 | 0.0077 | <0.0123 | 0.1850 | 0.0614 | 0.1310 | 0.0018 | 0.3680 |
| AQ6 | 0.0073 | <0.0123 | 0.1710 | 0.0478 | 0.1000 | 0.0139 | 0.3940 |
| AQ8 | 0.0074 | <0.0123 | 0.1350 | 0.0507 | 0.1150 | 0.0035 | 0.6380 |
| FC1 | 0.0088 | <0.0123 | 0.2060 | 0.0524 | 0.1070 | 0.0033 | 0.2660 |
| FC2 | 0.0091 | <0.0123 | 0.2340 | 0.0496 | 0.0997 | 0.0023 | 0.3460 |
| FC3 | <0.0024 | <0.0123 | 0.1610 | <0.0245 | 0.0592 | 0.0012 | 0.2980 |
| FC4 | <0.0024 | <0.0123 | 0.1860 | <0.0245 | 0.0634 | <0.0008 | <0.2310 |
| FC5 | 0.0092 | <0.0123 | 0.2645 | 0.0546 | 0.1090 | 0.0021 | <0.2310 |
| SQ1 | 0.0095 | <0.0123 | 0.3010 | 0.0497 | 0.1000 | 0.0081 | 0.2620 |
| SQ2 | 0.0110 | <0.0123 | 0.1320 | 0.0518 | 0.1010 | 0.0047 | 0.2220 |

Note: <#### indicates no values were detected above the lower certified reporting limit
µg/m³ micrograms per cubic meter

remedial activities. Maximum average and 24-hour concentrations measured in FY93 for each metal component are discussed below.

The highest average concentration of arsenic, $0.00067 \mu\text{g}/\text{m}^3$, occurred just south of South Plants at AQ6 (this is below the certified detection level). The maximum 24-hour arsenic measurement was $0.0139 \mu\text{g}/\text{m}^3$ at AQ6 on May 1, 1993. Arsenic was infrequently measured at RMA, but detections usually occurred at several or more stations at the same time. For example, on May 1, seven of eleven metal sampling sites measured arsenic at varying levels across RMA. A case study of this day is further reviewed in Section 4.4.4.2.

The maximum average cadmium value was only $0.0015 \mu\text{g}/\text{m}^3$, measured at five of the ten stations that detected cadmium during FY93. This value is only slightly above half the detection limit; two stations did not detect any cadmium. The maximum 24-hour value was $0.0110 \mu\text{g}/\text{m}^3$, measured at SQ1 on June 12, 1993, a high-event monitoring day (see Section 4.4.3.4 for further discussion of this episode).

During FY93, chromium was not detected at any monitoring site. Chromium was detected during FY92 on seven sampling days out of the possible 55 routine and high-event sample dates.

The highest annual average copper level for FY93, $0.1305 \mu\text{g}/\text{m}^3$, occurred at FC4. The maximum 24-hour value for copper, $0.3010 \mu\text{g}/\text{m}^3$, was measured at SQ1 during a high-event episode (see Section 4.4.3.11).

The highest FY93 average lead concentration was $0.0152 \mu\text{g}/\text{m}^3$ (less than the detection limit of $0.0245 \mu\text{g}/\text{m}^3$) at AQ2 at the northwestern perimeter, where the 24-hour maximum lead value was $0.0692 \mu\text{g}/\text{m}^3$. These values were consistent with previous CMP measurements. The highest FY92 average lead value was $0.0156 \mu\text{g}/\text{m}^3$ at AQ1, while the maximum lead concentration was $0.0443 \mu\text{g}/\text{m}^3$ at AQ1.

The FY93 maximum average zinc concentration, $0.0434 \mu\text{g}/\text{m}^3$, was measured at AQ2. The FY93 maximum 24-hour concentration of $0.1310 \mu\text{g}/\text{m}^3$ was measured at AQ5. These values were very close to FY92 levels.

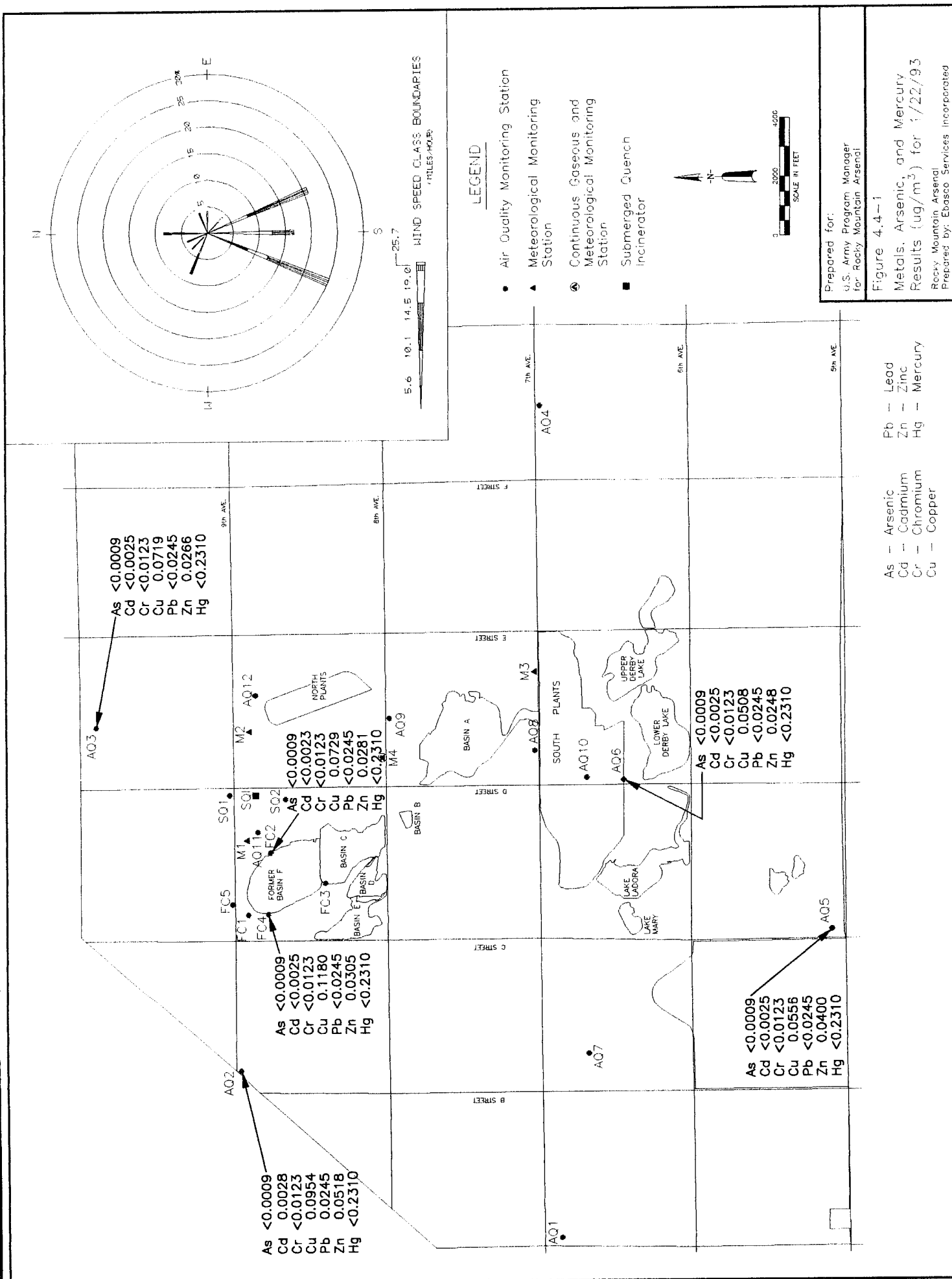
Mercury was detected on a number of individual sampling days during FY93, although on most days this element was below the detection limit. The highest average value was $0.1307 \mu\text{g}/\text{m}^3$, which is slightly above half the detection limit. This was measured at AQ8. On July 15, 1993, a maximum mercury level of $1.6710 \mu\text{g}/\text{m}^3$ was measured at AQ10, and a second maximum value of $0.7536 \mu\text{g}/\text{m}^3$ was measured at a mobile site (M601N) during a metals high-event episode (see Section 4.4.3.7). Note that these were one-time measurements and were not included in summary Table 4.4-4. It is possible that these values are the result of laboratory contamination, as several problems with filter blanks were noted during this period.

4.4.3 High-Event Monitoring Case Studies

During FY93, high-event monitoring was conducted at a number of potential metal source areas as shown in Table 4.4-2. These episodes are reviewed in the following sections. Several cases of interest are illustrated in figures showing the distribution of metal concentrations at the high-event sampling sites.

4.4.3.1 January 22, 1993—Basin F

This was a winter day that readily met high-event criteria for metals. Winds averaged more than 12 mph for the 24-hour period and peak gusts exceeded 33 mph. Prevailing strong winds were from the south-southwest through south-southeast, with lighter winds from other directions. Figure 4.4-1 shows the distribution of metals concentrations at six sites, including two downwind Basin F sites, FC2 and FC4; also shown is the 24-hour wind rose for this day. There were no measurable concentrations of chromium, lead, arsenic and mercury at RMA on this date. There was one low measurement of cadmium ($0.0028 \mu\text{g}/\text{m}^3$); the highest zinc concentration was $0.0518 \mu\text{g}/\text{m}^3$, both at AQ2. The highest copper concentration was $0.1180 \mu\text{g}/\text{m}^3$ at FC4 suggesting that the remediated Basin F area may be a very weak source of copper. Basin F sites averaged



slightly higher levels of copper as compared to the other RMA sites; however, AQ2 on the western boundary also showed levels comparable to the Basin F sites. In general, strong and gusty winds on this date indicated minimal metals impacts at the Basin F monitoring sites.

4.4.3.2 March 4, 1993—SQI

This was a late winter day with moderate winds, primarily from the northeast, and occasional gusts reaching more than 20 mph. Intense monitoring was conducted on this date for VOCs, SVOCS, and metals as a pre-SQI operational baseline evaluation. All metals concentrations at the SQI sites and at the mobile sites were well below normal levels except copper, which was $0.1120 \mu\text{g}/\text{m}^3$ downwind at SQ2; this was comparable to average values of copper in the Basin F area. There did not appear to be a significant metal source in the SQI area on this date. A more complete review of SQI metals results is provided in Section 4.10.

4.4.3.3 April 10, 1993—Basin A

This was an excellent metals high-event monitoring day in early spring. Sustained winds frequently exceeded 20 mph and peak gusts reached 37 mph. Winds were primarily from the northwest. Monitoring was conducted south of Basin A at AQ6, east and west of the Basin at two mobile sites, and at perimeter sites. All metals levels were measured below the certified limit with exception of copper and zinc. Zinc was highest at AQ2, $0.0310 \mu\text{g}/\text{m}^3$, suggesting off-post flow from the northwest, while copper was highest at AQ6, suggesting some influence off the old Basin floor. However, the level $0.0821 \mu\text{g}/\text{m}^3$ was below typical RMA values. The data suggest that Basin A, which has been historically considered a potential source of metals, evidenced little or no metals impacts on this windy day. Figure 4.4-2 shows the measured metals concentrations at each RMA sampling site and wind rose for the 24-hour period of monitoring.

4.4.3.4 June 12, 1993—SQI

On this date, the metals impacts surrounding the SQI were again evaluated in conjunction with other parameters. At this time, however, the SQI was fully operational and undergoing special stack emissions tests. The winds were primarily from the south and gusts were moderate,

occasionally reaching 30 mph. This was a typical late spring day with temperatures reaching 82°F. The highest levels of cadmium, copper, lead, and zinc were all measured at SQ2. However, with the winds blowing almost exclusively from the south (see Figure 4.4-3) and SQ1 levels consistently lower than SQ2, the implications are that the metals impacts did not come from SQI stack emissions; it is possible that Basin A or the South Plants complex to the south of SQ2 may have been a source of these emissions. It is also of interest that arsenic and mercury were sampled on this date at 12 monitoring sites. No arsenic was measured; however, one mercury measurement was sampled at AQ8 (just to the north of the South Plants). This level was 0.6380 $\mu\text{g}/\text{m}^3$ and was one of the higher mercury measurements during FY93.

4.4.3.5 June 24, 1993—Basin A

As a follow-up to the 12 June high-event episode, Basin A was intensely monitored on 24 June. Stations were located north, south, east, and west of the Basin. Winds were again from the south, and there were also strong east and north wind components. Average speeds were in excess of 11 mph and gusts reached 37 mph. The maximum temperature was 79° F. This time, there were no cadmium or mercury detections. Copper and zinc were highest along the southern RMA perimeter, and all metals levels were below detection in the Basin F area (see Figure 4.4-4). On this high-event day, Basin A did not appear to be a source of metals.

4.4.3.6 June 30, 1993—Basin F

This was an early summer day with temperatures above 90°F. Winds were moderate to strong with gusts reaching 30 mph. All metals concentrations were low or below the certified level with exception of mercury that was recorded at AQ2 at 0.247 $\mu\text{g}/\text{m}^3$. With mostly east and northeast winds on this date, a possible source of the mercury could have been the Basin F remediated area.

4.4.3.7 July 15, 1993—Basin A Soil Sifting

On this date, "soil sifting" activities of Basin A dirt were conducted in Building 545 in the South Plants. The area was surrounded with several permanent and mobile monitors, and sampling was

conducted for the 8 hours of operation. Most metals concentrations were low or below detection limits, although copper was slightly elevated, $0.1450 \mu\text{g}/\text{m}^3$ at the mobile M601N station.

On this metals high-event monitoring day, two high mercury readings were also recorded of $1.6710 \mu\text{g}/\text{m}^3$ at AQ10 and $0.7536 \mu\text{g}/\text{m}^3$ at mobile site M601N. It is possible that these values could be the result of laboratory contamination as several problems with filter blanks were noted during this period. The winds during the monitoring period were exclusively from the northwest through northeast. AQ10 was approximately 1,000 ft west-southwest of Building 545 and M601N was 100 feet east of the building. With respect to the soil sifting operations at Building 545, it would appear that mobile M701N, 100 feet to the south, would be more likely to experience impacts from the operation than AQ10 or M601N (see Figure 4.4-5). However, this sample site measured no detectable levels of mercury, as did none of the other sample sites. The contractor was not aware of any mercury in the sifted materials.

4.4.3.8 August 29, 1993—SQI

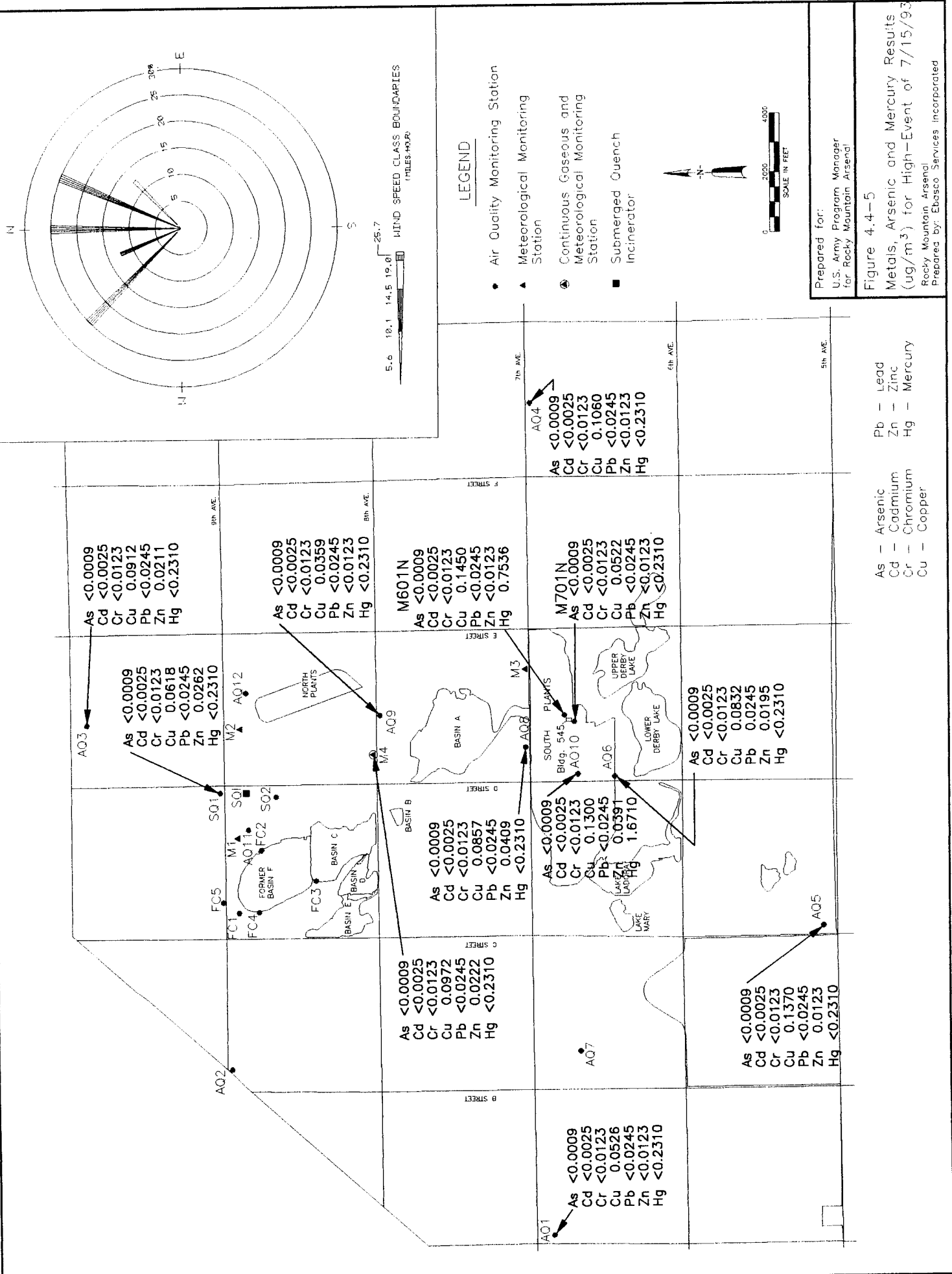
This was another SQI follow-on sampling assessment in conjunction with other parameters. Temperatures were in the low 80s and winds were primarily from the south quadrant at moderate speeds. All metals concentrations were below detection, with the exception of low copper and zinc levels (the only metals detected) at the RMA perimeter sites.

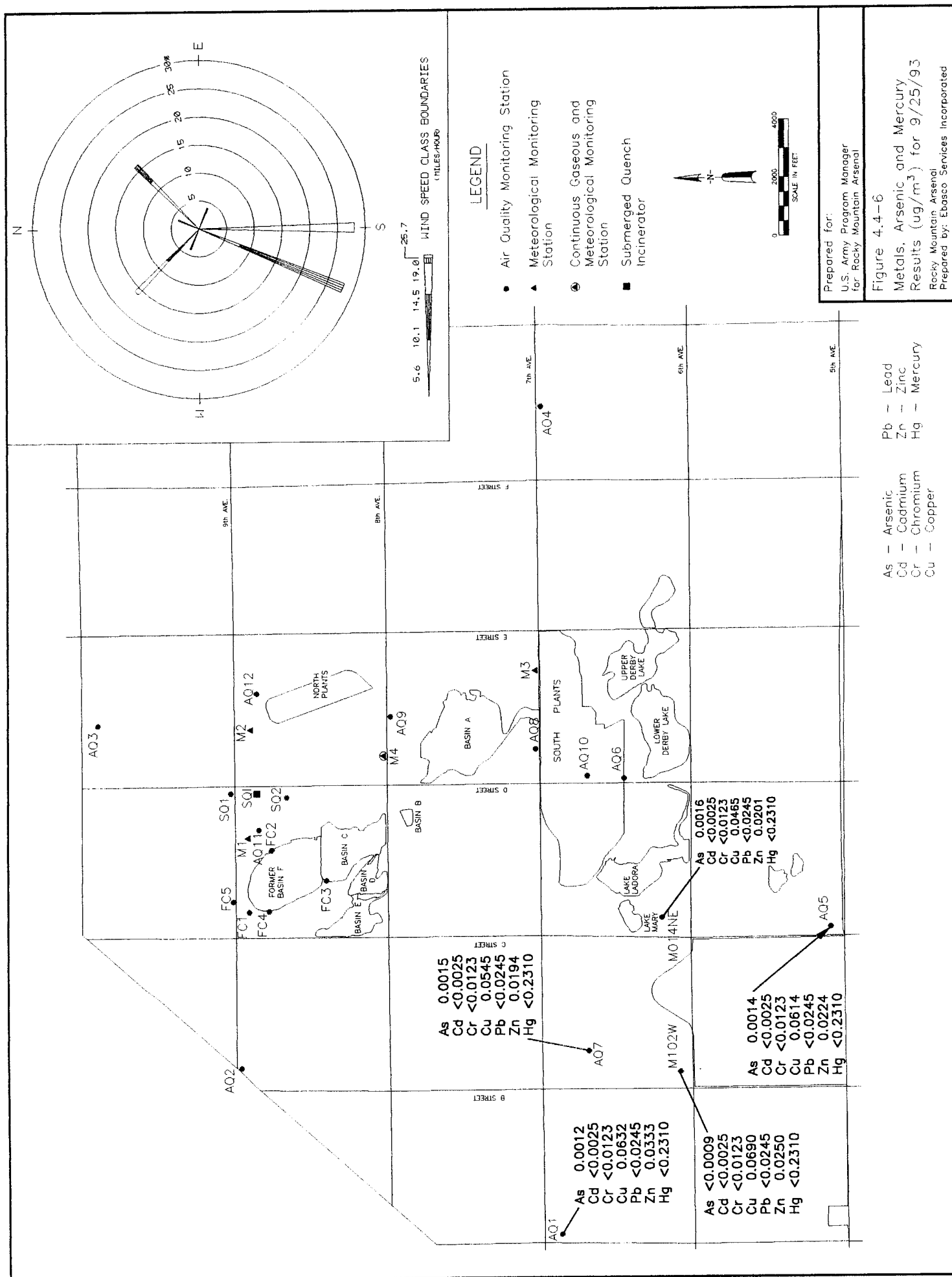
4.4.3.9 September 12, 1993—Basin F Waste Pile

This was a high event for mercury at the Basin F Waste pile. Winds were strong and persistent from the south-southwest. There were no mercury detections.

4.4.3.10 September 25, 1993—Brine Loading at Railroad Yard

Metals were sampled at the railroad yard complex on this early fall day. Winds were moderate from the south and south-southwest with lighter components from other directions; temperatures were in the 50s. Two mobile sites surrounded the activity. All concentrations were below the detection level or low; however, arsenic was detected at all of the RMA sites (see Figure 4.4-6).





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U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.4-6

Metals, Arsenic and Mercury
Results (ug/m³) for 9/25/93

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

As - Arsenic
Cd - Cadmium
Cr - Chromium
Cu - Copper
Pb - Lead
Zn - Zinc
Hg - Mercury

The arsenic levels were uniform at all six sampling sites including perimeter sites AQ1 and AQ5. With prevailing southerly winds, the arsenic source may have been off post.

4.4.3.11 September 28, 1993—Pond A Liner Cleaning

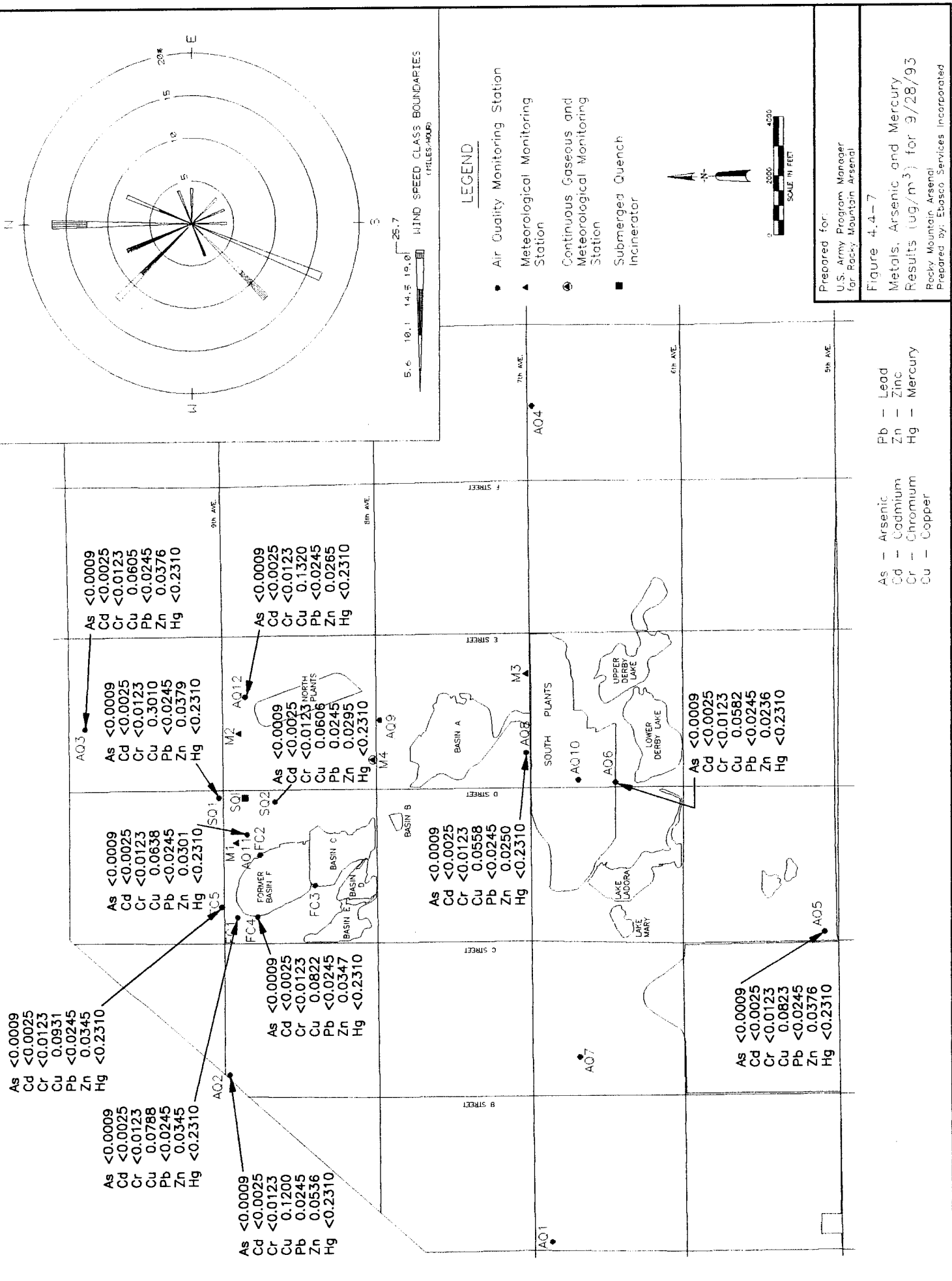
On this day, the liner over Pond A was cleaned, and metals and SVOC high-event monitoring was conducted. This was a mild fall day with variable winds at moderate speeds. Metals were monitored at 14 sites across RMA. All concentrations were low or below detection with exception of copper, which was measured at $0.3010 \mu\text{g}/\text{m}^3$ at site SQ1. This was the highest 24-hour copper measurement during the FY93 program. A likely explanation is that "Cutrine-Plus", a copper compound (copper alkanlamine complex), was used in the liner cleaning process. The high copper level detection at SQ1 illustrates the sensitivity of the sampling process. Ambient metals concentrations on this day are shown in Figure 4.4-7.

4.4.3.12 September 30, 1993—Section 36 Excavations

On this date, metals were monitored during excavation activities in Section 36. Winds were primarily from the south-southwest with gusts near 30 mph. This met typical high-event meteorological criteria for metals; however, all concentrations were low or below detection with exception of zinc, which was slightly above average at all stations. The highest zinc level of $0.0742 \mu\text{g}/\text{m}^3$ was at AQ1 with persistent south-southwest winds throughout the day, the implication is that the zinc source was from metropolitan Denver industrial activity.

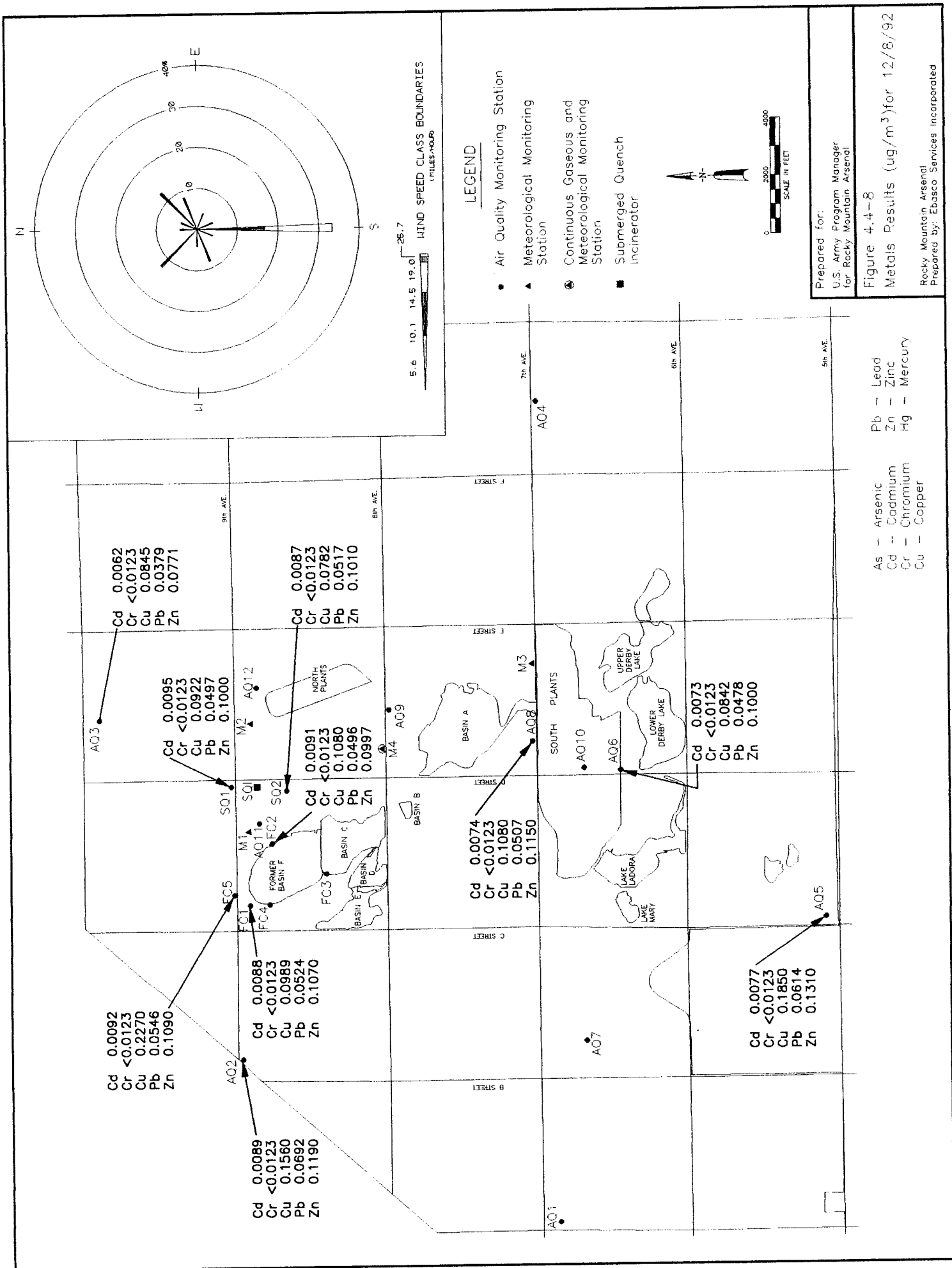
4.4.3.13 High-Event Summary Discussion

It would appear that in FY93, gusty winds did not contribute to increased metals levels as in the past under more intense RMA remediation activities. On occasion, several downwind sampling sites showed slightly higher levels than upwind sites, but ambient levels remained within typical RMA ranges. The fact that high metals concentrations were not detected during these high-event conditions or during selected construction remediation activities remains meaningful for the CAQMMP. One of the key objectives of the program is to detect potential health hazards as



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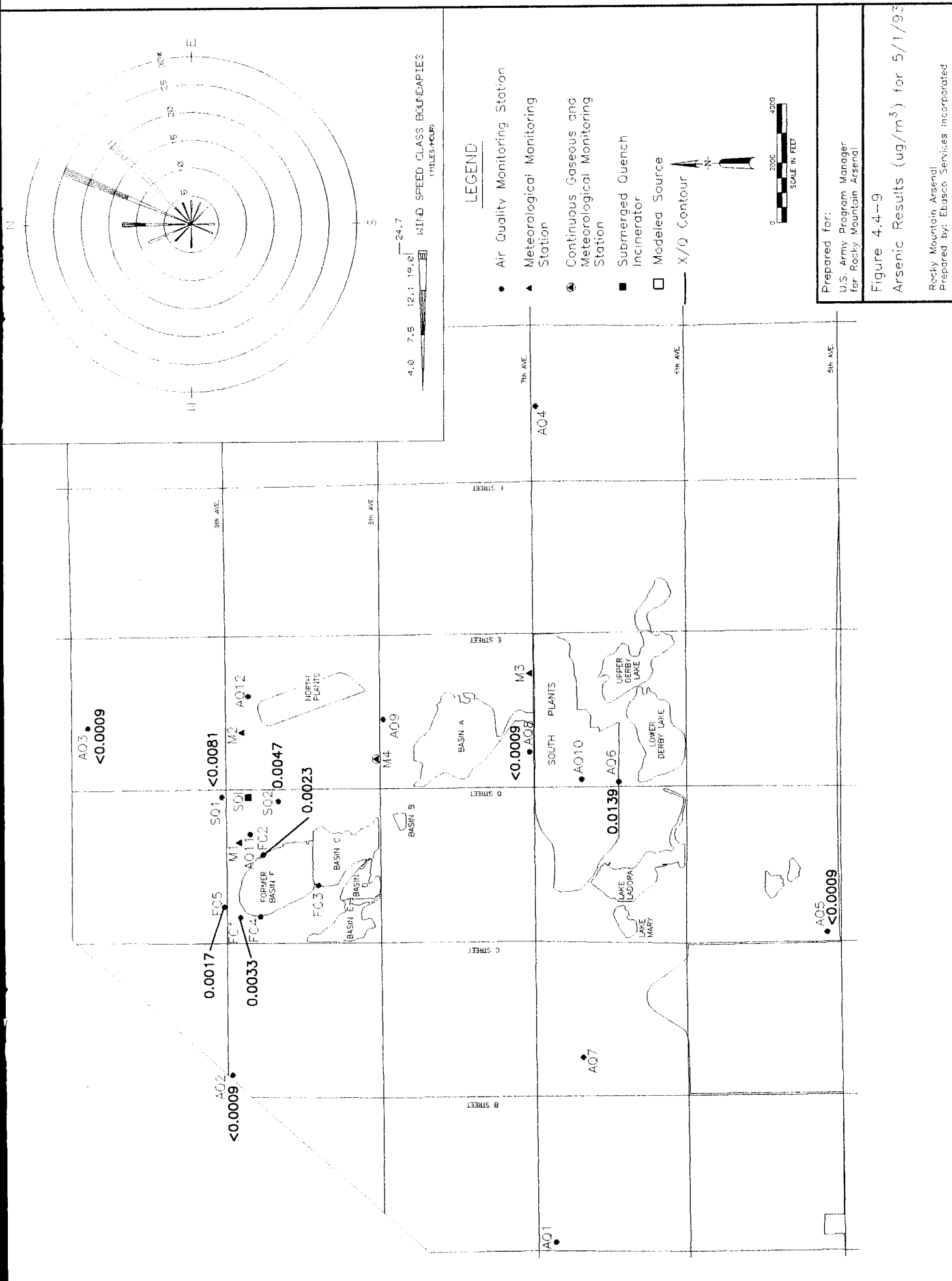
Figure 4.4-7
Metals, Arsenic and Mercury
Results (ug/m³) for 9/28/93
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



with other high wind days described above. A noted exception was arsenic, which was reported at seven of eleven metals sampling stations. Peak 24-hour arsenic concentrations of $0.0139 \mu\text{g}/\text{m}^3$, $0.0081 \mu\text{g}/\text{m}^3$, and $0.0047 \mu\text{g}/\text{m}^3$ were measured at monitoring sites AQ6, SQ1, and SQ2, respectively. These were the three highest arsenic concentrations measured during FY93. Figure 4.4-9 indicates the distribution of arsenic across RMA on this date. The arsenic source is difficult to identify. Although the most frequent concentrations were clustered near Basin F and the SQI, the highest concentration was several miles to the south at AQ6; also there were no arsenic detections at AQ8, which is located to the north of AQ6. Several similar days with broad distributions of arsenic concentrations (at lower levels) were observed during FY93.

4.4.5 Assessment of Basin F Metals Impacts

During the Basin F cleanup program, air monitoring stations immediately adjacent to the intense remedial activity were able to identify several metals elements (chromium, copper, zinc, and mercury) that reflected higher concentrations. Those stations at the RMA perimeter or farther from Basin F measured typical semi-urban levels reflecting impacts from the Denver metropolitan area. In the post-remedial period, metals impacts from the Basin F area have been minimal. On occasion, they have been slightly above Arsenal interior concentrations but well below levels detected during the Phase 1 and Phase 2 remediation periods. This has been well documented in the previous CMP/CAQMMP annual assessment reports. The trend continued in FY93. For comparison purposes, Table 4.4-5 provides a listing of average and 24-hour maximum metals levels for the Phase 1 and Phase 2 (1988-1989 remediation periods) and the current Phase 7 (FY93) period. Note that several metals (cadmium, lead, and arsenic) did not appear to reflect special impacts from the Basin F remediation. Copper and zinc reflect comparable annual averages in both the remedial and post-remedial phases. Most striking is that the 24-hour maximum values for chromium, copper, zinc, and mercury show significant decreases after completion of remediation activities. It is interesting that arsenic, which has not been identified as a Basin F source metal, measured a 24-hour maximum peak level (for the CAQMMP history) in FY93.



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for Rocky Mountain Arsenal

Figure 4.4--9

Arsenic Results ($\mu\text{g}/\text{m}^3$) for 5/1/93

Rocky Mountain Arsenal
Prepared by: Etasco Services Incorporated

Table 4.4-5 Summary of Average Metals and Mercury Results for Phases 1, 2-1, and 7 (µg/m³) Page 1 of 2

| Site | Phases | Cadmium | Chromium | Copper | Lead | Zinc | Arsenic | Mercury |
|---------|-----------------|---------|----------|--------|--------|--------|---------|---------|
| BF1/FC1 | Phase 1 | 0.0006 | 0.0128 | 0.1163 | 0.0148 | 0.0688 | 0.0008 | 1.6000 |
| | Phase 2 Stage 1 | 0.0009 | 0.0036 | 0.0569 | 0.0200 | 0.2243 | 0.0008 | 1.3000 |
| | Phase 7 | 0.0015 | ND | 0.0974 | 0.0148 | 0.0307 | 0.0007 | 0.1201 |
| BF2/FC2 | Phase 1 | 0.0008 | 0.0101 | 0.1516 | 0.0175 | 0.0727 | 0.0009 | 1.6000 |
| | Phase 2 Stage 1 | 0.0006 | 0.0037 | 0.0646 | 0.0220 | 0.0931 | 0.0011 | 1.2000 |
| | Phase 7 | 0.0015 | ND | 0.0936 | 0.0149 | 0.0281 | 0.0006 | 0.1223 |
| BF3/FC3 | Phase 1 | 0.0005 | 0.0176 | 0.0944 | 0.0138 | 0.0776 | 0.0007 | 1.5000 |
| | Phase 2 Stage 1 | 0.0007 | 0.0026 | 0.0566 | 0.0181 | 0.1952 | 0.0009 | 1.2000 |
| | Phase 7 | ND | ND | 0.0967 | ND | 0.0262 | 0.0005 | 0.1227 |
| BF4/FC4 | Phase 1 | 0.0006 | 0.0200 | 0.1258 | 0.0160 | 0.0779 | 0.0009 | 1.5000 |
| | Phase 2 Stage 1 | 0.0007 | 0.0028 | 0.0578 | 0.0202 | 0.1494 | 0.0009 | ND |
| | Phase 7 | ND | ND | 0.1037 | 0.1037 | 0.0270 | ND | ND |
| BF5 | Phase 1 | 0.0006 | 0.0026 | 0.0712 | 0.0144 | 0.0508 | 0.0012 | 1.2000 |
| | Phase 2 Stage 1 | 0.0006 | 0.0031 | 0.0522 | 0.0159 | 0.3260 | 0.0008 | ND |
| | Phase 7 | * | * | * | * | * | * | * |
| FC5 | Phase 1 | * | * | * | * | * | * | * |
| | Phase 2 Stage 1 | * | * | * | * | * | * | * |
| | Phase 7 | 0.0015 | ND | 0.1305 | 0.0147 | 0.0309 | 0.0006 | ND |

* Samples not collected at this site.

ND Not detected.

µg/m³ Micrograms per cubic meter

Note: Phase 1 is from March 22 to December 12, 1988.

Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989.

Phase 7 is from October 1, 1992 to September 30, 1993.

Table 4.4-5 Summary of 24-Hour Maximum Metals and Mercury Results for Phases 1, 2-1, and 7 (µg/m³) Page 2 of 2

| Site | Phases | Cadmium | Chromium | Copper | Lead | Zinc | Arsenic | Mercury |
|---------|-----------|---------|----------|--------|--------|--------|---------|---------|
| BF1/FC1 | Phase 1 | 0.0032 | 0.1520 | 0.3851 | 0.0258 | 0.5741 | 0.0025 | 4.7000 |
| | Phase 2-1 | 0.0037 | 0.0149 | 0.0839 | 0.0499 | 0.9703 | 0.0024 | 1.5000 |
| | Phase 7 | 0.0088 | ND | 0.2060 | 0.0524 | 0.1070 | 0.0033 | 0.2660 |
| BF2/FC2 | Phase 1 | 0.0149 | 0.1350 | 1.4980 | 0.0614 | 0.8106 | 0.0037 | 7.3000 |
| | Phase 2-1 | 0.0016 | 0.0083 | 0.1226 | 0.0428 | 0.4959 | 0.0031 | 3.0000 |
| | Phase 7 | 0.0091 | ND | 0.2340 | 0.0496 | 0.0997 | 0.0023 | 0.3460 |
| BF3/FC3 | Phase 1 | 0.0036 | 0.2083 | 0.5760 | 0.0292 | 0.5054 | 0.0018 | 7.3000 |
| | Phase 2-1 | 0.0016 | 0.0037 | 0.0858 | 0.0430 | 0.9151 | 0.0032 | 2.1000 |
| | Phase 7 | ND | ND | 0.1610 | ND | 0.0592 | 0.0012 | 0.2980 |
| BF4/FC4 | Phase 1 | 0.0037 | 0.2858 | 0.5349 | 0.0317 | 0.8651 | 0.0024 | 5.0000 |
| | Phase 2-1 | 0.0016 | 0.0038 | 0.1004 | 0.0436 | 1.1316 | 0.0026 | ND |
| | Phase 7 | ND | ND | 0.1860 | ND | 0.0634 | ND | ND |
| BF5 | Phase 1 | 0.0022 | 0.0087 | 0.1758 | 0.0434 | 0.2557 | 0.0136 | 2.1000 |
| | Phase 2-1 | 0.0016 | 0.0075 | 0.0975 | 0.0260 | 3.3576 | 0.0028 | ND |
| | Phase 7 | * | * | * | * | * | * | * |
| FC5 | Phase 1 | * | * | * | * | * | * | * |
| | Phase 2-1 | * | * | * | * | * | * | * |
| | Phase 7 | 0.0092 | ND | 0.2645 | 0.0546 | 0.1090 | 0.0021 | ND |

* Samples not collected at this site
 ND Not detected
 µg/m³ Micrograms per cubic meter

Note: Phase 1 is from March 22 to December 12, 1988
 Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989
 Phase 7 is from October 1, 1992 to September 30, 1993

4.4.6 Analysis of Metals Source Factors

Detectable concentrations of metals sampled in the RMA vicinity appeared to result from a variety of sources. Observed concentrations of several metals may have originated from deposition in the soil around Basin F and occasionally from Basin A; other higher concentrations may have resulted from industrial and transportation activities in the Denver metropolitan area. Inherent in all the observed metals concentrations at RMA are baseline levels typical of the area's semi-urban environment. With the exception of these industry-related components, the metals were probably transported primarily by wind-blown dust and soil particles. In a CMP study conducted in FY89 (RLSA 1990a), the relationship was assessed between total metals (the sum of concentrations of all target metals compounds) collected under the Basin F program and TSP levels at Basin F during all remediation phases for Site BF2. Although there were typical variations in this database (the overall correlation factor was 0.68), the general inference was that the higher the TSP levels, the higher the levels of metals collected in the sample filters.

An analysis of average and maximum metals concentrations for FY93 indicated that, as in past years, high metals concentrations at RMA monitoring sites resulted from a number of sources and occurred under meteorological conditions conducive to high particulate concentrations. In the CMP FY88 Data Assessment Report, Table 4.2-16 showed in almost every case that the highest metal concentrations for each element at each monitoring site were associated with average 24-hour winds in excess of 10 mph and wind gusts in excess of 25 mph (RLSA 1990b). In contrast, the majority of high metals episodes subsequent to FY89 have resulted from "brown cloud" inversion impacts from metropolitan Denver (WCC 1993). Very high TSP and PM-10 concentrations at all RMA monitoring sites were also recorded on these dates, even though wind speeds were not very strong or gusty. Prevailing winds in all cases were from the south or southwest, indicating that impacts were from metropolitan Denver. The December 8, 1992 case study in Section 4.4.4.1 illustrates this situation.

4.4.7 Seasonal Metals Impacts

While some higher metals concentrations continued to be a function of strong and gusty winds, especially downwind from local construction and remediation activities, metals concentrations were generally higher during fall and winter inversion periods than under high wind scenarios. A study conducted for the CAQMMP FY92 Annual Report (WCC 1993) for a 5-year period for stations AQ3 and AQ5 indicated that the highest lead, zinc, and arsenic levels for both the average and 24-hour maximum values occurred in the fall or winter seasons. During spring when winds were frequently the strongest, lower concentrations were experienced for most metals except chromium, which was detected primarily during the spring and summer. In contrast, copper concentrations were reported at their highest during summer.

4.4.8 Assessment of Metals Concentrations

Table 4.4-6 summarizes the locations of maximum metals concentrations from both the CMP/CAQMMP and Basin F monitoring programs throughout the FY88-FY93 periods. The EPA and the State of Colorado have promulgated standards and published air toxic guidelines only for lead, for which a Colorado monthly standard of $1.5 \mu\text{g}/\text{m}^3$ exists. During FY93, a maximum monthly average lead concentration of $0.0152 \mu\text{g}/\text{m}^3$ was measured at perimeter site AQ2; this was 1 percent of the EPA criterion.

4.4.9 Summary

Ambient metals concentrations distributed across RMA were generally proportional to TSP concentrations. In the past, relatively high metals concentrations were generally associated with strong and gusty winds; however, exceptions to this trend were apparent in FY93 due to a number of different influences including the status of ongoing remediation and the effectiveness of post-remediation mitigation efforts such as reseeded and soil stabilization treatments. The extreme maxima were associated with poor dispersion conditions and light winds during winter and occurred on days with high particulate levels over Denver. Basin F appeared to be a potential source of chromium, mercury, copper, and zinc during Phase 1 remedial activities, but concentrations were localized and decreased rapidly with distance from the source. After

Table 4.4-6 Maximum Metals Concentrations at Rocky Mountain Arsenal ($\mu\text{g}/\text{m}^3$) Page 1 of 1

| Metal | Maximum Long-term Average Concentration | Location | Phase | Maximum 24-Hour Concentration | Location | Phase |
|----------|---|----------|-------|-------------------------------|----------|-------|
| Arsenic | 0.0025 | AQ36 | P4 | 0.0139 | AQ6 | P7 |
| Cadmium | 0.0027 | FC2 | P3 | 0.0281 | AQ5 | P4 |
| Chromium | 0.0200 | BF4 | P1 | 0.2858 | BF4 | P1 |
| Copper | 0.1825 | AQ1 | P6 | 1.4980 | BF2 | P1 |
| Lead | 0.0573 | AQ7 | P3 | 0.0984 | AQ5 | P2-S1 |
| Zinc | 0.3260 | BF5 | P2-S1 | 3.3576 | BF5 | P2-S1 |
| Mercury | 1.8000 | BF2 | P1 | 7.3000 | BF2 | P1 |

P1 is from March 22 to December 12, 1998.

P2-S1 is from December 13, 1988 to February 15, 1989.

P3 is from May 5 to September 30, 1989.

P4 is from October 1, 1989 to September 30, 1990 (FY90).

P6 is from October 1, 1989 to September 30, 1991 (FY92).

P7 is from October 1, 1992 to September 30, 1993 (FY93).

remediation, metals levels in the vicinity of Basin F were typical of RMA baseline concentrations. None of the metals concentrations measured during the CAQMMP high-event days in FY93 indicated significant toxic or contamination levels.

4.5 ASBESTOS

Asbestos sampling was conducted at four sites on a monthly basis during FY93. Stations were selected based upon their proximity to potential sources of asbestos. Table 4.5-1 lists sample locations and data recovery rates.

In January of FY93, the target sample flowrate was lowered from 7 standard liters per minute (slpm) to 1.4 slpm due to the collection of excessive levels of particulate and other background material on the sample filter. Samples with fiber counts above 1,300 fibers per square millimeter (fmm^{-2}) and fiber counts from samples with greater than 50 percent of the filter area covered with particulates were reported as "cannot be read" (CBR). The reduction in flowrate appeared to have corrected this problem.

The estimated level of detection for this method is 7 fmm^{-2} (this corresponds to a concentration of 0.001 fibers per milliliter). All samples collected in FY93 had fiber counts below this minimum detection level. (Refer to Appendix D for complete listing of asbestos results). In summary, asbestos detections were minimal in FY88, FY89, FY90, and FY91 with no detections in FY92 and FY93. The concentration of asbestos in RMA ambient air is generally quite low.

4.6 VOLATILE ORGANIC COMPOUNDS

4.6.1 CAQMMP VOC Sampling, Analysis, and Reporting Strategies.

The CAQMMP Technical Plan calls for routine 6-day, monthly, and seasonal monitoring of VOCs; special weekly and monthly VOC monitoring at the air stripper; quarterly monitoring at the waste pile, Pond A, and tank vents; and an extensive VOC high-event monitoring program

Table 4.5-1 Synopsis of FY93 Asbestos Monitoring

Page 1 of 1

| Station | Samples Scheduled | Samples Collected | Percent Recovery |
|-------------------|-------------------|-------------------|------------------|
| AQ1 | 12 | 12 | 100 |
| AQ6 | 12 | 12 | 100 |
| AQ8 | 12 | 12 | 100 |
| AQ12 | 12 | 12 | 100 |
| Duplicate | 12 | 12 | 100 |
| Field Blank | 12 | 12 | 100 |
| <u>Trip Blank</u> | <u>12</u> | <u>12</u> | <u>100</u> |
| Total | 84 | 84 | 100 |

(discussed in Section 4.6.4). Details of the monitoring program have been previously provided in Section 3 and in Table 3.2-2. Table 4.6-1 provides a summary of the routine FY93 VOC program including the schedule and the recovery data at each of the sampling sites.

In FY91, the VOC sampling frequency was increased to every sixth day at 10 stations, and monthly and seasonally at five additional stations (including collocated stations). This sampling strategy has continued through FY93. As indicated in previous reports, low VOC certification ranges have, on occasion, resulted in values above the certification range. As a result, the laboratory has estimated VOC target analyte levels that were above the CRL using standard laboratory analytical methods of retrieval. The fact that certain target analytes were measured and analyzed above certification levels is a reflection of the limitations of the certification ranges and does not necessarily imply that these levels represented potential contamination risks. The available data, in fact, reflect the contrary. The data listing of VOCs collected in FY93 are provided in Appendix E. It is noted that those measurements listed as "greater than" (GT) do not meet IRDMIS certification criteria. In this section, however, all of the sample results are summarized and evaluated and the data above certified reporting limits have been incorporated into the analyses.

During FY93, VOC sampling included not only routine and high-event sampling, but real-time and integrated air sampling at Basin A vent sites, Pond A, and three storage tanks sites. In addition, VOC monitoring was conducted at the Basin A air stripper facility. These results are provided in Sections 4.8 and 4.9.

Two additional programs were conducted in FY93 involving VOCs. First, an SQI air quality support program was conducted in which VOC monitoring was a principal ingredient. Results of this investigation are provided in Section 4.10. Second, operational testing and air monitoring of the SVE system in the RMA motor pool area was performed from September 29 to October 1, 1993. Results of these tests are provided in Section 4.11.

Table 4.6-1 Summary of Volatile Organic Compound Monitoring for FY93

Page 1 of 1

| Station | Samples Scheduled | Samples Collected | Valid Samples | Field % Recovery | Overall % Recovery |
|-----------|-------------------|-------------------|---------------|------------------|--------------------|
| AQ1 | 4 | 5 | 5 | 125* | 125* |
| AQ2 | 61 | 55 | 52 | 90 | 85 |
| AQ3 | 61 | 53 | 50 | 87 | 82 |
| AQ4 | 4 | 3 | 3 | 75 | 75 |
| AQ5 | 61 | 56 | 52 | 92 | 85 |
| AQ6 | 61 | 53 | 51 | 87 | 84 |
| FC1 | 61 | 55 | 53 | 90 | 87 |
| FC2 | 61 | 53 | 50 | 87 | 82 |
| FC3 | 12 | 12 | 12 | 100 | 100 |
| FC4 | 12 | 12 | 10 | 100 | 83 |
| FC5 | 61 | 55 | 53 | 90 | 87 |
| SQ1 | 61 | 53 | 52 | 87 | 85 |
| SQ2 | 61 | 53 | 50 | 87 | 82 |
| Duplicate | 61 | 60 | 54 | 98 | 89 |
| Overall | 642 | 578 | 547 | 90 | 85 |

* Values over 100% reflect more samples taken than scheduled
% percent

During FY93, the problems with the CMO4 VOC method of field sampling using Tenax and Tenax/charcoal tubes and certification limitations have continued to be concerns for the CAQMMP. In recent years VOC sample flow rates and tube train configurations have been investigated and modified (WCC 1993). Nevertheless, two major concerns remained: (1) measurements above certification limits, as discussed above; and (2) breakthroughs of the more volatile organic compounds. These investigations were continued in FY93 and summarized in reports to PMRMA. In addition, alternative methods of VOC sampling, including, in particular, the use of passivated canisters, were investigated. During August and September of 1993, PMRMA authorized the testing of canisters in conjunction with Tenax/charcoal tubes at collocated sites. These results are currently being evaluated and it is anticipated that a concerted effort will be forthcoming to incorporate canisters into the VOC monitoring program. Discussions relating to VOC investigations and the canister tests are provided in Sections 4.6.8 and 4.6.9.

4.6.2 CAQMMP FY93 VOC Monitoring Results

VOCs were monitored in FY93 routinely and for high-event conditions. Tables 4.6-2 and 4.6-3 present summaries of results of the sampling of target VOC analytes at both fixed and mobile locations, including all high-event results. Table 4.6-2 shows the annual average and Table 4.6-3 shows the 24-hour maximum concentrations of each VOC at each site. For averaging purposes, those VOCs below the CRL were assigned a value of one-half of the detection limit (averages below the CRLs are shown in parentheses in Table 4.6-2).

Measured VOCs during FY93 were comparable to concentrations observed during previous years. However, several of the average and 24-hour maximum levels were higher than the previous year's results because estimates of concentrations above the certified limits were not reported during the FY92 period. This procedure was followed in all previous CMP assessments (FY88 through FY91) and was reestablished in order to maintain continuity of the program and also as to evaluate remedial progress. One additional explanation for higher 24-hour maximum VOC values in FY93 was the result of a single day's sampling measurements. As noted elsewhere in

Table 4.6-2 Summary of FY93 Average Volatile Organic Compound Concentrations (µg/m³)

| | 111TCE | 112TCE | 11DCLE | 12DCLE | 12DCE | BCHPD | C6H6 | CCL4 | CH2CL2 | CHCL3 |
|-----|--------|-------------------|-------------------|-------------------|-------------------|-------------------|-------|-------|---------|-------|
| AQ1 | 1.785 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.036) | <0.056 (0.028) | <0.056 (0.028) | 2.779 | 0.593 | 113.432 | 0.203 |
| AQ2 | 1.954 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.039) | <0.056 (0.028) | <0.056 (0.028) | 3.178 | 0.758 | 14.178 | 0.585 |
| AQ3 | 1.508 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.039) | <0.056 (0.028) | <0.056 (0.030) | 2.019 | 0.718 | 16.712 | 0.336 |
| AQ4 | 1.182 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.040) | <0.056 (0.028) | <0.056 (0.028) | 1.192 | 0.588 | 32.652 | 0.097 |
| AQ5 | 2.670 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.030) | <0.056 (0.028) | <0.056 (0.029) | 2.817 | 0.646 | 8.387 | 0.264 |
| AQ6 | 1.771 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.037) | <0.056 (0.028) | <0.056 (0.028) | 2.070 | 0.656 | 18.036 | 0.304 |
| FC1 | 1.945 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.032) | <0.056 (0.028) | <0.056 (0.040) | 3.304 | 0.781 | 8.484 | 1.115 |
| FC2 | 1.892 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.040) | <0.056 (0.028) | 0.085 | 2.651 | 0.743 | 7.667 | 1.755 |
| FC3 | 1.778 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.039) | <0.056 (0.028) | <0.056 (0.031) | 2.992 | 0.576 | 2.263 | 0.716 |
| FC4 | 2.200 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.047) | <0.056 (0.028) | <0.056 (0.028) | 3.339 | 0.748 | 2.355 | 0.556 |
| FC5 | 1.777 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.033) | <0.056 (0.028) | <0.056 (0.038) | 2.522 | 0.738 | 8.358 | 0.997 |
| SQ1 | 1.690 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.031) | <0.056 (0.028) | <0.056 (0.028) | 2.361 | 0.717 | 16.821 | 0.589 |
| SQ2 | 1.686 | <0.167 (0.084) | <0.056 (0.028) | <0.056 (0.030) | <0.056 (0.028) | <0.056 (0.028) | 2.178 | 0.716 | 13.586 | 0.565 |

µg/m³ - micrograms per cubic meter

< ### indicates concentration is less than lower certified reporting limit.

(###) indicates computed average when [average] is less than lower certified reporting limit.

| | | | | | |
|--------|---|--------------------------|--------|---|----------------------|
| 111TCE | - | 1,1,1-Trichloroethane | BCHPD | - | Bicycloheptadiene |
| 112TCE | - | 1,1,2-Trichloroethane | C6H6 | - | Benzene |
| 11DCLE | - | 1,1-Dichloroethane | CCL4 | - | Carbon Tetrachloride |
| 12DCLE | - | 1,2-Dichloroethane | CH2CL2 | - | Methylene Chloride |
| 12DCE | - | trans-1,2-Dichloroethene | CHCL3 | - | Chloroform |

Table 4.6-2 Summary of FY93 Average Volatile Organic Compound Concentrations (µg/m³)

| | CLC6H5 | DBCP | DCPD | DMDS | ETC6H5 | MEC6H5 | MIBK | TCLEE | TRCLE | XYLENES |
|-----|-------------------|-------------------|-------------------|-------------------|--------|--------|-------------------|-------|-------------------|---------|
| AQ1 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.698 | 5.278 | <0.741 (0.370) | 0.576 | <0.060 (0.040) | 3.319 |
| AQ2 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.774 | 5.980 | <0.741 (0.370) | 0.645 | <0.060 (0.043) | 3.712 |
| AQ3 | <0.060 (0.031) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.506 | 3.956 | <0.741 (0.370) | 0.394 | <0.060 (0.038) | 2.372 |
| AQ4 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.301 | 3.009 | <0.741 (0.370) | 0.424 | <0.060 (0.030) | 1.453 |
| AQ5 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.779 | 6.085 | <0.741 (0.370) | 0.782 | <0.060 (0.045) | 3.745 |
| AQ6 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.487 | 4.020 | <0.741 (0.370) | 0.436 | <0.060 (0.035) | 2.334 |
| FC1 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.687 | 5.149 | <0.741 (0.370) | 0.632 | <0.060 (0.047) | 3.306 |
| FC2 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.639 | 4.776 | <0.741 (0.370) | 0.640 | <0.060 (0.044) | 3.087 |
| FC3 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.924 | 6.383 | <0.741 (0.370) | 0.704 | 0.068 | 4.701 |
| FC4 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.857 | 6.068 | <0.741 (0.370) | 0.769 | 0.063 | 4.256 |
| FC5 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.611 | 4.578 | <0.741 (0.370) | 0.550 | <0.060 (0.042) | 2.914 |
| SQ1 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.656 | 4.775 | <0.741 (0.370) | 0.612 | <0.060 (0.048) | 2.979 |
| SQ2 | <0.060 (0.030) | <0.462 (0.231) | <0.412 (0.206) | <0.218 (0.114) | 0.517 | 3.960 | <0.741 (0.370) | 0.464 | <0.060 (0.037) | 2.388 |

µg/m³ - micrograms per cubic meter

< #### indicates concentration is less than the lower certified reporting limit

(####) indicates computed average when [average] is less than the lower certified reporting limit

| | | | | |
|----------|----------------------|---------|---|------------------------|
| CLC6H5 - | Chlorobenzene | MEC6H5 | - | Toluene |
| DBCP - | Dibromochloropropane | MIBK | - | Methyl isobutyl ketone |
| DCPD - | Dicyclopentadiene | TCLEE | - | Tetrachloroethene |
| DMDS - | Dimethyl disulfide | TRCLE | - | Trichloroethene |
| ETC6H5 - | Ethyl benzene | XYLENES | - | Total xylenes |

RMA/0975 10/19/94 2:58 pm bpw

Table 4.6-3 Summary of FY93 24-Hour Maximum Volatile Organic Compound Concentrations ($\mu\text{g}/\text{m}^3$)

| | 111TCE | 112TCE | 11DCLE | 12DCLE | 12DCE | BCHPD | C6H6 | CCL4 | CH2CL2 | CHCL3 |
|-----|--------|--------|--------|--------|--------|--------|--------|-------|---------|--------|
| AQ1 | 3.934 | <0.167 | <0.056 | 0.069 | <0.056 | <0.056 | 6.404 | 0.932 | 560.000 | 0.547 |
| AQ2 | 7.373 | <0.167 | <0.056 | 0.300 | <0.056 | <0.056 | 11.982 | 5.400 | 390.295 | 13.570 |
| AQ3 | 4.854 | <0.167 | <0.056 | 0.301 | 0.100 | 0.148 | 9.800 | 1.500 | 395.000 | 2.433 |
| AQ4 | 1.400 | <0.167 | <0.056 | 0.063 | <0.056 | <0.056 | 2.380 | 1.000 | 97.000 | 0.199 |
| AQ5 | 20.093 | <0.167 | <0.056 | 0.086 | <0.056 | 0.104 | 11.682 | 1.136 | 295.455 | 2.889 |
| AQ6 | 6.161 | <0.167 | <0.056 | 0.130 | <0.056 | <0.056 | 9.548 | 1.380 | 440.870 | 1.898 |
| FC1 | 6.160 | <0.167 | <0.056 | 0.074 | <0.056 | 0.205 | 31.430 | 2.100 | 293.023 | 5.070 |
| FC2 | 6.512 | <0.167 | <0.056 | 0.250 | <0.056 | 0.626 | 13.488 | 1.470 | 184.700 | 4.870 |
| FC3 | 2.830 | <0.167 | <0.056 | 0.155 | <0.056 | 0.061 | 7.143 | 0.823 | 11.556 | 1.790 |
| FC4 | 3.930 | <0.167 | <0.056 | 0.180 | <0.056 | <0.056 | 8.559 | 1.150 | 10.455 | 1.200 |
| FC5 | 5.090 | <0.167 | <0.056 | 0.140 | <0.056 | 0.168 | 9.211 | 2.350 | 273.300 | 2.910 |
| SQ1 | 13.130 | <0.167 | <0.056 | 0.078 | <0.056 | <0.056 | 33.830 | 2.050 | 305.000 | 6.110 |
| SQ2 | 5.652 | <0.167 | <0.056 | 0.094 | 0.200 | <0.056 | 11.304 | 2.100 | 254.579 | 3.650 |

489

 $\mu\text{g}/\text{m}^3$ micrograms per cubic meter

< ### indicates the concentration is less than the lower certified reporting limit

| | | | | | |
|--------|---|--------------------------|--------|---|----------------------|
| 111TCE | - | 1,1,1-Trichloroethane | BCHPD | - | Bicycloheptadiene |
| 112TCE | - | 1,1,2-Trichloroethane | C6H6 | - | Benzene |
| 11DCLE | - | 1,1-Dichloroethane | CCL4 | - | Carbon Tetrachloride |
| 12DCLE | - | 1,2-Dichloroethane | CH2CL2 | - | Methylene Chloride |
| 12DCE | - | trans-1,2-Dichloroethene | CHCL3 | - | Chloroform |

Table 4.6-3 Summary of FY93 24-Hour Maximum Volatile Organic Compound Concentrations (µg/m³)

Page 2 of 2

| | CLC6H5 | DBCP | DCPD | DMDS | ETC6H5 | MEC6H5 | MIBK | TCLEE | TRCLE | XYLENES |
|-----|--------|--------|--------|--------|--------|--------|--------|--------|-------|---------|
| AQ1 | <0.060 | <0.462 | <0.412 | <0.218 | 2.010 | 13.340 | <0.741 | 1.630 | 0.078 | 9.310 |
| AQ2 | <0.060 | <0.462 | <0.412 | <0.218 | 4.480 | 29.032 | <0.741 | 4.590 | 0.380 | 22.120 |
| AQ3 | <0.068 | <0.462 | <0.412 | <0.218 | 5.000 | 30.000 | <0.741 | 2.700 | 0.227 | 23.000 |
| AQ4 | <0.060 | <0.462 | <0.412 | <0.218 | 0.754 | 6.667 | <0.741 | 0.913 | 0.060 | 3.670 |
| AQ5 | <0.060 | <0.462 | <0.412 | <0.218 | 4.720 | 51.402 | <0.741 | 7.477 | 0.303 | 22.430 |
| AQ6 | <0.060 | <0.462 | <0.412 | <0.218 | 2.190 | 26.066 | <0.741 | 3.610 | 0.122 | 10.200 |
| FC1 | <0.060 | <0.462 | <0.412 | <0.218 | 4.060 | 32.093 | <0.741 | 5.630 | 0.250 | 18.605 |
| FC2 | <0.060 | <0.462 | <0.412 | <0.218 | 3.740 | 24.186 | <0.741 | 5.020 | 0.202 | 19.535 |
| FC3 | <0.060 | <0.462 | <0.412 | <0.218 | 2.150 | 13.839 | <0.741 | 1.390 | 0.115 | 11.700 |
| FC4 | <0.060 | <0.462 | <0.412 | <0.218 | 2.320 | 17.568 | <0.741 | 2.260 | 0.141 | 13.000 |
| FC5 | <0.060 | <0.462 | <0.412 | <0.218 | 4.350 | 23.830 | <0.741 | 3.300 | 0.178 | 20.560 |
| SQ1 | <0.060 | <0.462 | <0.412 | <0.218 | 14.120 | 87.560 | <0.741 | 16.780 | 0.657 | 51.390 |
| SQ2 | <0.060 | <0.462 | <0.412 | <0.218 | 4.230 | 29.800 | <0.741 | 3.370 | 0.180 | 19.810 |

µg/m³ micrograms per cubic meter

< #### indicates the concentration is less than the lower certified reporting limit

| | | | | | |
|--------|---|----------------------|---------|---|------------------------|
| CLC6H5 | - | Chlorobenzene | MEC6H5 | - | Toluene |
| DBCP | - | Dibromochloropropane | MIBK | - | Methyl isobutyl ketone |
| DCPD | - | Dicyclopentadiene | TCLEE | - | Tetrachloroethene |
| DMDS | - | Dimethyl disulfide | TRCLE | - | Trichloroethene |
| ETC6H5 | - | Ethyl benzene | XYLENES | - | Total xylenes |

this report, December 8, 1992 was a particularly bad air quality day in metropolitan Denver. Levels of suspended particulates, metals, and criteria gaseous pollutants reached their highest levels of the year, both in downtown Denver and at RMA. With respect to VOCs, seven analytes, 1,1,1-trichloroethane, benzene, toluene, ethylbenzene, trichloroethene, tetrachloroethene, and xylene, measured their highest 24-hour concentrations for the year on this date. Many concentrations were close to FY88 (Phase 1) Basin F levels, and two VOCs, ethylbenzene and xylene, exceeded the peak Basin F remedial period 24-hour measurements. The December 8, 1992 case study will be reported in further detail in Section 4.6.5.1.

Another VOC that reached its highest average and 24-hour levels during FY93 was methylene chloride. This compound was detected at high levels at a number RMA monitoring sites between April 16 and June 13, 1993. No specific location or activity at RMA has been identified as a potential source of the methylene chloride emissions. Potential on- and off-post sources have been investigated (and these investigations continue). See Section 4.6.5.2 for a more detailed discussion of the methylene chloride results.

Table 4.6-4 provides a listing of where long-term (annual) average and maximum 24-hour concentrations of individual VOC analytes were located in FY93. Although many of the highest average concentrations were measured at Basin F sites, average concentrations were, for the most part, uniformly distributed across RMA during FY93 (as shown in the composite results, Table 4.6-2). Those analytes previously identified as potential emissions from Basin F (benzene, bicycloheptadiene, chloroform, ethylbenzene, toluene, and dimethyldisulfide) were well below Basin F remediation levels and only slightly higher at Basin F than at other RMA sampling locations. 1,1,1-trichloroethane and tetrachloroethene were highest at the south perimeter (Station AQ5).

Many of the 24-hour maximum values occurred at RMA perimeter sites. However, six analytes (benzene, ethylbenzene, toluene, tetrachloroethene, trichloroethene, and total xylenes) were measured at the highest 24-hour maximum levels at SQ1. It is important to note that these

Table 4.6-4 Maximum Volatile Organic Compound Concentrations at Rocky Mountain Arsenal During CAQMMP, CMP, Basin F, and IRA-F Programs ($\mu\text{g}/\text{m}^3$) Page 1 of 1

| Analyte | Maximum 24-Hour Concentration | Location | Date | Maximum Annual Avg. Conc. | Location |
|--------------------------|-------------------------------------|----------|----------|---------------------------------|----------|
| 1,1,1-Trichloroethane | 20.093 | AQ5 | 12/08/92 | 2.670 | AQ5 |
| 1,1,2-Trichloroethane | < 0.170 | NA | NA | < 0.170 | NA |
| 1,1-Dichloroethane | < 0.056 | NA | NA | < 0.056 | NA |
| 1,2-Dichloroethane | 0.301 | AQ3 | 01/07/93 | < 0.056 | NA |
| Trans-1,2-Dichloroethene | 0.200 | SQ2 | 06/15/93 | < 0.056 | NA |
| Bicycloheptadiene | 1.440 | FC2D | 08/29/93 | 0.085 | FC2 |
| Benzene | 33.830 | SQ1 | 12/08/92 | 3.339 | FC4 |
| Carbon Tetrachloride | 5.400 | AQ2 | 09/04/93 | 0.781 | FC1 |
| Methylene Chloride | 560.000 | AQ1 | 06/03/93 | 18.036* | AQ6* |
| Chloroform | 13.570 | AQ2 | 09/04/93 | 1.755 | FC2 |
| Chlorobenzene | 0.068 | AQ3 | 02/24/93 | < 0.060 | NA |
| Dibromochloropropane | 0.980 | FC1D | 08/23/93 | < 0.463 | NA |
| Dicyclopentadiene | 0.450 | FC2D | 08/29/93 | < 0.410 | NA |
| Dimethyldisulfide | < 0.220 | NA | NA | < 0.220 | NA |
| Ethylbenzene | 14.120 | SQ1 | 12/08/92 | 0.924 | FC3 |
| Toluene | 87.560 | SQ1 | 12/08/92 | 6.383 | FC3 |
| Methyl isobutyl ketone | < 0.740 | NA | NA | < 0.740 | NA |
| Tetrachloroethene | 16.780 | SQ1 | 12/08/92 | 0.782 | AQ5 |
| Trichloroethene | 0.657 | SQ1 | 12/08/92 | 0.068 | FC3 |
| Xylenes | 51.390 | SQ1 | 12/08/92 | 4.701 | FC3 |

$\mu\text{g}/\text{m}^3$ - micrograms per cubic meter

< ### - indicates average concentration is less than the lower certified reporting limit

NA - not applicable

* - AQ1 and AQ4 measured higher averages (see Table 4.6-2); however, these were based on limited samples, 5 and 3 cases respectively, while the AQ6 annual average was based on 71 cases.

analytes were all measured on December 8, 1992, prior to the operation of the SQI. As will be discussed in Section 4.6.5, these concentrations were part of the severe brown cloud influx from metropolitan Denver resulting from an intense surface inversion over the area. Two VOC analytes unique to Basin F, bicycloheptadiene and dicyclopentadiene, measured their highest FY93 24-hour concentrations at Basin F station FC2D; however, concentration levels were one to two orders of magnitude below peak levels detected during Basin F remediation. Comparisons of Basin F remediation levels and FY93 levels for VOCs will be discussed further in the next section.

4.6.3 Basin F Remediation Progress

The CMP/CAQMMP has monitored Basin F post-remedial air quality impacts continuously since the conclusion of the Basin F cleanup program in May of 1989. Routine VOC monitoring has been conducted each year at five or more sampling sites surrounding the remediation area. Seven remedial and post-remedial evaluation phases have been defined as shown in Table 4.1-1; FY93 is Phase 7 of this evaluation. Previous annual CMP reports have listed the sequential data for each post-remedial phase to assess remedial progress. It is now fairly well established that those VOC contaminants identified at higher concentrations during remediation have been reduced to near, or very slightly above, regional background levels. Changes in VOC ambient air concentration levels over the past several years have reflected, for the most part, regional air quality variations and other minor remediation activities at RMA, rather than continuing influences of Basin F. For the purpose of brevity in this report, Table 4.6-5 compares the annual average and 24-hour maximum VOC concentrations at Basin F stations during the Phase 1 and Phase 2 (Stage 1) remediation periods with the most recent Phase 7 post-remediation period (rather than including data from all of the other post-remedial phases). For those VOCs that have been identified as potential Basin F source emissions, FY93 concentration levels are significantly reduced or below detection levels. This is demonstrated graphically in Figure 4.6-1, which shows comparisons between Phase 1 remediation average concentrations for bicycloheptadiene, chloroform, dicyclopentadiene, dimethylsulfide, and toluene, and FY93 Phase 7 results. Also noted from these data (Table 4.6-5) is the decrease in concentration levels (even during Phase 1)

Table 4.6-5 Summary of Average Volatile Organic Compound Results for Phases 1, 2-1, and 7 (µg/m³) Page 1 of 2

| Site | Phases | 111TCE | 112TCE | 11DCLE | 12DCLE | 12DCE | BCHPD | C6H6 | CCL4 | CH2CL2 | CHCL3 |
|---------|-----------------|--------|--------|--------|--------|-------|-------|------|------|--------|-------|
| BF1/FC1 | Phase 1 | 9.44 | ND | ND | 0.04 | ND | 1.82 | 2.45 | 0.65 | 10.48 | 2.14 |
| | Phase 2 Stage 1 | 5.50 | ND | ND | ND | ND | 0.25 | 3.97 | 1.04 | 4.01 | 0.77 |
| | Phase 7 | 1.94 | ND | ND | 0.03 | ND | 0.04 | 3.30 | 0.78 | 8.48 | 1.12 |
| BF2/FC2 | Phase 1 | 8.05 | ND | 0.02 | 0.30 | ND | 9.96 | 3.49 | 0.61 | 7.57 | 6.95 |
| | Phase 2 Stage 1 | 5.85 | ND | ND | 0.06 | ND | 0.88 | 3.79 | 0.92 | 2.52 | 1.89 |
| | Phase 7 | 1.89 | ND | ND | 0.04 | ND | 0.09 | 2.65 | 0.74 | 7.67 | 1.76 |
| BF3/FC3 | Phase 1 | 6.94 | ND | ND | 0.03 | ND | 0.49 | 1.94 | 0.52 | 7.17 | 0.80 |
| | Phase 2 Stage 1 | 5.03 | ND | ND | 0.02 | ND | 0.09 | 3.32 | 0.88 | 3.52 | 0.48 |
| | Phase 7 | 1.78 | ND | ND | 0.04 | ND | 0.03 | 2.99 | 0.58 | 2.26 | 0.72 |
| BF4/FC4 | Phase 1 | 10.53 | 0.05 | ND | ND | ND | 0.55 | 2.54 | 0.42 | 6.36 | 0.91 |
| | Phase 2 Stage 1 | 5.22 | ND | ND | 0.03 | ND | 0.13 | 3.78 | 0.75 | 3.64 | 0.39 |
| | Phase 7 | 2.20 | ND | ND | 0.05 | ND | ND | 3.34 | 0.75 | 2.36 | 0.56 |
| BF5 | Phase 1 | 8.54 | ND | ND | 0.03 | ND | 0.43 | 2.19 | 0.53 | 7.77 | 0.88 |
| | Phase 2 Stage 1 | 4.18 | ND | ND | 0.02 | ND | 0.47 | 2.45 | 0.97 | 2.12 | 0.33 |
| | Phase 7 | * | * | * | * | * | * | * | * | * | * |
| FC5 | Phase 1 | * | * | * | * | * | * | * | * | * | * |
| | Phase 2 Stage 1 | * | * | * | * | * | * | * | * | * | * |
| | Phase 7 | 1.78 | ND | ND | 0.03 | ND | 0.04 | 2.52 | 0.74 | 8.36 | 1.00 |

* Samples not collected at this site.

ND Not detected.

µg/m³ Micrograms per cubic meter

Note: Phase 1 is from March 22 to December 12, 1988.

Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989.

Phase 7 is from October 1, 1992 to September 30, 1993.

| | | | | | |
|--------|---|--------------------------|--------|---|----------------------|
| 111TCE | - | 1,1,1-Trichloroethane | BCHPD | - | Bicycloheptadiene |
| 112TCE | - | 1,1,2-Trichloroethane | C6H6 | - | Benzene |
| 11DCLE | - | 1,1-Dichloroethane | CCL4 | - | Carbon Tetrachloride |
| 12DCLE | - | 1,2-Dichloroethane | CH2CL2 | - | Methylene Chloride |
| 12DCE | - | trans-1,2-Dichloroethene | CHCL3 | - | Chloroform |

Table 4.6-5 Summary of Average Volatile Organic Compounds Results for Phases 1, 2-1, and 7 (µg/m³) Page 2 of 4

| Site | Phases | CLC6H5 | DBCP | DCPD | DMDS | ETC6H5 | MEC6H5 | MIBK | TCLEE | TRCLE | XYLENES |
|---------|-----------------|--------|------|------|------|--------|--------|------|-------|-------|---------|
| BF1/FC1 | Phase 1 | 0.05 | * | 1.60 | 7.86 | 1.09 | 9.34 | ND | 2.12 | 0.26 | 5.61 |
| | Phase 2 Stage 1 | 0.02 | * | 0.45 | ND | 1.59 | 5.24 | ND | 1.55 | 0.17 | 6.75 |
| | Phase 7 | ND | ND | ND | ND | 0.69 | 5.45 | ND | 0.63 | 0.05 | 3.31 |
| BF2/FC2 | Phase 1 | 0.12 | * | 4.53 | 5.34 | 1.58 | 20.95 | ND | 5.07 | 0.12 | 8.57 |
| | Phase 2 Stage 1 | 0.03 | * | 0.32 | 0.02 | 1.49 | 5.05 | ND | 1.57 | 0.11 | 6.35 |
| | Phase 7 | ND | ND | ND | ND | 0.64 | 4.78 | ND | 0.64 | 0.04 | 3.09 |
| BF3/FC3 | Phase 1 | ND | * | 0.52 | 0.50 | 0.81 | 8.27 | ND | 1.13 | 0.21 | 4.23 |
| | Phase 2 Stage 1 | ND | * | 0.18 | ND | 1.45 | 4.81 | ND | 1.37 | 0.15 | 6.40 |
| | Phase 7 | ND | ND | ND | ND | 0.92 | 6.38 | ND | 0.70 | 0.07 | 4.20 |
| BF4/FC4 | Phase 1 | 0.03 | * | 0.48 | 0.91 | 0.92 | 8.51 | ND | 1.13 | 0.29 | 5.02 |
| | Phase 2 Stage 1 | ND | * | 0.21 | ND | 1.94 | 5.02 | ND | 1.59 | 0.17 | 8.17 |
| | Phase 7 | ND | ND | ND | ND | 0.86 | 6.07 | ND | 0.77 | 0.06 | 4.26 |
| BF5 | Phase 1 | 0.02 | * | 0.34 | 0.92 | 0.77 | 5.47 | ND | 1.21 | 0.24 | 3.94 |
| | Phase 2 Stage 1 | ND | * | 0.06 | ND | 0.85 | 3.30 | ND | 0.94 | 0.09 | 3.72 |
| | Phase 7 | * | * | * | * | * | * | * | * | * | * |
| FC5 | Phase 1 | * | * | * | * | * | * | * | * | * | * |
| | Phase 2 Stage 1 | * | * | * | * | * | * | * | * | * | * |
| | Phase 7 | ND | ND | ND | ND | 0.61 | 4.58 | ND | 0.55 | 0.04 | 2.91 |

* Samples not collected at this site.
 ND Not detected.
 µg/m³ Micrograms pwer cubic meter
 Note: Phase 1 is from March 22 to December 12, 1988.
 Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989.
 Phase 7 is from October 1, 1992 to September 30, 1993.

CLC6H5 - Chlorobenzene
 DBCP - Dibromochloropropane
 DCPD - Dicyclopentadiene
 DMDS - Dimethyldisulfide
 ETC6H5 - Ethylbenzene
 MEC6H5 - Methyl isobutyl ketone
 MIBK - Tetrachloroethene
 TCLEE - Trichloroethene
 TRCLE - Total Xylenes
 XYLENES - Total Xylenes

Table 4.6-5 Summary of 24-Hour Maximum Volatile Organic Compounds Results for Phases 1, 2-1, and 7 (µg/m³) Page 4 of 4

| Site | Phases | CLC6H5 | DBCP | DCPD | DMDS | ETC6H5 | MEC6H5 | MIBK | TCLEE | TRCLE | XYLENES |
|---------|-----------------|--------|------|-------|-------|--------|--------|------|-------|-------|---------|
| BF1/FC1 | Phase 1 | 0.45 | * | 4.58 | 36.72 | 4.95 | 35.55 | ND | 16.02 | 1.47 | 25.94 |
| | Phase 2 Stage 1 | 1.55 | * | 3.36 | ND | 4.38 | 13.32 | ND | 3.19 | 0.57 | 17.26 |
| | Phase 7 | ND | ND | ND | ND | 4.06 | 32.09 | ND | 5.63 | 0.25 | 18.61 |
| BF2/FC2 | Phase 1 | 0.80 | * | 29.12 | 24.06 | 8.91 | 90.10 | ND | 23.17 | 0.78 | 49.91 |
| | Phase 2 Stage 1 | 0.06 | * | 2.48 | 0.08 | 4.20 | 11.43 | ND | 4.48 | 0.34 | 15.77 |
| | Phase 7 | ND | ND | ND | ND | 3.74 | 24.19 | ND | 5.02 | 0.20 | 19.54 |
| BF3/FC3 | Phase 1 | ND | * | 6.04 | 4.97 | 3.98 | 49.41 | ND | 2.30 | 1.00 | 10.51 |
| | Phase 2 Stage 1 | ND | * | 1.27 | ND | 3.75 | 11.14 | ND | 2.83 | 0.34 | 15.57 |
| | Phase 7 | ND | ND | ND | ND | 2.15 | 13.84 | ND | 1.39 | 0.16 | 11.70 |
| BF4/FC4 | Phase 1 | 0.39 | * | 3.74 | 8.07 | 2.39 | 49.91 | ND | 3.40 | 1.81 | 10.39 |
| | Phase 2 Stage 1 | ND | * | 1.37 | ND | 8.97 | 19.52 | ND | 5.52 | 0.65 | 28.28 |
| | Phase 7 | ND | ND | ND | ND | 2.32 | 17.57 | ND | 2.26 | 0.14 | 13.00 |
| BF5 | Phase 1 | 0.05 | * | 2.15 | 3.95 | 2.48 | 14.49 | ND | 3.32 | 1.00 | 10.52 |
| | Phase 2 Stage 1 | ND | * | 0.60 | ND | 1.99 | 8.42 | ND | 2.58 | 0.17 | 10.39 |
| | Phase 7 | * | * | * | * | * | * | * | * | * | * |
| FC5 | Phase 1 | * | * | * | * | * | * | * | * | * | * |
| | Phase 2 Stage 1 | * | * | * | * | * | * | * | * | * | * |
| | Phase 7 | ND | ND | ND | ND | 4.35 | 23.83 | ND | 3.30 | 0.18 | 20.36 |

* Samples not collected at this site.

ND Not detected.

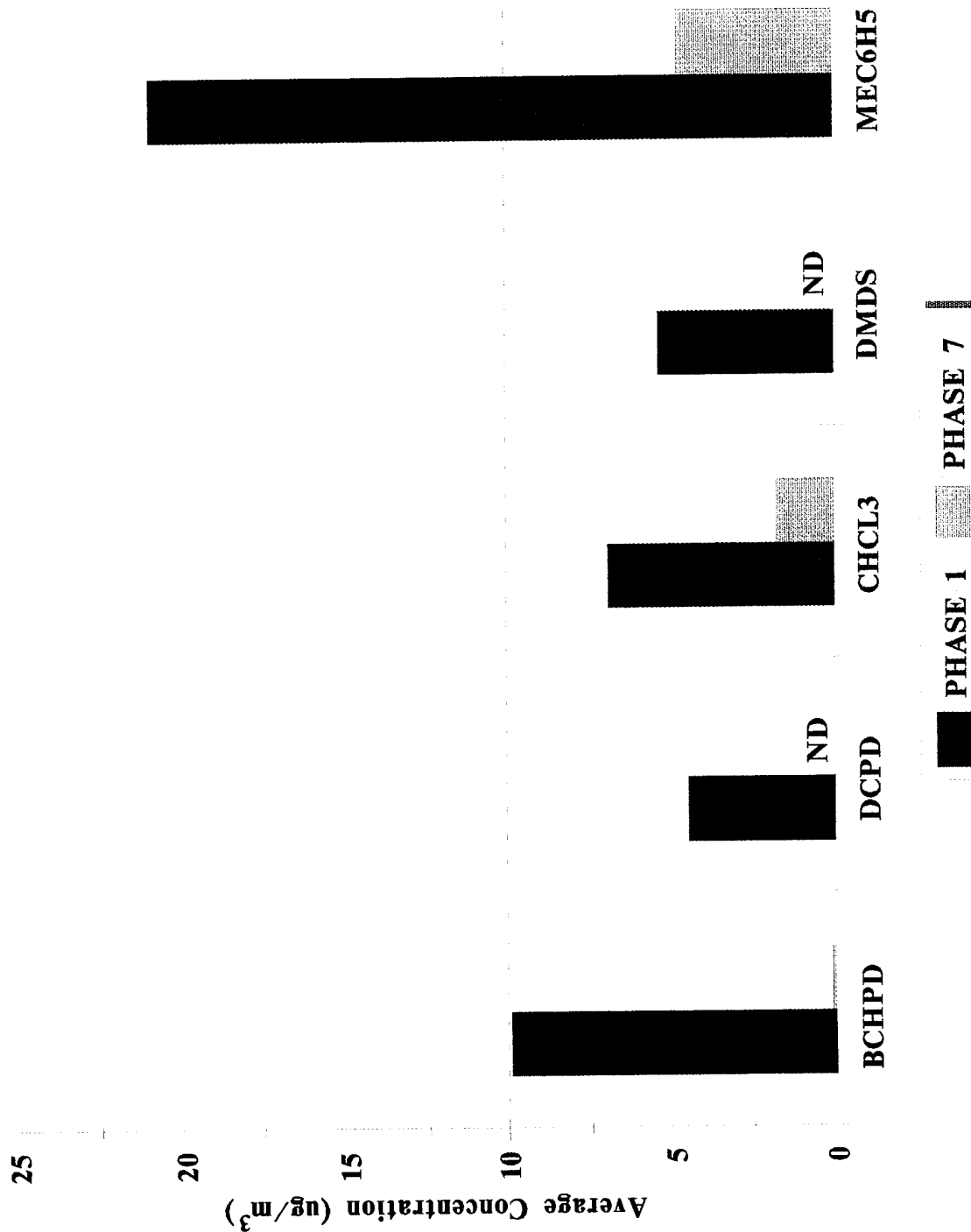
µg/m³ Micrograms per cubic meter

Note: Phase 1 is from March 22 to December 12, 1988.

Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989.

Phase 7 is from October 1, 1992 to September 30, 1993.

| | | | | |
|----------|----------------------|---------|---|------------------------|
| CLC6H5 - | Chlorobenzene | MEC6H5 | - | Toluene |
| DBCP - | Dibromochloropropane | MIBK | - | Methyl isobutyl ketone |
| DCPD - | Dicyclopentadiene | TCLEE | - | Tetrachloroethene |
| DMDS - | Dimethyldisulfide | TRCLE | - | Trichloroethene |
| ETC6H5 - | Ethylbenzene | XYLENES | - | Total Xylenes |



BCHPD - Bicycloheptadiene
 DCPD - Dicyclopentadiene
 CHCL3 - Chloroform
 DMDS - Dimethyldisulfide
 MEC6H5 - Toluene
 ND - Not detected during this phase

Prepared for:
 U.S. Army Program Manager
 for Rocky Mountain Arsenal
 Figure 4.6-1
 FC2 Average VOC Concentration
 for Phases 1 and 7
 Rocky Mountain Arsenal
 Prepared by: Ebasco Services Incorporated

with distance from the Basin F source (e.g., BF2 versus BF5). This suggests that Basin F was the principal source of these contaminants and that impacts were localized and decreased beyond the immediate vicinity of the remediation activities.

Another approach to assessing Basin F impacts is to evaluate monitoring data during worst-case meteorological conditions. During active remediation periods, higher concentrations for many contaminants were downwind from the flow across Basin F, and were also consistent with air dispersion modeling evaluations using Basin F as a potential area source. Several Basin F air monitoring scenarios continue to be employed in the high-event program (to be discussed in the next section) and, as will be noted, significant concentrations of contaminants are no longer discovered downwind from Basin F, though on occasion some residual levels above baseline are measured.

4.6.4 VOC High-Event Monitoring

VOCs were monitored in FY93 routinely, bi-weekly, monthly, seasonally, and also in accordance with high-event, worst-case meteorological conditions. For the most part, the identification of high events for VOCs attempted to meet specified meteorological criteria, e.g., temperatures above 75°F at the start of the monitoring period and winds less than 5 mph. Since the wind criteria of less than 5 mph was extremely difficult to predict during the summer seasons, especially during mid-afternoon when convective activity was at its highest levels, emphasis, as in the past, was placed on mobilizing for light wind days, generally between 5 and 8 mph. Because of the variability of light winds, each high-event monitoring episode concentrated on a particular potential source with several samplers, including mobile monitors, surrounding the source area.

Experience with earlier CMP/CAQMMP programs also formed the basis for modifications and improvements to the high-event program. Although VOC monitoring was initiated on very warm days with temperatures in excess of 75°F, the monitoring was generally continued though the 24-hour period in order to capture the impacts of VOCs released under the warm temperatures and

then trapped under inversion conditions. Table 4.6-6 provides a summary of the FY93 high-event program for VOCs, including dates, selected target areas, sampling duration, and the sampling sites employed.

Also, in the past several years the interpretation of the high-event program has been broadened to include not only worst-case meteorological conditions but also special activities or contingency events that might warrant special monitoring (such as a leaking valve or pipeline). To the extent possible, appropriate meteorological conditions were sought, but in certain instances this was not possible.

FY93 VOC high-event episodes are described in the following sections. Several cases of interest are further illustrated in figures including selected ambient concentrations, wind roses, and dispersion modeling as appropriate. For those cases not illustrated, a detailed listing of all concentrations for the day of the event is provided in Appendix E.

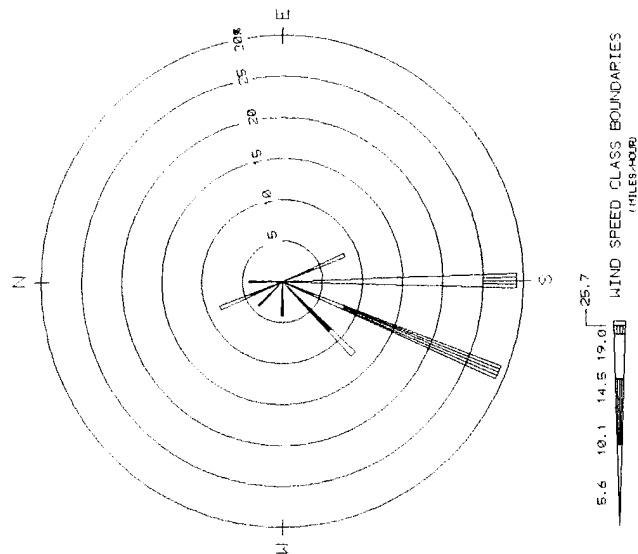
4.6.4.1 December 11, 1993—Seasonal High-Event

This was a late fall VOC seasonal high-event (the Technical Plan requires four such seasonal events). On this day, the winds were moderate to strong and persistently from the south. Temperatures ranged from below freezing to near 50°F at mid day. All the RMA perimeter sites were sampled and Basin F was also targeted with two downwind sampling sites. The highest levels of benzene, methylene chloride, and trichloroethene were measured at AQ1 on the western perimeter; the highest levels of toluene, ethylbenzene, and total xylenes were measured at the northwest perimeter site AQ2; and the highest tetrachloroethene was measured at perimeter site AQ5 at the southern boundary. 1,1,1-trichloroethane was highest at AQ6 in the southern portion of RMA; accordingly, with the wind flow as shown in Figure 4.6-2, it would appear that the source of this contamination is metropolitan Denver. Since this day was three days after the previously described severe pollution/inversion day of December 8, 1992, some residual air quality impacts may still have been evident. Two contaminants were identified as potential Basin F source emissions: 1) bicycloheptadiene was measured downwind from Basin F at FC2

Table 4.6-6 FY93 Volatile Organic Compound High-Event Sampling Locations

Page 1 of 1

| Date | Target Area | Duration | Sampling Locations |
|----------|---|------------------------|---|
| 12/11/92 | Basin F and Seasonal | 1100-1100 | AQ1, AQ2, AQ3, AQ4, AQ5, AQ6, FC1, FC2 |
| 2/24/93 | Seasonal | 1200-1200 | AQ1, AQ4 |
| 3/4/93 | Submerged Quench Incinerator (Baseline) | 1200-1200 | AQ2, AQ3, AQ4, AQ5, AQ6, FC2, SQ1, SQ2, M1E |
| 5/13/93 | Hydrazine Plant | 1100-1100 | AQ2, AQ6, M201W, M301N, M401E, M501S |
| 6/3/93 | Seasonal | 1100-1100 | AQ1, AQ4 |
| 6/12/93 | Submerged Quench Incinerator (Trial Burn) | 0700-0700 | AQ3, AQ6, SQ1, SQ2, FC2, M125W |
| 6/24/93 | Basin F Excavation Activities | 1300-1300 | AQ5, AQ8, AQ9, FC1, M336E, M436W |
| 6/30/93 | Basin F | 1100-1100 | FC1, FC2, FC3, FC4, AQ11, M226E |
| 7/15/93 | Basin A Soil Sifting | 0700-1500 (8 hours) | AQ1, AQ4, AQ5, AQ9, AQ10, M601N, M701N |
| 7/30/93 | Section 27 Water Treatment Plant | 1000-1000 | AQ2, FC1, FC2, M122SW, M127NW, M227NW, AQ1 |
| 8/11/93 | South Plants Area | 1200-1200 | AQ5, AQ6, AQ8, AQ9, M801N, M102E |
| 8/29/93 | Submerged Quench Incinerator | 1300-1300 | AQ3, AQ6, SQ1, SQ2, FC2, M225W |
| 9/12/93 | Basin F Waste Pile | 1200-1200 | FC2, FC3, FC4, M326E, M426S, M526SE |



LEGEND

- Air Quality Monitoring Station
- ▲ Meteorological Monitoring Station
- Continuous Gaseous and Meteorological Monitoring Station
- Submerged Quench Incinerator
- Modeled Area Source

X/Q Contour



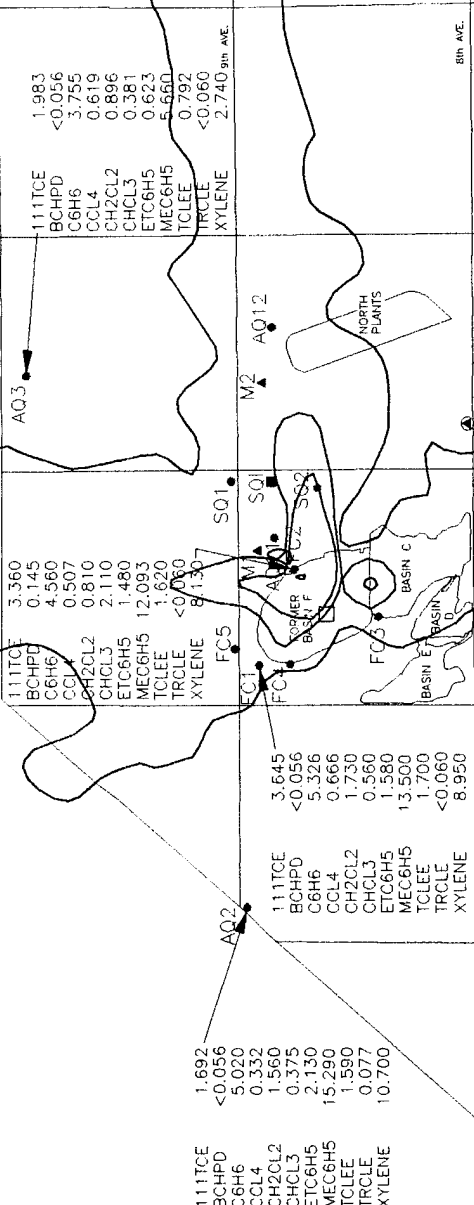
Prepared for:

U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.6-2

VOC Results ($\mu\text{g}/\text{m}^3$) and X/Q
Dispersion Pattern for 12/11/92

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



111TCE - 1,1,1-Trichloroethane
BCHPD - Bicycloheptadiene
C6H6 - Benzene
CCL4 - Carbon Tetrachloride
CH2CL2 - Methylene Chloride
CHCL3 - Chloroform
ETC6H5 - Ethylbenzene
MEC6H5 - Toluene
TCLEE - Tetrachloroethene
TRCLE - Trichloroethene
XYLENE - Total Xylenes

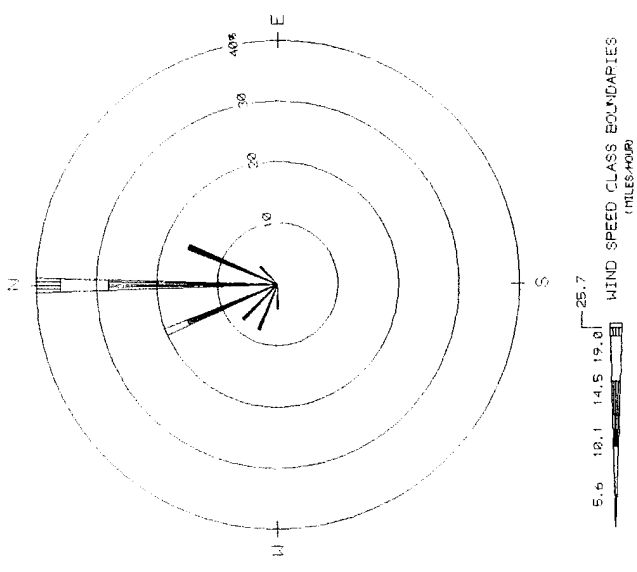
(and FC2D, not shown), and 2) chloroform was measured at highest levels at both FC2 and FC1. Bicycloheptadiene, a contaminant unique to Basin F, was measured at a low level, $0.154 \mu\text{g}/\text{m}^3$, on this less than ideal meteorological day for VOC detection. Carbon tetrachloride was also highest at FC1 but the concentration level was comparable to perimeter sites. Figure 4.6-2 provides a display of all concentrations, the atmospheric dispersion pattern and the 24-hour wind rose for this day. Note that FC1, and particularly FC2, are within the potential impact of Basin F as reflected by the X/Q contours from a theoretical Basin F area source.

4.6.4.2 February 24, 1993—Basin F (also Seasonal High-Event)

On this day, VOCs were measured at four Basin F sites, the SQI sites, and RMA perimeter sites. Winds were primarily from the north (see Figure 4.6-3) and temperatures were in the low 30s. Winds were light (2 to 3 mph) through much of the sampling period. Higher levels of 1,1,1-trichloroethane, carbon tetrachloride, methylene chloride, and chloroform were all measured at Basin F Site FC5; other maximum values were primarily at boundary sites. It is difficult to explain the higher concentrations at FC5 as winds were exclusively from the north, upwind from the Basin F area. An explanation may be its close proximity to Pond A, or possibly the recirculation of contaminants in the Denver metropolitan basin under persistent light winds and drainage conditions on this day.

4.6.4.3 March 4, 1993—Submerged Quench Incinerator

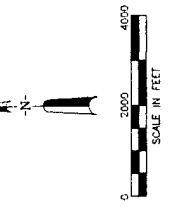
On this day, VOC high-event monitoring was conducted to investigate the SQI area baseline prior to operation of the incinerator. Monitoring was conducted north and south of the SQI at SQ1 and SQ2, respectively, west of the SQI at FC2, and east of the SQI at a mobile site, M1E. Temperatures were in the 40s, and winds were from the northeast at moderate to light speeds. Most VOC concentrations were either below detection, below average, or near baseline levels, except carbon tetrachloride, which was measured at $0.926 \mu\text{g}/\text{m}^3$, slightly above the annual average value, and 1,2-dichloroethane which was measured at a low concentration level, $0.078 \mu\text{g}/\text{m}^3$; both were measured at FC2. There appeared to be no significant baseline impacts in the SQI area.



LEGEND

- Air Quality Monitoring Station
- ▲ Meteorological Monitoring Station
- Continuous Gaseous and Meteorological Monitoring Station
- Submerged Quench Incinerator
- Modeled Area Source

X/O Contour

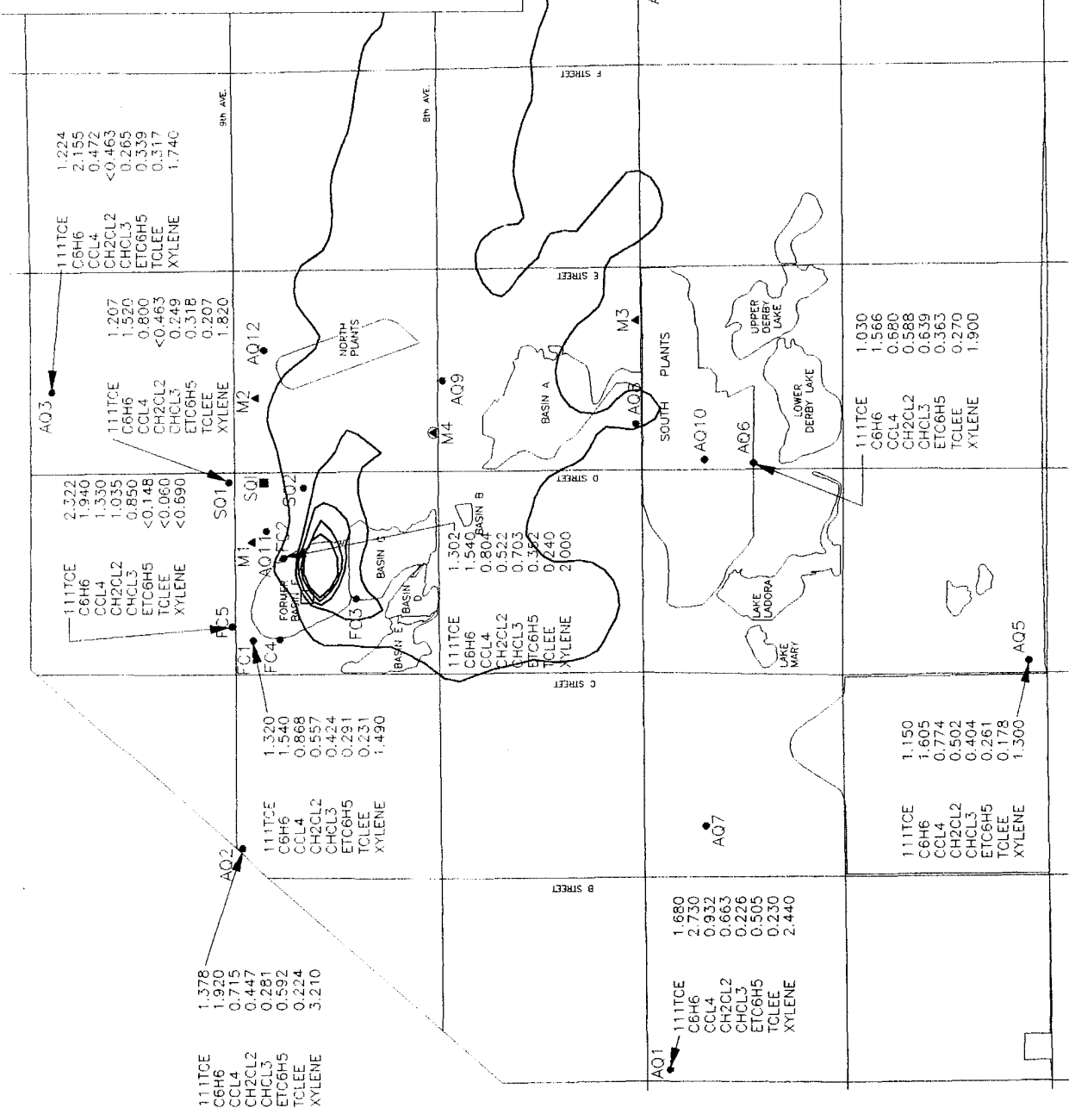


Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.6-3

VOC Results ($\mu\text{g}/\text{m}^3$) and X/Q
Dispersion Pattern for 2/24/93

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



111TCE - 1,1,1-Trichloroethane
BCHPD - Bicycloheptadiene
C6H6 - Benzene
CCL4 - Carbon Tetrachloride
CH2CL2 - Methylene Chloride
CHCL3 - Chloroform
ETC6H5 - Ethylbenzene
MEC6H5 - Toluene
TCLEE - Tetrachloroethene
TRCLE - Trichloroethene
XYLENE - Total Xylenes

4.6.4.4 May 13, 1993—Hydrazine Plant Area

On this day, high-event monitoring was conducted in the vicinity of the remediated hydrazine plant area where VOC contaminants had previously been identified. Extensive monitoring was conducted as this high event also coincided with a routine sampling day. Four mobile sites surrounded the old hydrazine area (M301N, M501S, M4013, and M201W), and the program included SQ1, SQ2, AQ2, AQ3, AQ5 and AQ6. Winds were primarily from the southeast on this day and were moderate to light. Temperatures reached the low 70s (see Figure 4.6-4).

Several results of interest were noted on this sampling day. The old hydrazine plant area, which was the principal area of investigation, indicated little or no residual VOC impacts, however, slightly higher concentrations for several VOCs occurred at FC1 downwind from the remediated Basin F area. These included ethylbenzene, benzene, carbon tetrachloride, chloroform, and xylene. This was a more favorable high-event sampling day than the previous fall and winter sampling events and suggests that Basin F remains a low-level source for several VOCs. Note that the VOC potential source in the figure is centered over the old hydrazine plant area. If the source were centered over Basin F, the higher X/Q contours shown in Figure 4.6-4 would coincide with the location of Station FC1.

Of particular interest on this day were the moderate-to-high levels of methylene chloride which had begun to be measured at RMA sites starting in April 1993 and continuing to mid-June 1993. This compound was measured at all RMA sites with the lowest levels at the hydrazine sites and the highest levels at perimeter sites AQ5 and AQ2, which measured $52.6 \mu\text{g}/\text{m}^3$ and $52.0 \mu\text{g}/\text{m}^3$ respectively. High methylene chloride concentration levels will be discussed further in Section 4.6.5.2.

4.6.4.5 June 3, 1993—Seasonal High-Event

This was a late spring seasonal high-event that included perimeter sites AQ1, AQ3, AQ4, and the SQI special network. Temperatures were in the 60s and winds were moderate and

occasionally gusty. This was not an ideal day for high-event VOC monitoring and most analytes were reported at low or non-detectable levels. Of special interest on this day, however, was that very high methylene chloride concentrations were measured at all sites, including the highest level of the year, 560.0 $\mu\text{g}/\text{m}^3$ at AQ1, and 440 $\mu\text{g}/\text{m}^3$ at AQ6. This case study will be discussed further in Section 4.6.5.2 (and illustrated in Figure 4.6-9).

4.6.4.6 June 12, 1993—SQI Stack Testing

This was a follow-on high-event sampling episode at the SQI during a period when the facility was fully operational and undergoing special stack testing. Temperatures reached the low 80s during the monitoring period and winds were primarily from the south at low-to-moderate speeds. Monitoring was conducted at the SQI special network (SQ1, SQ2, AQ3, and AQ6), as well as at FC2 and FC5. Highest concentrations occurred at slightly above-average levels at the FC2 site. With respect to the SQI, levels at all sites were close to uniform. SQ2 concentrations to the south were slightly higher than SQ1 levels to the north. With winds primarily from the south, this would not suggest that the SQI facility was a source of VOCs. Methylene chloride was again measured at moderate levels with the highest levels, 29 $\mu\text{g}/\text{m}^3$, occurring at SQ2 and also at AQ3 on the northern perimeter. Complete results of the SQI special monitoring program are discussed in Section 4.6.10.

4.6.4.7 June 24, 1993—Basin A Excavation Activities

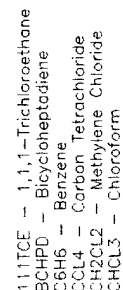
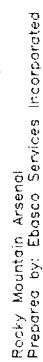
High-event monitoring was initiated on this day because of excavation activities in the Basin A area. Winds were moderate with occasional gusts and temperatures reached the upper 70s. Wind directions were from the north, east, and south with potential dispersion impacts to the northwest through southwest. Basin A was surrounded using AQ9 to the north, AQ8 to the south, and mobile sites to the east and west. Four FC sites were also sampled, as were perimeter sites AQ2 and AQ5. Most of the compounds appeared to be uniform upwind and downwind from the Basin A area. The one exception was chloroform which measured 4.0 $\mu\text{g}/\text{m}^3$ at AQ8. Chloroform has been identified in the Basin A and South Plants areas in previous sampling.

4.6.4.8 June 30, 1993—Basin F

This was a very broad high-event monitoring day that encompassed 15 sampling sites, including five Basin F sites; a mobile site, M226E, directly east of Basin F; the SQI interior sites; and three perimeter sites. Temperatures reached a maximum of 92°F and winds were moderate to light, primarily from the northeast. The dispersion pattern from a potential Basin F source spread across the remediated basin floor from northeast to southwest (see Figure 4.6-5). On this day many of the highest VOC concentrations were measured at Basin F sites, especially at FC3 downwind. As the wind was exclusively from the northeast quadrant on this day, and away from downtown Denver, the higher VOC levels could be attributed to residual impacts from Basin F. Nevertheless, as stated elsewhere, concentration levels were one or two orders of magnitude below peak remediation values. It is interesting that methylene chloride levels were all below detection except for one very low measurement of 0.710 $\mu\text{g}/\text{m}^3$ at FC3.

4.6.4.9 July 15, 1993—Basin A Soil Sifting Building 545, South Plants

This was an 8-hour VOC high-event episode during soil sifting activities in Building 545 in South Plants. Mobile monitoring sites were located close to the north, east, and west of the building; AQ10 to the south and AQ8 to the north were also used. Winds were exclusively from the northern quadrant on this day and were generally light; it is likely that there were variable wind circulations within the South Plant building complex. Temperatures reached the mid-80s. The case study is illustrated in Figure 4.6-6, with a potential VOC source in the area of Building 545. Although many VOC compounds were below detection, those that were measured, including toluene, tetrachloroethene, benzene, carbon tetrachloride, and chloroform, were all highest in the network surrounding Building 545. In particular, chloroform reached a peak level of 9.70 $\mu\text{g}/\text{m}^3$ at mobile site M601ND, 100 feet east of the building. A high chloroform level of 7.1 $\mu\text{g}/\text{m}^3$ was also measured at AQ8 just to the north of South Plants. With the winds exclusively from the northern quadrant, it would appear that the source of chloroform was the Basin A or Basin F areas. It is possible that there may have been close-in impacts from the soil sifting activities, but comparable levels of these compounds have previously been measured within the South Plants complex.



4.6.4.10 July 30, 1993—Section 27 Water Treatment Plant

At the request of RMA, high-event monitoring was conducted around the groundwater treatment plant in Section 27, at the northwest corner of RMA. Several mobile sampling stations, in addition to AQ2, were in close proximity to the facility. Winds were variable and generally light; temperatures reached the mid 90s. The highest concentrations for most compounds were measured at AQ5 at the south perimeter or at Basin F. Concentrations at the sites surrounding the water treatment site were below average with exception of toluene, which measured $5.20 \mu\text{g}/\text{m}^3$ at mobile site M12714W, and also at nearby site AQ2. However, on this day, site AQ5 on the south perimeter measured a toluene level of $7.14 \mu\text{g}/\text{m}^3$, indicating a broad source over the metropolitan area. Based on these data, the water treatment plant does not appear to be impacting the air quality in its vicinity.

4.6.4.11 August 11, 1993—South Plants

On this date, high-event monitoring was conducted at the South Plants. Winds were light to moderate and were bimodal in direction, from the southeast and northwest. Temperatures reached the low 80s. Monitoring was conducted at three mobile sites within the South Plants and at 13 other sampling sites across RMA. The highest VOC levels on this day were measured at perimeter sites AQ2 and AQ5 and at Basin F sites. South Plants mobile sites were comparable to levels measured at South Plants during the July 15 soil sifting activities (see Section 4.6.4.9 above) except for chloroform, which was very low on August 11 at the South Plants sampling sites compared to a peak level of $9.7 \mu\text{g}/\text{m}^3$ on July 15. On the other hand, chloroform was again high at AQ8 reaching a maximum concentration of $10.00 \mu\text{g}/\text{m}^3$; Basin F or Basin A could have been a source of chloroform emissions on this day.

4.6.4.12 August 29, 1993—SQI facility

On this day, high-event monitoring was conducted in the vicinity of the SQI and also at Basin F and perimeter sites. Winds were from the south-southeast through the northwest and were light to moderate. Temperatures reached the low 80s. The dispersion pattern indicated a potential

spread of pollutants from south to north. With respect to potential impacts from the SQI, VOC levels were low and there appeared to be little difference between concentrations at SQ1 north of the facility and SQ2 south of the facility. Several VOCs including benzene, ethylbenzene, toluene, tetrachloroethene, total xylenes, and bicycloheptadiene, measured values well above averages of this day in the vicinity of Basin F. This is illustrated in Figure 4.6-7. Of special interest, collocated passivated canisters were also tested on this day. Results were comparable with several individual exceptions. The canister tests will be discussed in Section 4.6.9.

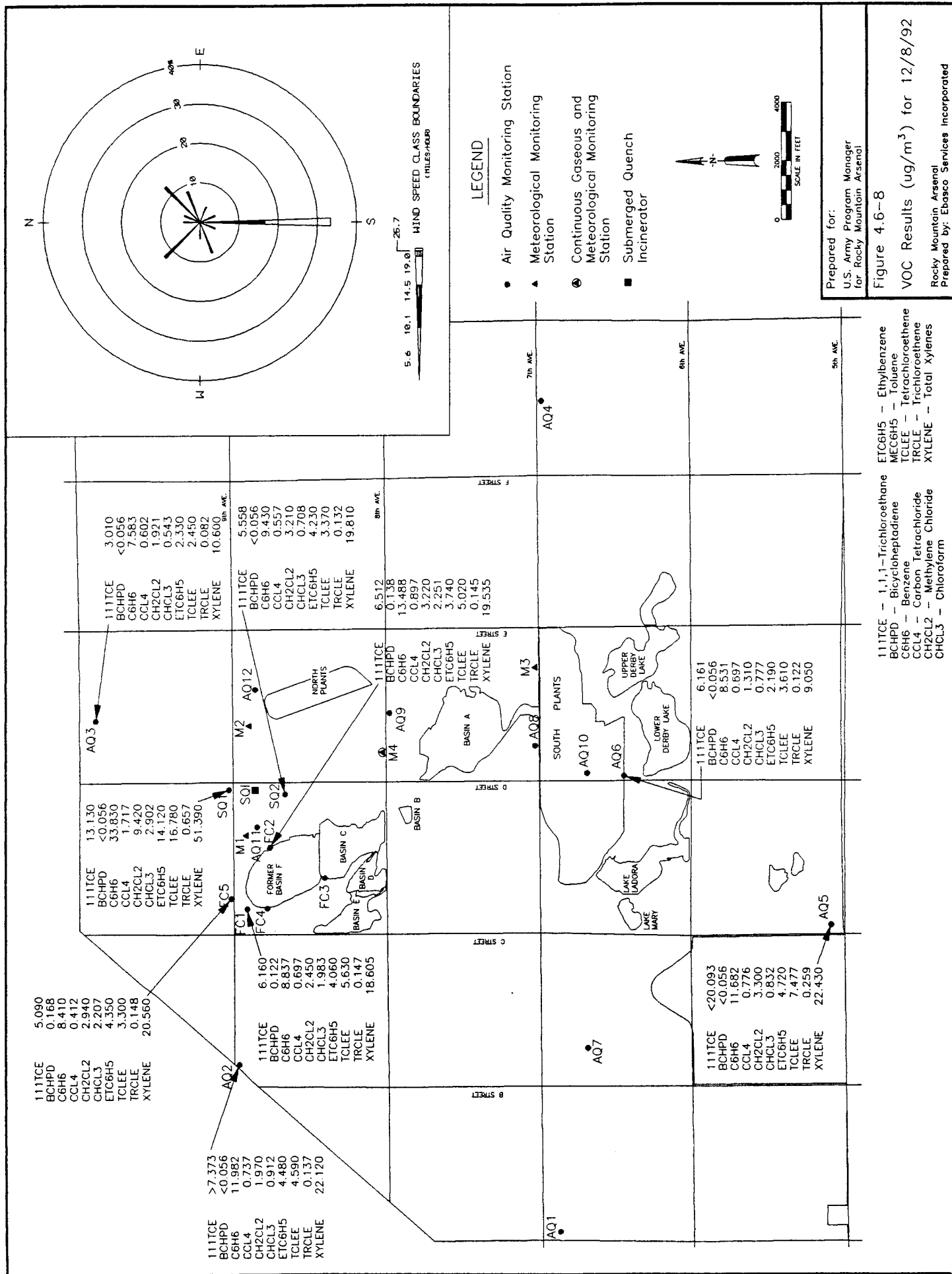
4.6.4.13 September 12, 1993—Basin F

VOC high-event monitoring was conducted at Basin F on this date. Wind speeds were moderate to strong (this was conducted in conjunction with SVOC high-event monitoring where strong winds were a criteria sought), and temperatures reached the high 80s. Interestingly, all concentrations were considerably below the levels measured in the high-event episode 2 weeks earlier (Section 4.6.4.12). One explanation is that the winds were much stronger (reaching 30 mph) and not conducive to the trapping of VOCs, but this event also demonstrated the variability of VOCs at a given location, especially under variable meteorological conditions.

4.6.5 Additional Special VOC Monitoring Results

4.6.5.1 December 8, 1992

The above examples demonstrate the variability of VOC measurements across RMA, as well as the likelihood that some residual emissions exist in special areas such as Basin F, South Plants, and Basin A. It should be stressed that while higher concentrations of some VOCs are unique to areas such as Basin A and Basin F, they are well below levels that were achieved during active remediation. This can be put in a different perspective by examining potential impacts from metropolitan Denver, and particularly the consequences of an intense inversion day such as December 8, 1992 (which has been discussed previously). On this date, a very strong ground inversion existed over the Denver area for an extended period of time. Winds were light (under 5 mph) and primarily from the south. Temperatures remained below freezing the entire day. Figure 4.6-8 shows the VOC concentrations measured at RMA and the wind rose and dispersion



pattern; it can be assumed that the dispersion source was a broad area centered over downtown Denver and was advected to the north and over RMA. As discussed previously, the highest annual levels of TSP, PM-10, and criteria pollutants were measured in Denver and/or at RMA on this date. With respect to VOCs, seven compounds (1,1,1-trichloroethene, trichloroethene, tetrachloroethene, benzene, toluene, ethylbenzene, and total xylenes) measured their highest 24-hour maximum levels of the year on this date. Almost all individual stations measured peak values. Many concentrations were close to remediation peak levels and two, ethylbenzene and xylene, exceeded these levels.

In addition to the December 8, 1992 episode, several other analytes during FY93 measured annual peak values at RMA perimeter sites including chloroform, carbon tetrachloride, chlorobenzene, methylene chloride and 1,2-dichloroethane.

CDH and EPA do not routinely monitor VOCs in Denver, although several studies have been conducted, and an increasing regulatory effort has been made to identify metropolitan Denver VOC emission sources and toxicity levels. This information has appreciably assisted in interpreting and evaluating data collected under the CAQMMP. An interesting observation of the December 8, 1992 inversion episode is that the RMA monitoring network has, in essence, identified significant potential VOCs within the metropolitan air shed such as those analytes listed above (at least within the limitations of the CAQMMP VOC target list). Several target compounds historically identified at RMA were below detection during FY93 and one may conclude that they were not a factor in Denver or RMA during this year. Several other analytes discussed above (e.g., bicycloheptadiene and chloroform) undoubtedly represent unique local sources within RMA, but there may also be emission sources within the metropolitan area. One of the more challenging tasks of the CAQMMP is to identify the source of various pollutant constituents (including VOCs, metals, pesticides, and particulates) and to distinguish between on- and off-post sources. For example, methylene chloride is a compound that was observed (measured) at significant levels starting at about the middle of the FY93 program. The source

of these methylene chloride emissions is currently being investigated by the CAQMMP; results of initial investigations are discussed below.

4.6.5.2 Methylene Chloride Results and Analysis

During a period of several months in FY93, methylene chloride concentrations were measured at levels well above historical remediation and other RMA baseline average and maximum values. High concentrations were measured intermittently at almost all of the CAQMMP VOC sampling stations. Many 24-hour concentrations were in excess of $100 \mu\text{g}/\text{m}^3$ and a maximum concentration of $556 \mu\text{g}/\text{m}^3$ was measured at AQ1 on June 3, 1993. The previous maximum 24-hour level of methylene chloride was $434 \mu\text{g}/\text{m}^3$ measured at BF3 during the Basin F remediation program (there was one other high concentration of $179 \mu\text{g}/\text{m}^3$ at BF6 during the remediation program, and all other methylene chloride maximum levels were well below $100 \mu\text{g}/\text{m}^3$).

Table 4.6-7 shows all methylene chloride concentrations measured at RMA between April 1, 1993 and July 12, 1993; during this time the third-day special sampling operations were conducted at Stations AQ3, AQ6, SQ1, and SQ2 to characterize the ambient air quality during startup of the SQI operations (see Section 4.6.10). Figure 4.6-9 shows the distribution of average and maximum 24-hour concentrations across RMA during this monitoring period (note, as shown in Table 4.6-7, the number of sampling days are different for individual stations). Finally, Figure 4.6-10 shows a bar graph of the maximum 24-hour concentrations for the same periods. The following interesting observations can be discerned from these data:

- High-to-moderate methylene chloride levels (in excess of $5 \mu\text{g}/\text{m}^3$) occurred intermittently and were generally reflected at all the RMA sampling locations on the same day.
- Peak concentrations (in excess of $100 \mu\text{g}/\text{m}^3$) occurred at almost all VOC sampling stations including perimeter sites AQ1, AQ2, AQ3, and AQ5; one time monitoring at AQ4 measured a methylene chloride concentration of $97 \mu\text{g}/\text{m}^3$.

Table 4.6-7 Methylene Chloride Results (µg/m³)

| Date | AQ1 | AQ2 | AQ3 | AQ4 | AQ5 | AQ6 | FC1 | FC2 | FC3 | FC4 | FC5 | SQ1 | SQ2 | FBLK (µg) | Trip (µg) | FC1 Duplicate |
|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|-----------|-----------|---------------|
| 4/1/93 | | < 0.463 | < 0.463 | | < 0.463 | < 0.463 | < 0.463 | < 0.463 | | | < 0.463 | < 0.463 | < 0.463 | < 0.100 | < 0.100 | < 0.463 |
| 4/7/93 | | 0.523 | 1.060 | | 3.386 | 2.586 | 0.614 | 2.090 | 2.240 | 3.460 | 2.320 | 0.972 | 1.700 | 0.358 | < 0.100 | 1.310 |
| 4/13/93 | | 4.103 | 2.582 | | 6.908 | 6.217 | 2.552 | 2.518 | | | 1.710 | 13.574 | 1.047 | 0.198 | < 0.100 | 1.991 |
| 4/16/93 | | | 73.783 | | | 2.950 | | | | | | 61.822 | 56.468 | 0.284 | < 0.100 | |
| 4/19/93 | | 3.900 | 1.500 | | 2.500 | 4.100 | 1.300 | 3.790 | | | < 0.463 | < 0.463 | 5.505 | 0.170 | < 0.100 | 0.510 |
| 4/22/93 | | | 1.600 | | | 0.660 | | | | | | < 0.463 | < 0.463 | 0.120 | 0.220 | |
| 4/25/93 | | < 0.463 | < 0.463 | | 0.490 | < 0.463 | < 0.463 | < 0.463 | | | < 0.463 | < 0.463 | < 0.463 | < 0.100 | < 0.100 | < 0.463 |
| 4/28/93 | | > 4.630 | > 4.630 | | | < 0.463 | | | | | | < 0.463 | < 0.463 | < 0.100 | < 0.100 | |
| 5/1/93 | | 20.957 | 4.800 | | 2.100 | 9.123 | 1.200 | 32.301 | | | 3.700 | 5.000 | 7.900 | 0.140 | < 0.100 | 14.140 |
| 5/4/93 | | | 30.698 | | | 13.269 | | | | | | 28.636 | 6.000 | < 0.100 | < 0.100 | |
| 5/7/93 | | 100.437 | 11.538 | | 21.127 | 78.431 | 44.390 | 65.909 | 11.556 | 10.455 | 27.014 | 10.423 | 13.980 | < 0.100 | < 0.100 | 12.084 |
| 5/10/93 | | | 25.860 | | | 8.700 | | | | | | 16.071 | 30.000 | < 0.100 | < 0.100 | |
| 5/13/93 | | 52.000 | 27.570 | | 52.600 | 9.589 | 14.000 | 11.000 | | | 5.200 | 24.775 | 8.411 | < 0.100 | < 0.100 | 22.000 |
| 5/16/93 | | | 16.200 | | | 7.000 | | | | | | 16.100 | 10.500 | < 0.100 | < 0.100 | |
| 5/19/93 | | 390.295 | 332.743 | | 295.455 | 368.182 | 293.023 | 184.700 | | | 273.300 | 272.207 | 254.579 | 2.710 | 0.630 | 134.200 |
| 5/22/93 | | | 30.000 | | | 111.000 | | | | | | 89.000 | 90.000 | 0.600 | 0.480 | |
| 5/25/93 | | | 8.790 | | | 21.740 | 2.400 | 14.100 | | | 14.600 | 8.019 | 4.408 | < 0.100 | < 0.100 | 1.100 |
| 5/28/93 | | | 3.660 | | | 1.800 | | | | | | 2.070 | 3.800 | 0.400 | < 0.100 | |
| 5/31/93 | | 0.710 | < 0.463 | | < 0.463 | < 0.463 | < 0.463 | 1.300 | | | 1.000 | 0.850 | 0.660 | < 0.100 | < 0.100 | 1.100 |
| 6/3/93 | 560.000 | | 51.000 | 97.000 | | 440.870 | | | | | | 210.000 | 110.580 | < 0.100 | < 0.100 | |
| 6/6/93 | | 70.000 | | | 14.540 | 31.000 | 20.200 | 28.000 | | | 37.000 | 16.900 | 30.000 | 0.170 | 0.430 | 34.000 |
| 6/9/93 | | | 395.000 | | | 31.640 | | | | | | 305.000 | 191.000 | 0.350 | < 0.100 | |
| 6/10/93 | | | 50.000 | | | 60.000 | | | | | | 42.000 | 45.000 | 0.200 | < 0.100 | |
| 6/11/93 | | 40.560 | 29.200 | | 2.300 | 4.700 | 14.100 | 4.100 | | | 10.000 | 27.500 | 29.100 | < 0.100 | < 0.100 | |
| 6/12/93 | | | 4.740 | | | 4.300 | | | | | | 6.500 | 4.460 | < 0.100 | < 0.100 | |
| 6/15/93 | | | 2.150 | | | 3.550 | | | | | | 2.600 | 2.650 | < 0.100 | < 0.100 | |
| 6/18/93 | | < 0.463 | < 0.463 | | < 0.463 | 0.470 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | 0.450 | < 0.100 | < 0.100 | < 0.463 |
| 6/21/93 | | | 0.700 | | | 1.100 | | | | | | < 0.463 | 0.670 | 0.340 | 0.210 | |
| 6/24/93 | | < 0.463 | < 0.463 | | < 0.463 | 0.900 | < 0.463 | 2.570 | | | 5.300 | < 0.463 | < 0.463 | 0.016 | < 0.100 | 1.500 |
| 6/30/93 | | < 0.463 | < 0.463 | | < 0.463 | < 0.463 | < 0.463 | < 0.463 | 0.710 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.100 | 0.130 | < 0.463 |
| 7/6/93 | | < 0.463 | < 0.463 | | < 0.463 | < 0.463 | < 0.463 | 0.900 | | | 0.540 | < 0.463 | 1.100 | 0.280 | < 0.100 | < 0.463 |
| 7/9/93 | | | < 0.463 | | | < 0.463 | < 0.463 | | | | | < 0.463 | < 0.463 | < 0.100 | < 0.100 | < 0.463 |
| 7/12/93 | | 1.790 | 0.580 | | < 0.463 | < 0.463 | 0.450 | < 0.463 | | | < 0.463 | < 0.463 | 1.700 | 0.200 | < 0.100 | < 0.463 |
| 7/15/93 | 3.900 | | < 0.463 | < 0.463 | < 0.463 | < 0.463 | | | | | < 0.463 | < 0.463 | < 0.463 | 0.130 | < 0.100 | |
| 7/18/93 | | < 0.463 | < 0.463 | | 1.200 | 0.770 | | < 0.463 | | | < 0.463 | 0.560 | < 0.463 | < 0.100 | < 0.100 | < 0.463 |
| 7/21/93 | | | 0.770 | | | 0.690 | < 0.463 | | | | | 1.000 | 0.660 | 0.230 | < 0.100 | |

µg/m³ - Micrograms per cubic meter

<## - indicates the concentration is less than lower certified reporting limit

FBLK - Field Blank

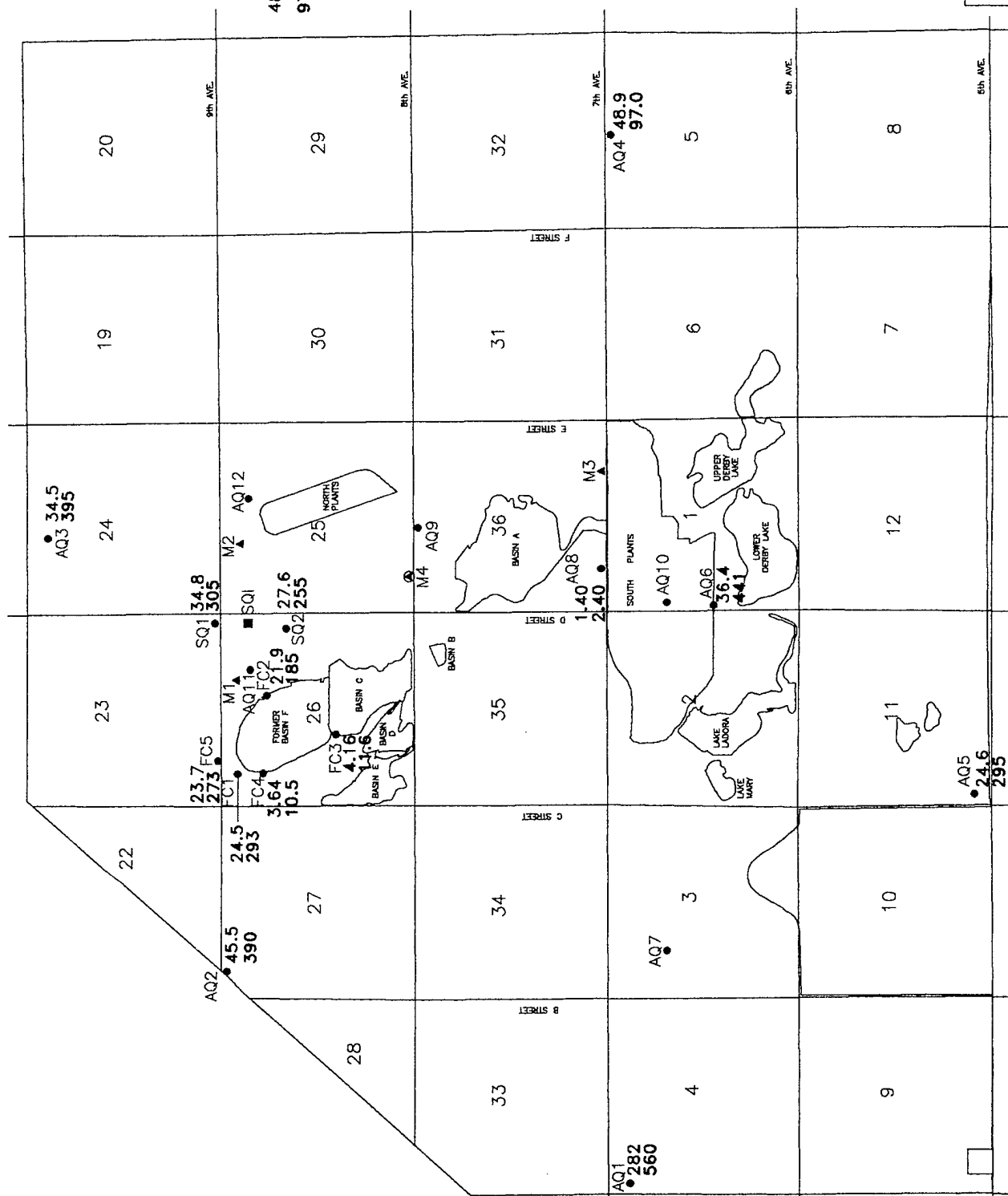
RMA/1043 10/06/94 9:20 am bpw

LEGEND

- Air Quality Monitoring Station
- ▲ Meteorological Monitoring Station
- ⊙ Continuous Gaseous and Meteorological Monitoring Station
- Submerged Quench Incinerator

48.9 24 Hr. Average Concentration

97.0 24 Hr. Max. Concentration



Prepared for:

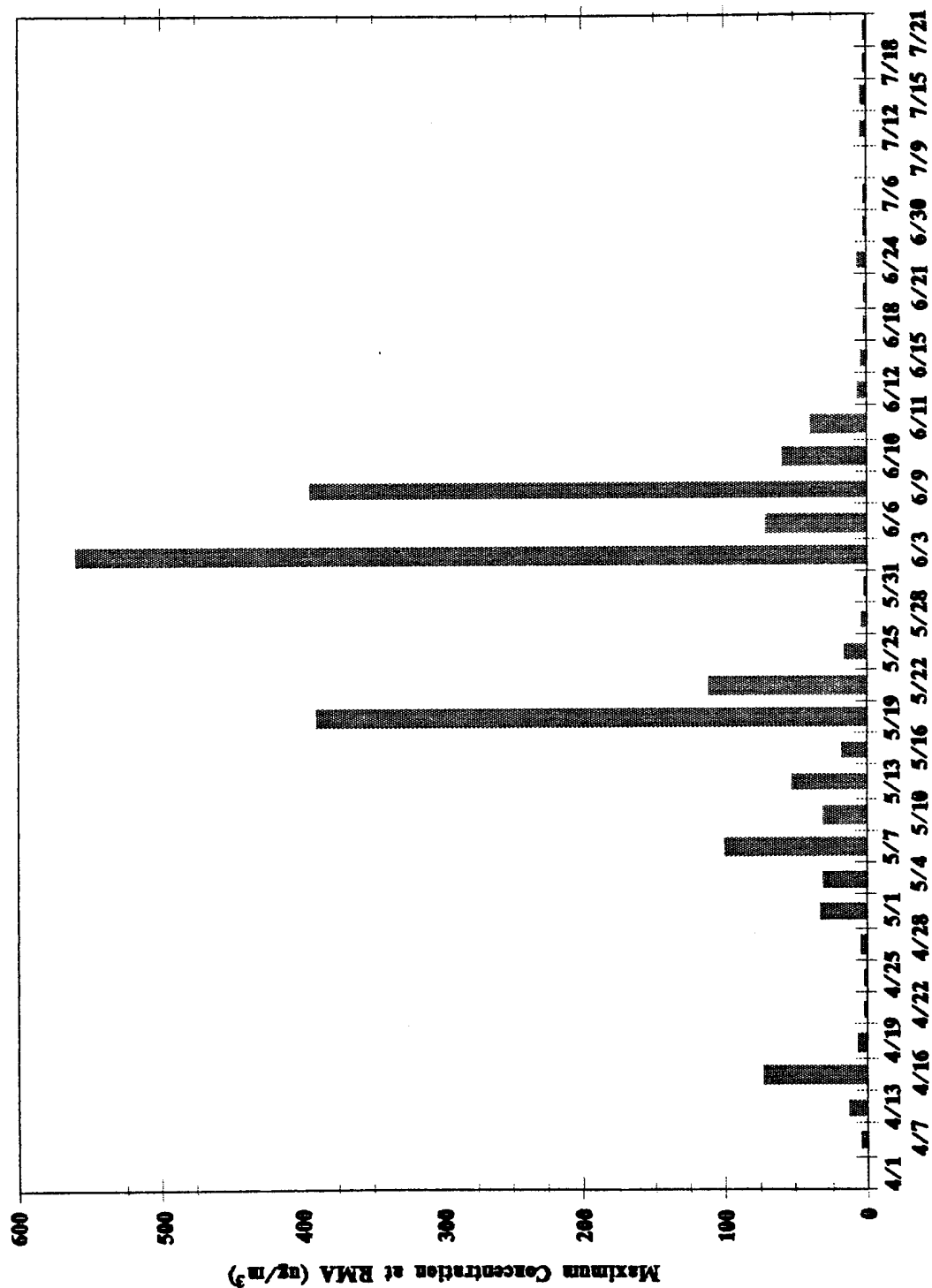
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.6~9

Methylene Chloride Concentration Results
in $\mu\text{g}/\text{m}^3$ for the Period

4/16/93 - 7/21/93

Rocky Mountain Arsenal
Prepared by: **Ebasco Services Incorporated**



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.6-10

Maximum Methylene Chloride
Concentrations at RMA

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

- Peak concentrations on any given day did not correlate with the dispersion pattern from Basin F, the A-Neck Air Stripper, the SQI or any other potential RMA sources. Several very high concentrations at AQ1, AQ2, AQ4, and AQ5 were outside the wind and dispersion patterns from RMA sources (this will be discussed further).
- High methylene chloride concentrations ceased abruptly after July 12 and did not reach a concentration in excess of $5 \mu\text{g}/\text{m}^3$ through the remainder of FY93 (higher levels have been experienced in the spring of FY92, and again in the first half of FY94, before this report was completed).
- Methylene chloride was the only parameter that was measured at anomalous levels when compared to the long-term baseline.

4.6.5.2.1 Methylene Chloride Dispersion Assessments

Assuming that the high methylene chloride concentrations resulted from a potential RMA source, a number of modeling scenarios were investigated including Basin F and SQI ground sources and a possible SQI stack source. In no case, however, could a relationship be shown between sampled methylene chloride emissions and a potential RMA source. Two examples are demonstrated below:

1. On June 3, 1993, winds were primarily from the east-southeast, while other directions were infrequently reflected. Figure 4.6-11 shows the wind rose and the dispersion pattern (for the 24-hour period of monitoring) from the SQI stack as a potential source. The principal dispersion impacts are shown to be to the west-northwest of the stack with some lesser impacts to the south. Methylene chloride measurements are shown in Figure 4.6-11 and it is noted that the highest methylene chloride concentration reported was at AQ1 completely outside the dispersion pattern. The second highest concentration ($441 \mu\text{g}/\text{m}^3$) was at AQ6, which showed a minimal dispersion potential from the SQI stack. AQ4, also removed from the dispersion impact influence, measured a concentration of $97 \mu\text{g}/\text{m}^3$. The high concentrations at stations AQ1, AQ5, and AQ4 cannot be explained by the wind and dispersion patterns on this date as coming from any RMA source.

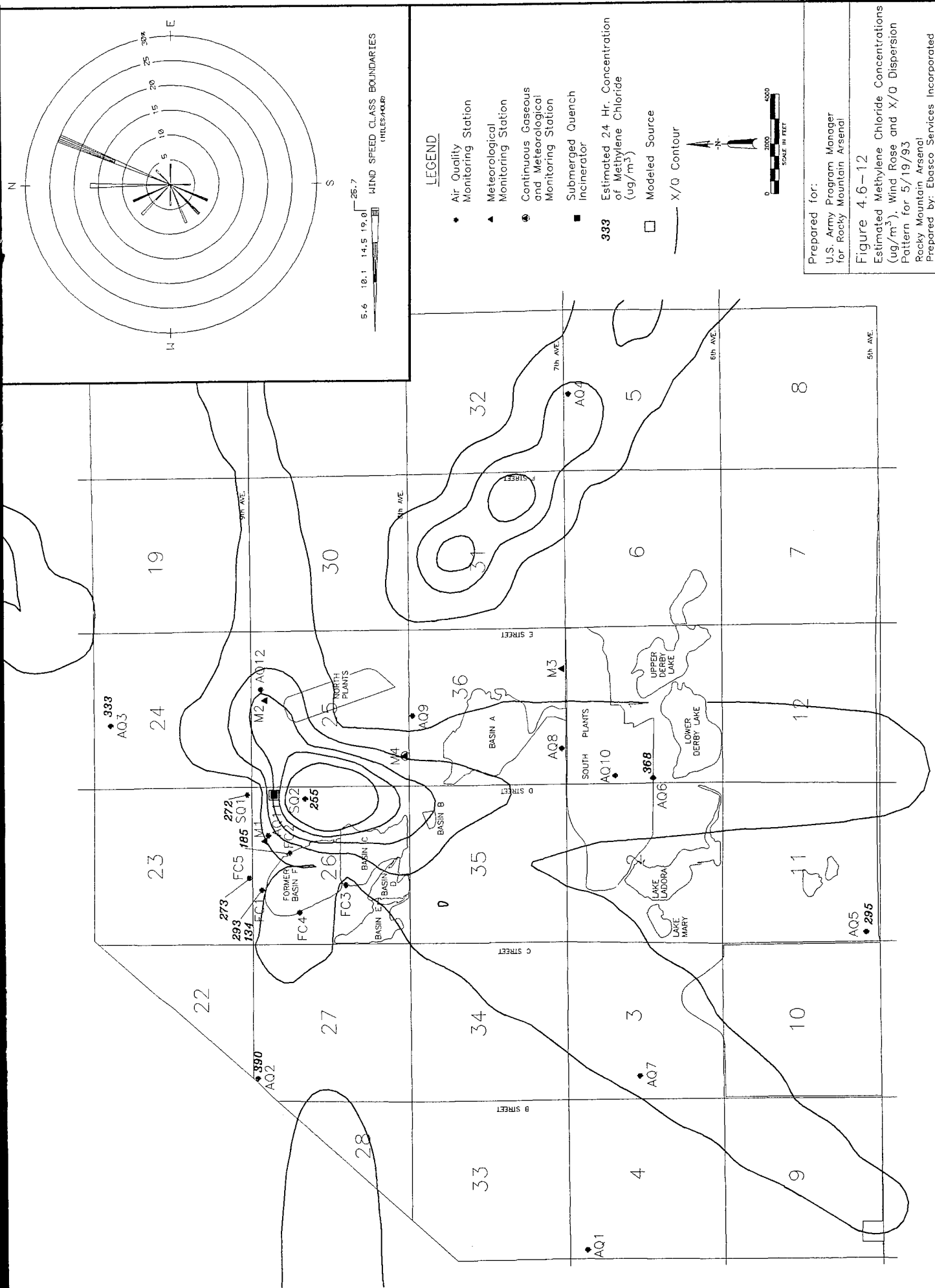
2. On May 19, 1993, a similar example can be demonstrated. On this date winds were primarily from the north and north-northeast, with other directions again minimally represented. Figure 4.6-12 shows the winds and dispersion pattern for the period of sampling. The highest concentrations were reported at AQ2 ($390 \mu\text{g}/\text{m}^3$) and AQ6 ($368 \mu\text{g}/\text{m}^3$), which were both outside the principal dispersion potential. Stations SQ1, SQ2, AQ3, FC1, FC2, and FC5 all measured moderately high and somewhat uniform methylene chloride concentrations but with no correlation to the dispersion patterns.

Similar disparities between methylene chloride high concentrations, RMA potential emissions, dispersion patterns, and meteorological conditions were observed for the other methylene chloride anomalous concentration days shown in Table 4.6-7.

4.6.5.2.2 Potential Methylene Chloride Sources

The source of methylene chloride as of the date of this report has not been detected. The CAQMMP has continued investigations and analyses to determine if the source is at RMA, but real-time, close-in monitoring of the Basin F waste pile, Pond A, and the storage tanks have indicated no increased levels of methylene chloride emissions. It is noted that there have been no anomalous methylene chloride concentrations measured at RMA in the 5-year post-remedial period, with the exception of the spring of FY92. Other sources of methylene chloride at RMA that could produce ambient concentrations at levels measured during the 3-month sampling period have been surveyed and none identified. The SQI has been ruled out as a possible source of methylene chloride for several reasons:

- High methylene chloride concentrations were noted 12 days prior to the SQI startup operations.
- SQI stack testing during trial burns between May 7 and May 25, 1993 indicated a methylene chloride emission rate of $2.20\text{E-}05 \text{ g/sec}$.
- Dispersion modeling indicated that the maximum potential concentrations from the stack using the above emission rate was less than the observed concentrations at downwind receptors by a factor of 10^{-6} .



- The Basin F fluid does not contain methylene chloride at concentrations that, if leaked from piping, could account for observed concentrations at downwind receptors.
- As the Basin F fluid would contain a mixture of other Basin F compounds, these compounds would also be detected in the ambient sample analysis (however, only anomalous levels of methylene chloride were detected).
- Any leaks would result in the dispersion of a low level or ground source, and the highest concentrations would be located close to the tanks at the Basin F and SQI monitoring sites and diminish downwind. In fact, the highest levels were frequently measured at the southern and western perimeter sites (farthest from this potential source).

The CAQMMP investigated the possibility that the high levels of methylene chloride could be a laboratory contamination problem. Laboratory operations were audited and no discrepancies were noted, although methylene chloride has often been considered a laboratory contaminant. It should be noted, however, that there are shortcomings in the CMO4 analytical method for measuring VOCs, as discussed in Section 4.6.8. In addition to breakthroughs of the Tenax and Tenax/charcoal tubes, the concentrations measured were greater than the certified reporting ranges and consequently all concentrations provided in this discussion are "estimated" values so are not certified (estimated levels of methylene chloride have been reported previously in CAQMMP annual reports and the FY93 values are considerably higher).

A distinct possibility exists that the methylene chloride source is off-post. Table 4.6-8 indicates 75 VOC sources in excess of 1 ton per year, that have been identified in the Denver metropolitan area; many of these are clustered close to RMA (see Figure 4.6-13). A number of the VOC sources included methylene chloride, although precise emissions levels have not been officially required for reporting. Such reporting will, in fact, be required in 1994. CAQMMP will continue its investigations in FY94 and update its findings in the next annual report.

4.6.6 Metropolitan Denver Area VOC Emissions

Many of the ambient VOC concentrations measured at RMA interior monitoring stations during FY89, FY90, FY91, FY92, and FY93 were below values measured at Denver metropolitan

Table 4.6-8 Volatile Organic Compound Sources with Emissions greater than 15 Tons Per
Year in the RMA Vicinity

Page 1 of 2

| Facility | City | Update Year | Map No. | Estimate (TPY) |
|--|---------------|----------------|------------|-------------------|
| Chemical Systems Technology Inc. | Denver | 90 | 1 | 6655.0 |
| Colorado Refining Company / Total Petroleum | Commerce City | 91 | 2 | 988.6 |
| Conoco Inc. - Denver Refinery | Commerce City | 91 | 3 | 739.8 |
| Denver Industrial Sales & Service | Denver | 90 | 4 | 256.7 |
| Wyco Pipe Line Company Dupont Terminal | Dupont | 92 | 5 | 236.3 |
| Pillow Kingdom Mfg | Denver | 91 | 6 | 213.0 |
| Swedish Medical Center | Englewood | 91 | 7 | 155.2 |
| Union Chemicals Division Union Oil of Ca | Denver | 92 | 8 | 143.0 |
| Chase Terminal Co. | Aurora | 92 | 9 | 120.5 |
| Mastercraft | Aurora | 92 | 10 | 120.0 |
| The Gates Rubber Co. | Denver | 90 | 11 | 117.7 |
| Amoco Production Co. - Wattenberg Plant | Watkins | 92 | 12 | 115.8 |
| Schafer Commercial Seating Inc. | Denver | 92 | 13 | 100.0 |
| Central Products Co. | Brighton | 91 | 14 | 94.2 |
| The Electron Corporation | Littleton | 91 | 15 | 87.2 |
| Pillow Kingdom Mfg | Denver | 91 | 16 | 73.0 |
| Public Service Company - Cherokee | Denver | 91 | 17 | 71.8 |
| Atlas/Judds Brothers | Denver | 89 | 18 | 69.9 |
| Rainbo Bread Co | Commerce City | 92 | 19 | 65.0 |
| Diamond Shamrock Inc. | Commerce City | 91 | 20 | 62.1 |
| Sinclair Oil Denver Products Terminal | Henderson | 92 | 21 | 59.2 |
| Advance Foam Plastics Inc | Denver | 92 | 22 | 58.1 |
| Energy Thermal Systems Inc. | Denver | 90 | 23 | 55.6 |
| Colorado Interstate Gas Co - Watkins Station | Watkins | 91 | 24 | 55.0 |
| Accurate Plastics | Longmont | 92 | 25 | 49.0 |
| Lowry Air Force Base | Denver | 90 | 26 | 44.8 |
| Modern Plastics Inc. | Englewood | 91 | 27 | 42.8 |
| Emc-Henderson Terminal | Henderson | 92 | 28 | 42.0 |
| W.J. Whatley, Inc | Commerce City | 92 | 29 | 41.7 |
| E/M Corporation 12555 E. 37th Ave. | No city name | 90 | 30 | 37.0 |
| Pharmaceutical Basics, Inc | Denver | 91 | 31 | 35.6 |
| Hughes and Company | Commerce City | 90 | 32 | 33.6 |
| Phillips Pipeline Co | Commerce City | 91 | 33 | 33.2 |
| Power Assist Co. Inc. | Denver | 90 | 34 | 33.1 |
| The Denver Post Corp | Denver | 92 | 35 | 31.0 |
| Quik Print of Denver | Denver | 90 | 36 | 30.5 |
| AT&T Info. Systems | Westminster | 90 | 37 | 30.1 |
| Import Auto Body | Denver | 90 | 38 | 30.0 |

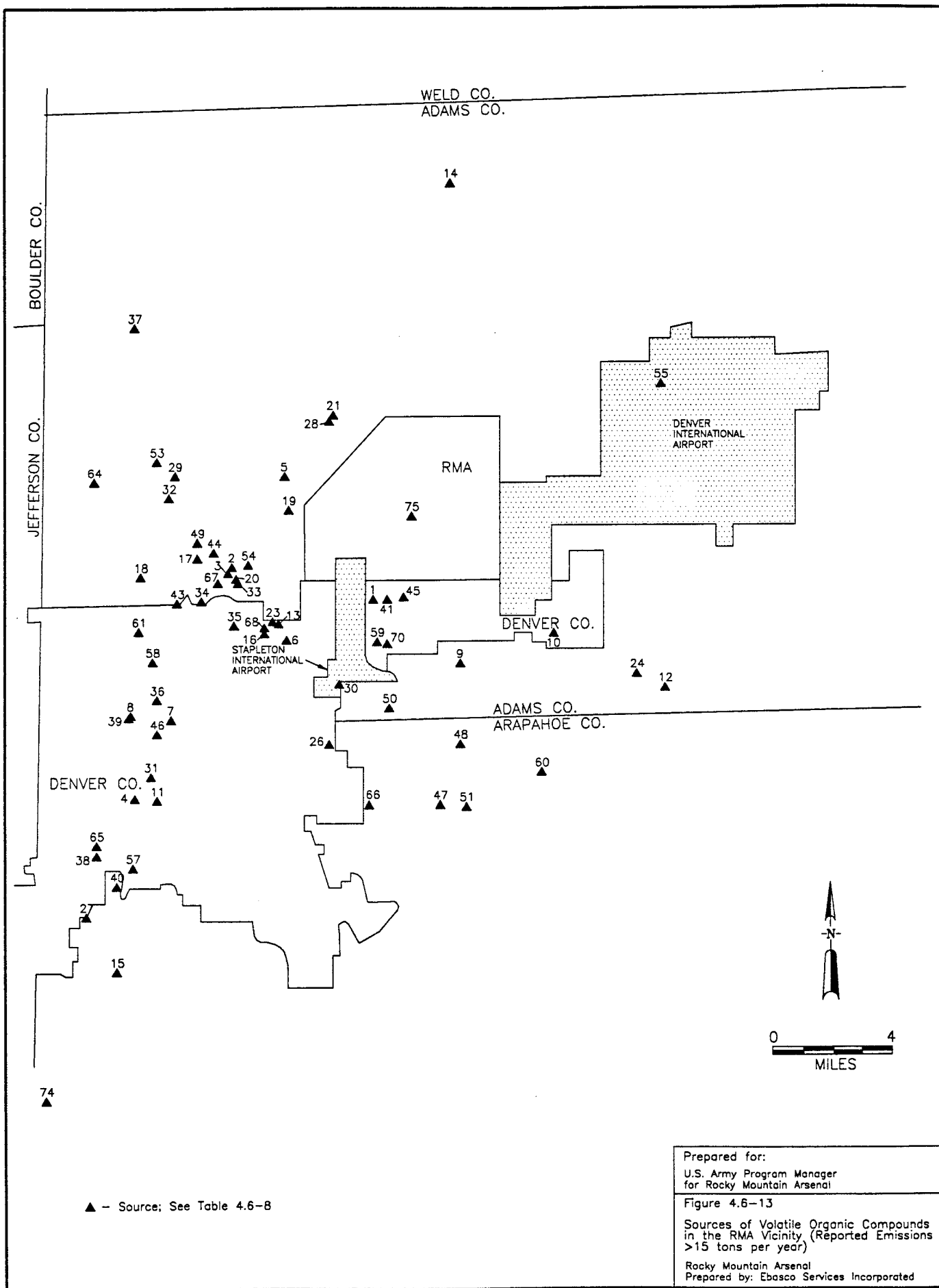
Source: CDH 1993

Table 4.6-8 Volatile Organic Compound Sources with Emissions greater than 15 Tons Per
Year in the RMA Vicinity

Page 2 of 2

| Facility | City | Update Year | Map No. | Estimate (TPY) |
|--|---------------|----------------|------------|-------------------|
| Cohler-Mclister | Denver | 90 | 39 | 29.2 |
| Sterling Stainless Tube Corp | Englewood | 88 | 40 | 29.0 |
| Pease Industries Inc Montbello Ind Park | Denver | 90 | 41 | 27.8 |
| Styro-Molders | Aurora | 92 | 42 | 27.6 |
| Ardco Corporation | Denver | 92 | 43 | 27.4 |
| Metro Wastewater Reclamation District | Denver | 91 | 44 | 25.4 |
| Conoco Inc. S. Federal Blvd. | No city name | 89 | 45 | 24.9 |
| Conoco 8073 S. Broadway Littleton | Littleton | 90 | 46 | 24.9 |
| Phillips 66 Co 13780 E 6th Ave Aurora | Aurora | 90 | 47 | 23.7 |
| Phillips 66-14401 E. Colfax Ave., Aurora | Westminster | 90 | 48 | 23.7 |
| Phillips 66 | Arvada | 92 | 49 | 23.3 |
| Silver State Cleaners | Denver | 92 | 50 | 22.0 |
| Amoco Gas Station #3691 | Aurora | 92 | 51 | 21.1 |
| Phoenix Press | Denver | 91 | 52 | 21.0 |
| Phillips 66 | Broomfield | 92 | 53 | 19.4 |
| United Asphalts Inc 4306 East 60th | No city name | 90 | 54 | 19.1 |
| Koch Hydrocarbon Co. Third Creek Plant | Brighton | 92 | 55 | 18.0 |
| Total Petroleum, Inc. #2498 | Denver | 92 | 56 | 17.9 |
| Public Service Company - Arapahoe | Denver | 91 | 57 | 17.4 |
| B-D Chemical Co. Inc. | Denver | 90 | 58 | 17.2 |
| Kwal Paints Inc. | Denver | 90 | 59 | 17.2 |
| Buckley ANG Base - Colo. Air Nat'l Guard | Aurora | 91 | 60 | 17.2 |
| Brannan S & G | Denver | 90 | 61 | 17.0 |
| Total Petroleum Inc. #2344 | Westminster | 91 | 62 | 17.0 |
| Total Petroleum Inc. #2355 | Denver | 91 | 63 | 16.7 |
| Exxon Corporation-7995 N Federal Blvd | Westminster | 90 | 64 | 16.6 |
| Exxon Corp 2110 S Federal Denver | Denver | 90 | 65 | 16.5 |
| Exxon Corp 1090 S Havana St Aurora | Aurora | 90 | 66 | 16.5 |
| Republic Paperboard Company | Commerce City | 92 | 67 | 16.5 |
| Stemco Manufacturing | Denver | 92 | 68 | 16.4 |
| Denver Forms | No city name | 91 | 69 | 16.0 |
| Frito Lay Inc. | Denver | 89 | 70 | 15.9 |
| Total Petroleum, Inc. #2463 | Littleton | 91 | 71 | 15.4 |
| Total Petroleum Inc. #2348 | Denver | 91 | 72 | 15.2 |
| Total Petroleum Inc. #2337 | Denver | 91 | 73 | 15.1 |
| T.A. Pelsue | No city name | 90 | 74 | 15.0 |
| Rocky Mountain Arsenal | Commerce City | 92 | 75 | 1.1 |

Source: CDH 1993



stations during various studies (See Tables 4.6-14 and 4.6-15 in the FY91 CMP Annual Assessment Report, and Tables 4.6-11 and 4.6-12 in this report). Measured ambient concentrations of various VOC analytes that were elevated during Basin F remediation activities decreased to typical urban background levels during the post-remedial monitoring periods, and have continued to occur at these levels during FY93. Some ambient VOC concentrations measured at the RMA monitoring sites during FY93 continued to be influenced by residual emissions from Basin F and from undisturbed areas of the South Plants and Basin A (as in previous post-remedial monitoring periods). However, many of the temporal differences in measured ambient VOC concentrations at CAQMMP sampling sites appeared to result from varying meteorological conditions that affected on-site transport of background emissions not located on RMA.

Table 4.6-8 and Figure 4.6-13 present numerous VOC point sources identified by CDH that may impact air quality monitored within RMA (CDH 1993). Emission sources included in this list were those exceeding 15 tpy of VOCs and represent approximately 80 percent of all VOC emissions identified in Adams, Arapahoe, and Denver counties. This list of sources included several major industries (Chemical Systems Technology, Inc., Colorado Refining Company, Conoco Denver Refinery, Sinclair Pipeline Company, Pillow Kingdom, Wyco Pipeline Company, Continental Airlines, and Gates Rubber Company) and numerous smaller industrial operations, chemical plants, paint manufacturing facilities, and large gasoline service stations. In addition, there are many VOC point sources below 10 tpy in the metropolitan Denver area that include dry cleaning establishments, paint stores, automobile maintenance shops, and small service stations. Background VOC emissions also originate from mobile sources including gasoline- and diesel-burning vehicles, construction equipment, and aircraft operating from nearby Stapleton International Airport (and the soon-to-be operational Denver International Airport), and are likely a significant portion of the total VOCs present in the metropolitan area. Another source of VOC emissions not readily quantifiable for documentation in an emissions inventory is from feedlot facilities located north of RMA.

Table 4.6-9 classifies releases of Denver area VOC sources by specific toxicity criteria according to health-related characteristics (Citizens Fund 1992). The release quantities are expressed in terms of total pounds of air, water, and any other release medium at a particular facility. The relative toxicity of each VOC release cannot be evaluated strictly by emissions quantity; other factors that need to be considered include chemical reactivity in the ambient air and the relative toxicity hazard of each VOC.

Specific VOC releases from metropolitan Denver emission sources are likely to have contributed to observed ambient VOC concentrations at the RMA air monitoring sites. Several of the major sources are within 2 miles of the RMA boundary such as Colorado Refining Company, Conoco Refinery, Sinclair Pipeline Company, Pillow Kingdom, Wyco Pipeline Company, and Continental Airlines. The number, the diversity of emissions, and the geographical extent of the emission sources complicates the attribution of specific off-site emission sources to observed VOC concentrations at specific RMA monitoring sites. As noted in various examples provided in this report, a number of observed maximum VOC concentrations at the RMA perimeter monitoring sites occurred concurrently with prevailing wind flow from potential off-site emission sources in the direction of the RMA monitoring sites. High-event monitoring during FY93 resulted in observed ambient concentrations of certain VOC analytes that were highest at perimeter sites AQ1, AQ2, AQ3, AQ4, and AQ5. The December 8, 1992 intense inversion episode, however, resulted in the pervasive influence of various VOC constituents across all of the RMA sampling sites, as was the possible incursions of methylene chloride during the late spring of FY93. As the CAQMMP continues to monitor on-going and future remediation activities, a continuing effort will be made to identify point and area sources both on and off post.

4.6.7 Summary of VOC Results

Table 4.6-10 summarizes maximum average and maximum 24-hour concentrations throughout the CMP/CAQMMP program history. The State of Colorado has not published air toxics regulations or promulgated standards. Promulgated state air toxics standards are not anticipated until amendments to Section 112 of the Clean Air Act are finalized in 1996.

Table 4.6-9 Total Pounds of Releases of Toxic Chemicals, 1990 by Facility and Toxicity
for Denver and Adams Counties

Page 1 of 3

| Facility | Zip | Carcinogen | Heritable Mutagen | Developmental Toxin | Reproductive Toxin | Acute Toxin | Chronic Toxin | Neuro- Toxin |
|---------------------------------------|-------|------------|----------------------|------------------------|-----------------------|----------------|------------------|-----------------|
| A.B. Hirschfield Press, Inc. | 80216 | 0 | 0 | 17,236 | 17,236 | 0 | 17,236 | 0 |
| A.R. Wifley & Sons Inc. | 80205 | 10 | 0 | 5 | 5 | 0 | 10 | 0 |
| Aaa Plating Inc. | 80205 | 0 | 0 | 0 | 0 | 750 | 750 | 0 |
| Altac Corp. | 80601 | 0 | 0 | 114,297 | 114,297 | 0 | 0 | 0 |
| Asarco Inc. Globe Plant | 80216 | 0 | 0 | 0 | 0 | 751 | 751 | 0 |
| Atlas Energy Products | 80229 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Boyles Galvanizing Co. | 80222 | 0 | 0 | 0 | 0 | 60,403 | 60,403 | 0 |
| Bryant Electric Inc. | 80216 | 250 | 0 | 0 | 0 | 0 | 250 | 0 |
| Cai Technologies | 80204 | 0 | 0 | 0 | 0 | 505 | 0 | 0 |
| Chemical & Metal Industries, Inc. | 80216 | 59,819 | 0 | 59,819 | 76,897 | 13,298 | 47,531 | 0 |
| Chemical Systems Technology Inc. | 80239 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Chemrex Inc. | 80022 | 6,960 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cobitco Inc. | 80216 | 0 | 0 | 0 | 0 | 95 | 95 | 0 |
| Coca-Cola Bottling Co. | 80205 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Colorado Paint Co. | 80216 | 0 | 0 | 1,500 | 1,500 | 0 | 1,010 | 500 |
| Colorado Refining Co. | 80022 | 10,155 | 0 | 24,275 | 19,255 | 4,026 | 20,321 | 0 |
| Conoco Denver Refinery | 80022 | 3,753 | 0 | 33,602 | 33,602 | 370 | 37,496 | 3,727 |
| Continental Airlines Base Maintenance | 80207 | 3,046 | 0 | 11,156 | 11,156 | 0 | 645,271 | 0 |
| Cryenco | 80229 | 0 | 0 | 21,100 | 21,100 | 0 | 0 | 0 |
| Delta Rocky Mountain Petroleum | 80640 | 0 | 0 | 0 | 0 | 0 | 10 | 0 |
| Denver Metal Finishing | 80216 | 0 | 0 | 9,865 | 9,865 | 0 | 0 | 0 |
| Dow Chemical Co. | 80010 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |

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Table 4.6-9 Total Pounds of Releases of Toxic Chemicals, 1990 by Facility and Toxicity
for Denver and Adams Counties

Page 2 of 3

| Facility | Zip | Carcinogen | Heritable Mutagen | Developmental Toxin | Reproductive Toxin | Acute Toxin | Chronic Toxin | Neuro- Toxin |
|-------------------------------|-------|------------|----------------------|------------------------|-----------------------|----------------|------------------|-----------------|
| DPC Industries Inc. | 80022 | 0 | 0 | 0 | 0 | 3,414 | 3,414 | 0 |
| Eaton Metal Products Co. Inc. | 80216 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Electromedics Inc. | 80112 | 17,502 | 17,502 | 17,502 | 17,502 | 17,502 | 190,490 | 0 |
| Electron Corp. | 80120 | 0 | 0 | 1,030 | 1,030 | 0 | 0 | 3,139 |
| Firestone Building Products | 80216 | 0 | 0 | 0 | 0 | 0 | 5 | 0 |
| Jackson Ice Cream Co. | 80204 | 0 | 0 | 0 | 0 | 5,500 | 5,500 | 0 |
| Koppers Industries Inc. | 80216 | 229,650 | 0 | 60 | 60 | 0 | 60 | 0 |
| Kwal-Howells Inc. | 80239 | 0 | 0 | 0 | 0 | 0 | 750 | 0 |
| Lakewood Brick & Tile Co. | 80214 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Majestic Metals Inc. | 80229 | 0 | 0 | 11,900 | 11,900 | 0 | 11,900 | 0 |
| Martin Marietta Astronautics | 80127 | 0 | 0 | 153,270 | 153,270 | 500 | 79,600 | 40,600 |
| Medtronic Hemotec Inc. | 80112 | 0 | 0 | 0 | 0 | 0 | 7,510 | 0 |
| Meas Fiberglass Inc. | 80022 | 5,840 | 5,840 | 0 | 0 | 0 | 5,840 | 0 |
| Metal Treating & Research Co. | 80229 | 0 | 0 | 0 | 0 | 0 | 0 | 500 |
| Metrum Information Storage | 80122 | 0 | 0 | 11,811 | 11,811 | 0 | 18,213 | 0 |
| Mid-America Plating Inc. | 80216 | 0 | 0 | 0 | 0 | 32 | 21 | 0 |
| Mile High Equipment Co. | 80239 | 0 | 0 | 250 | 250 | 64,500 | 64,500 | 0 |
| Modine Western-Rocky Mountain | 80223 | 0 | 0 | 109 | 109 | 0 | 0 | 0 |
| Ner Data Products Inc. | 80216 | 0 | 0 | 5,389 | 5,389 | 0 | 0 | 0 |
| Norgren Co. | 80120 | 0 | 0 | 0 | 0 | 5 | 5 | 0 |
| Pease Industries Inc. | 80239 | 0 | 0 | 109,000 | 109,000 | 0 | 76,000 | 76,000 |
| Pepsi-Cola Bottling Co. | 80216 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 4.6-9 Total Pounds of Releases of Toxic Chemicals, 1990 by Facility and Toxicity
for Denver and Adams Counties

Page 3 of 3

| Facility | Zip | Carcinogen | Heritable Mutagen | Developmental Toxin | Reproductive Toxin | Acute Toxin | Chronic Toxin | Neuro- Toxin |
|--------------------------------|-------|------------|----------------------|------------------------|-----------------------|----------------|------------------|-----------------|
| Protecto Wrap Co. | 80223 | 0 | 0 | 997 | 997 | 0 | 10 | 5 |
| Purina Mills Inc. | 80640 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Redfield Riflescopes Inc. | 80224 | 0 | 0 | 32,700 | 32,700 | 0 | 0 | 0 |
| RMO Inc. | 80204 | 0 | 0 | 0 | 0 | 4 | 4 | 0 |
| Safeway Milk Plant | 80216 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sampsonite Corp. | 80239 | 0 | 0 | 145,389 | 145,389 | 0 | 0 | 0 |
| Sas Circuits Inc. | 80127 | 0 | 0 | 0 | 0 | 500 | 500 | 0 |
| Sashco Inc. | 80022 | 0 | 0 | 250 | 250 | 0 | 0 | 0 |
| Sashco Inc. | 80223 | 0 | 0 | 0 | 0 | 0 | 1,000 | 0 |
| Scott's Liquid Gold Inc. | 80239 | 0 | 0 | 4,251 | 4,251 | 0 | 0 | 0 |
| Shafer Commercial Seating Inc. | 80216 | 350 | 350 | 44,341 | 44,655 | 36 | 48,515 | 21,656 |
| Silver Engineering Works Inc. | 80011 | 438 | 0 | 249 | 249 | 0 | 438 | 0 |
| Stanley Aviation Corp. | 80010 | 40,347 | 0 | 71,334 | 71,334 | 0 | 40,347 | 0 |
| Thompson Pipe & Steel Co. | 80216 | 0 | 0 | 0 | 0 | 0 | 0 | 250 |
| W.J. Whatley Inc. | 80022 | 2,400 | 2,400 | 0 | 0 | 0 | 42,871 | 0 |
| Whirlpool Kitchens Inc. | 80011 | 0 | 0 | 64,665 | 64,665 | 0 | 12,766 | 0 |
| Whitaker Power Storage Systems | 80207 | 0 | 0 | 0 | 0 | 215 | 215 | 0 |
| Wright & McGill Co. | 80216 | 30 | 0 | 30 | 30 | 40 | 70 | 0 |
| Zimmerman Metals Inc. | 80216 | 2,805 | 0 | 5 | 5 | 0 | 2,805 | 250 |

Table 4.6-10 Maximum Concentrations and Locations of Volatile Organic Compounds
During CAQMMP, CMP, Basin F, and IRA-F Programs

Page 1 of 1

| VOC | Maximum Long-Term Average ($\mu\text{g}/\text{m}^3$) | Location | Phase | Maximum 24-Hour Concentration ($\mu\text{g}/\text{m}^3$) | Location | Phase |
|--------------------------|---|----------|-------|---|----------|-------|
| 1,1,1-Trichloroethane | 10.53 | BF4 | P1 | 72.54 | BF1 | P1 |
| 1,1,2-Trichloroethane | 0.46 | BF2 | P2-S2 | 4.00 | BF2 | P2-S2 |
| 1,1-Dichloroethane | 0.14 | CMP/BF3 | P1 | 0.27 | CMP/BF3 | P1 |
| Trans-1,2-Dichloroethene | 0.02 | BF7 | P2-S1 | 0.20 | SQ2 | P7 |
| 1,2-Dichloroethane | 0.49 | AQ5 | P2-S2 | 3.80 | RIFS2 | P2-S1 |
| Bicycloheptadiene | 9.96 | BF2 | P1 | 39.46 | BF2 | P1 |
| Benzene | 10.53 | CMP/BF1 | P1 | 44.29 | CMP/BF1 | P1 |
| Carbon Tetrachloride | 1.99 | RIFS2 | P2-S1 | 9.76 | RIFS2 | P1 |
| Methylene Chloride | 24.01 | AQ5 | P3 | 560.00 | AQ1 | P7 |
| Chloroform | 6.95 | BF2 | P1 | 37.15 | BF2 | P2-S2 |
| Chlorobenzene | 0.47 | CMP/BF2 | P1 | 2.53 | CMP/BF2 | P1 |
| Dibromochloropropane | 2.90 | CMP/BF2 | P1 | 17.04 | CMP/BF2 | P1 |
| Dicyclopentadiene | 4.53 | BF2 | P1 | 29.12 | BF2 | P1 |
| Dimethyldisulfide | 7.86 | BF1 | P1 | 36.72 | BF1 | P1 |
| Ethylbenzene | 2.39 | CMP/BF2 | P1 | 14.12 | SQ1 | P7 |
| Toluene | 23.55 | BF2 | P1 | 158.04 | BF2 | P1 |
| Methyl isobutyl ketone | 0.47 | AQ6 | P3 | 1.13 | FC4 | P4 |
| Tetrachloroethane | 5.29 | BF1 | P2-S2 | 24.02 | BF1 | P2-S2 |
| Trichloroethene | 6.98 | BF1 | P2-S2 | 40.43 | BF1 | P2-S2 |
| Xylenes | 8.57 | BF2 | P1 | 51.39 | SQ1 | P7 |

Note: Data from both within and above the certified range are summarized here. Long-term refers to averages by phase.

- P1 = Phase 1, is from March 22 to December 12, 1988.
 P2-S1 = Phase 2, Stage 1, is from December 13, 1988 to February 15, 1989.
 P2-S2 = Phase 2, Stage 2, is from February 16, 1989 to May 5, 1989.
 P3 = Phase 3, is from May 6, 1989 to September 30, 1989.
 P4 = Phase 4, is from October 1, 1989 to September 30, 1990.
 P5 = Phase 5, is from October 1, 1990 to September 30, 1991.
 P6 = Phase 6, is from October 1, 1991 to September 30, 1992.
 P7 = Phase 7, is from October 1, 1992 to September 30, 1993.

$\mu\text{g}/\text{m}^3$ Microgram per cubic meter

4.6.7.1 Other Comparisons

EPA conducted an air toxic study in the Denver area during the summer of 1987 and the winter of 1987-1988 (EPA 1989b). This study was discussed in the CMP FY89 Contamination Assessment Report and a comparison was made between maximum VOC concentrations measured in the Denver study and those measured at RMA (RLSA 1990b). An update of the RMA results is provided in Table 4.6-11 (last column). It is also anticipated that with recent CDH and EPA emphasis on toxic organic compounds, an update of the Denver data will be shortly forthcoming.

Where toxic guidelines vary significantly, perhaps a more meaningful evaluation would be a comparison of the RMA-measured VOC compounds with those of typical urban VOC levels. Table 4.6-12 compares the CAQMMP VOC results for FY93 to published EPA studies for both the Denver metropolitan area and a summary of "air toxic" results from studies across the country. CAQMMP concentrations were generally found to be less than the concentrations reported by the EPA urban airshed studies. Nevertheless, while the average maximum levels of compounds identified as concerns at the RMA boundaries, and other compounds identified as potential source emissions from Basin F, were equivalent to or below Denver and other urban source levels, the intent of this assessment is not to suggest that some potentially toxic air compounds may not have been emitted from various RMA sources and remedial activities. Analysis of results again suggests continued monitoring of these compounds with additional emphasis placed on identifying specific RMA local sources from new remediation activities, as well as delineating these from regional emission sources. Concentrations of some compounds may be the result of previous RMA production activities; others are typical industrial and traffic related compounds and are released throughout the Denver urban metropolitan area. RMA VOC measurements need to be evaluated quantitatively in terms of both internal and external sources. The CAQMMP will continue to address this issue.

Table 4.6-11 Comparison of EPA Air Toxic Study and Rocky Mountain Arsenal Results
for Volatile Organic Compounds

Page 1 of 1

| Analyte | Maximum Concentration for Denver Study* ($\mu\text{g}/\text{m}^3$) | Maximum Concentration at RMA ($\mu\text{g}/\text{m}^3$) (Phases 1-3) | Maximum Concentration at RMA ($\mu\text{g}/\text{m}^3$) (Phase 4) | Maximum Concentration at RMA ($\mu\text{g}/\text{m}^3$) (Phase 5) | Maximum Concentration at RMA ($\mu\text{g}/\text{m}^3$) (Phase 6) | Maximum Concentration at RMA ($\mu\text{g}/\text{m}^3$) (Phase 7) |
|------------------------------------|--|--|---|---|---|---|
| 1,2-Dichloroethane | 49 | 4 | 0.2 | 0.2 | 0.8 | 0.3 |
| 1,1,1-Trichloroethane | 44 | 73 | 7.2 | 5.8 | 5.1 | 20 |
| Carbon Tetrachloride | 3 | 10 | 1.9 | 1.8 | 1.3 | 5.4 |
| Trichloroethene | 4 | 40 | 0.5 | 0.2 | 0.7 | 0.7 |
| Tetrachloroethene | 14 | 24 | 6.3 | 2.3 | 2.8 | 17 |
| Benzene | 83 | 44 | 7.5 | 9.1 | 5.4 | 34 |
| Toluene | 294 | 158 | 16.9 | 10.8 | 5.9 | 88 |
| Ethylbenzene | 22 | 13 | 5.3 | 2.2 | 2.6 | 14 |
| o-Xylene | 13 | 8 | 5.9 | -- | -- | -- |
| m- and p-Xylene (Total Xylenes) | 252 | 50 | 9.5 | 10.1 | 14.2 | 51 |
| Chlorobenzene | 83 | 3 | 0.1 | 0.1 | 0.1 | 0.1 |

* Source: U.S. EPA, 1989. Report on the Air Toxics Monitoring Program for the Denver Metropolitan Area. Integrated Environmental Management Project. Report One, Data Summary. Region VIII, Denver, Colorado

-- Not analyzed during this phase
 $\mu\text{g}/\text{m}^3$ Micrograms per cubic meter

Phase 1-3 March 22, 1988 to September 30, 1989.
Phase 4 October 1, 1989 to September 30, 1990.
Phase 5 October 1, 1990 to September 30, 1991.
Phase 6 October 1, 1991 to September 30, 1992.
Phase 7 October 1, 1992 to September 30, 1993.

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Table 4.6-12 Ambient Volatile Organic Compound Concentrations from Various Studies

Page 1 of 1

| Target Name | 1987 EPA NMOC Study Means $\mu\text{g}/\text{m}^3$ | 1988 EPA UATMP Mean $\mu\text{g}/\text{m}^3$ | FY93 CMP* Basin F (FC-2) Mean $\mu\text{g}/\text{m}^3$ | FY93 CMP* interior (AQ6) Mean $\mu\text{g}/\text{m}^3$ | FY 93 CMP* Perimeter (AQ2) Mean $\mu\text{g}/\text{m}^3$ |
|----------------------------|---|---|---|---|---|
| 1,1,1-Trichloroethane | 18.6 | 6.7 | 1.892 | 1.771 | 1.954 |
| 1,1,2-Trichloroethane | -- | 3.0 | 0.084 | 0.084 | 0.084 |
| 1,1-Dichloroethane | 13.7 | 2.0 | 0.028 | 0.028 | 0.028 |
| 1,2-Dichloroethane | 25.1 | 8.1 | 0.040 | 0.037 | 0.039 |
| Benzene | 6.5 | 5.2 | 2.851 | 2.070 | 3.178 |
| Bicycloheptadiene | -- | -- | 0.085 | 0.028 | 0.028 |
| Carbon Tetrachloride | -- | 2.1 | 0.743 | 0.656 | 0.758 |
| Chlorobenzene | 7.4 | 3.0 | 0.030 | 0.030 | 0.030 |
| Chloroform | -- | 30.8 | 1.755 | 0.304 | 0.585 |
| Dicyclopentadiene | -- | -- | 0.206 | 0.206 | 0.206 |
| Dimethylidulfide | -- | -- | 0.114 | 0.114 | 0.114 |
| Ethylbenzene | -- | 7.0 | 0.639 | 0.487 | 0.774 |
| Methylene Chloride | 78.4 | 18.1 | 7.667 | 18.036 | 14.178 |
| Methylisobutyl Ketone | -- | -- | 0.370 | 0.370 | 0.370 |
| m-Xylene | 59.1 | 18.7 | -- | -- | -- |
| N-Nitrosodimethylamine | -- | -- | -- | -- | -- |
| o-Xylene (12DBM) | 7.4 | 5.7 | -- | -- | -- |
| p-Xylene | 59.1 | 18.9 | 3.087 ⁺ | 2.334 ⁺ | 3.712 ⁺ |
| Tetrachloroethylene | 34.1 | 13.1 | 0.640 | 0.436 | 0.645 |
| Toluene | 32.0 | 16.4 | 4.776 | 4.020 | 5.980 |
| Trans-1,2-Dichloroethylene | 9.7 | 4.4 | -- | -- | -- |
| Trichloroethylene | 20.8 | 10.4 | 0.044 | 0.035 | 0.043 |

Source: U.S. EPA, 1988. 1987 Nonmethane Organic Compound and Air Toxics Monitoring Program, Final Report, Vol II, "Toxic Species." EPA-450/4-88-012. Research Triangle Park, North Carolina: U.S. EPA Office of Research and Development.

* Data from both within and above the certified range are reported here.

Legend: NMOC = Non methane Organic Compound
UATMP = Urban Air Toxics Monitoring Program
CMP = Comprehensive Monitoring Program

-- Not analyzed during this phase
+ Monitoring result based on total xylenes
 $\mu\text{g}/\text{m}^3$ Micrograms per cubic meter

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In general, compounds at RMA did not appear to present toxic health risks significantly different from the urban environment of metropolitan Denver. Organic compounds that were unique to Basin F and other remediation areas decreased rapidly with distance from these sources and were at lower levels at the RMA boundaries.

4.6.8 VOC Performance Evaluation

In response to the PMRMA Laboratory Support Division (LSD) request, a review of information pertaining to sampling and analysis of VOCs was carried out under the CAQMMP in order to evaluate the performance characteristics of currently utilized methods and to examine potential alternatives for any appropriate improvements. A report on this evaluation was submitted to LSD in May 1993. This section presents a general discussion of performance evaluation observations, an examination of collected data and potential adjustments, and the applicability of alternative methods.

4.6.8.1 Background Information

USATHAMA method CM04 is presently used to analyze VOCs sampled in ambient air. The method has undergone several changes in the course of the CMP and the follow-on CAQMMP at RMA. For example, prior to January 1991 VOCs were analyzed using Midwest Research Institute's (MRI) Method E-7, which employed a glass cartridge instead of the longer stainless steel cartridge currently used. A backup cartridge packed with both Tenax and activated carbon was also utilized by MRI. During the original CMP, the front and back cartridge concentrations were reported as a total concentration for the sample.

Another feature of the initial analysis procedure was that when sample measurements were above the CRL, concentrations were estimated by extrapolating the concentration curve above the certified range. Such results represented a "best estimate" of the measured concentration; the data were not certified in accordance with PMRMA/USATHAMA standards, but were reported as estimates in the CMP Annual Reports (1988, 1989, 1990, 1991) indicating maximum

concentrations. This procedure of estimating was useful in augmenting the long-term data record for baseline and remedial progress documentation.

Starting in January 1991, the Environmental Science and Engineering (ESE) laboratory introduced a new design for the Tenax and Tenax/carbon sampling cartridges. Stainless steel tubing, with approximately the same diameter but greater length than previously employed by MRI, was packed with approximately 3.5 grams of Tenax, which was a greater amount than that used in the glass cartridges; a back tube was also used and packed with Tenax and activated carbon. MRI's front cartridges were packed with 1.6 grams of Tenax; their back cartridges were packed with 1.0 gram of Tenax and 1.0 gram of activated carbon (approximate amounts). Along with the cartridge changes, the analytical procedure also changed. ESE's written method, CM04, stated that the front cartridge would be analyzed and approximately 10 to 20 percent of the back cartridges would be analyzed to maintain quality control and to account for any breakthrough of target analytes from the front to the back cartridge. If the back cartridge contained more than 15 percent of the weight detected on the front cartridge, another back cartridge from the same lot was to be analyzed.

ESE analyzed nearly 20 percent of the back cartridges for the FY91 air program. When analyzed, the detections from the back cartridge were automatically added to the front cartridges and reported as one concentration. (In FY93, because of the frequent occurrence of "breakthroughs", LSD required that the laboratory analyze all backup cartridges.)

While many of the target analytes were not seen as back tube detections, several back tube detections including 1,1,1-trichloroethane, benzene, carbon tetrachloride, and chloroform, were reported in excess of 30 percent. Also, many of the back cartridge detections were over the 15 percent breakthrough level indicator (20 percent is currently considered as the appropriate indicator) suggesting that significant breakthrough of these compounds was taking place. Consequently, the FY91 Report stated that VOC concentrations listed in the Installation

Restoration Data Management Information System (IRDMIS) were suspected to be low for certain compounds.

During the latter half of 1991, the CMP operator changed the flow rate from 200 standard cubic centimeters per minute (sccm, ref. 25°C, 760 mm Hg) to 150 sccm. A report provided by ESE (July 1992) confirmed the effectiveness of reducing breakthrough by decreasing the flow rate.

In response to PMRMA/LSD's request for evaluation of the performance characteristics of the current VOC sampling and analysis methods (and to examine potential alternative methods), the next approach was to evaluate the latest VOC data available from the laboratory (October 1992 to January 1993). These were then compared with the last set of data provided under ESE laboratory analysis (FY91). Results again confirmed the effectiveness of decreasing breakthrough by reducing flow rates, but also indicated an increase in "greater than" (above CRL) measurements, as well as an increased frequency of concentrations below the detection limits for certain analytes. Key findings were as follows:

- (1) With respect to breakthrough, improvements were obtained as a result of decreasing the target volume from flowrates of 200 sccm to 150 sccm. This improvement was especially noticeable for benzene, 1,2-dichloroethane, trans-1,2-dichloroethene, and total xylenes. Of the certified analytes, carbon tetrachloride, chloroform, and 1,1,1-trichloroethane continued to exhibit breakthrough levels greater than 20 percent at frequencies of 16 percent or greater. In the transition to the reduced sample volume, breakthrough levels and frequency of breakthrough occurrence also increased substantially for methylene chloride, which is reported as a non-certified analyte.
- (2) Following reduction of the target sample volume, exceedence of the upper CRL occurred more frequently for certain analytes. Compounds exhibiting the greatest incidence of CRL exceedence included benzene, toluene, 1,1,1-trichloroethane, and total xylenes. The increased frequency of CRL exceedence for these compounds may be associated with

improved collection performance on the front Tenax cartridge as a result of the lowered sampling volume.

- (3) There are a large number of variables involved in selecting the most appropriate method for VOC analysis. The technical analysis provided to RMA concerning possible changes and improvements to the existing CMO4 method indicated that an alternative VOC analysis method must consider factors such as method reliability, consistency with past remedial progress data, practicality, economics, application of the latest state-of-the-art techniques, and most important, the objectives of the CAQMMP. These considerations are discussed below.

4.6.8.2 Pertinent Technical Guidance Factors

Guidance for sampling VOCs with Tenax-GC is given in Method TO-1 of the EPA Compendium of Methods for the Determination of Toxic Organic Compounds in Ambient Air. Method TO-1 indicates that, for the purpose of general guidance, breakthrough levels of less than 20 percent are acceptable. According to Method IP-1B in the EPA Compendium of Methods for the Determination of Air Pollutants in Indoor Air, which pertains to sampling VOCs with Tenax, temperature is often the major factor influencing breakthrough volumes for an analyte when its ambient concentration is less than 100 ppm. Given that the typical ambient concentration of VOCs at RMA is less than 100 ppm, temperature and sample volume will chiefly dictate the point at which breakthrough would occur.

A complicating factor in assessing actual breakthrough levels in the CM04 method is the presence of a dissimilar adsorbent in the back cartridge. The carbon packing in the backup cartridge broadens the potential for increased adsorption of very volatile compounds such as chloroform and 1,1,1-trichloroethane in that section. As such, the determination of breakthrough performance of the VOC sample cartridge sets is not totally straightforward for some of the highly volatile compounds that would have low breakthrough volumes with Tenax and high breakthrough volume with activated carbon.

Methods TO-1 and IP-1B provide estimated breakthrough volumes at 100°F for many of the analytes in the CM04 list. The methods state that the estimated breakthrough volumes may be used as a rough guidance and should probably be reduced by a certain factor to accommodate their approximate accuracy. For the purposes of this review, "safe" sampling volumes for each compound were calculated from the TO-1 and IP-1B suggested volumes using a reduction factor of 33 percent and incorporated a front cartridge and sorbent weight of 4 grams. Based on these calculations, it appeared that the "safe" sampling volumes of several of the volatile halogenated organics in the CM04 analyte list are less than the present sampling volume.

Additional factors to be considered in the use of solid adsorbents for VOC sampling are the potential formation of artifacts under certain conditions and the importance of analyte polarity. As indicated in Method CM04, the presence of oxidants in the sample atmosphere can produce artifacts through reactions with the organics adsorbed on Tenax-GC. Activated carbon is also cited in the EPA compendiums as being prone to artifact formation. This effect may result in the formation of detectable levels of compounds that were not originally present in the ambient air. A complete evaluation of artifact effects and their significance in this program would necessitate further detailed review. However, it is not apparent at this time that artifact formation has been significant in the analytical results. Method TO-1 is described as being applicable to nonpolar compounds. Adsorbent collection and adsorption efficiencies and analytical performance may be affected to varying degrees by the polarity of an analyte. When using the TO-1 or CM04 types of methodologies, nonpolar compounds are generally considered to be less problematic than polar compounds with respect to sampling and analysis performance. Most of the compounds in the CM04 method have nonpolar characteristics with respect to the sampling and analysis approaches, with the exception of methyl isobutyl ketone, and perhaps dicyclopentadiene and bicycloheptadiene.

4.6.8.3 Potential CM04 Method Adjustments and Alternatives

Among the approaches considered for improving the existing CM04 method for VOC collection and analyses, were changes to the sampling analyses and variations of the adsorbents used in the

sampling analyses. The difficulty appears to be that those factors that are optimum for one group of analytes are inappropriate for another group of the 20 target analytes.

The combined factors of "safe" breakthrough volume, range of CM04 certified reporting limits, and observed measurement performance variations suggest the need for distributive sampling at multiple flow rates. There would be a need for special testing using collocated sampling at multiple flow rates and using different types of adsorbents; the distributive sampling approach would require the initial determination of breakthrough volumes and certifiable reporting ranges for each analyte, preferably based upon gas-phase spike testing. The interpretation of data collected routinely from collocated distributed volume samples would entail a substantial amount of additional effort in order to select the most representative values. It would also involve equipment revisions, significantly increased media and analytical costs, and added data reduction and reporting. Nevertheless, Methods TO-1 and IP-1B state that distributive sampling is typically recommended for VOC sampling programs.

4.6.8.4 Whole-Air Sampling

Whole-air sampling (i.e., canisters) involves the collection of sample air directly into a rigid or non-rigid container. Rigid containers include glass sample bulbs and passivated stainless steel canisters, and non-rigid containers include Tedlar or Teflon sampling bags. Sampling bags are seldom used in situations where reliable retention of very low concentration VOCs is critical, especially if analysis is performed several hours or days after sample collection. Passivated canisters are now commonly used for sampling VOCs and some SVOCs at both low and high concentrations. Passivated canisters are popularly referred to as "Summa canisters" referring to the proprietary metallurgical electropolishing process developed by Molectrics, Inc. The electropolishing of the sampling canisters' inner surface substantially enhances preservation of sampling integrity. Holding times for samples collected in Summa passivated canisters are typically longer than those for adsorbent cartridges. Multiple analyses are possible with canister samples. Sample collection with canisters is performed either by pressurization with a pumping system or vacuum collection with a low flow regulator. Many configurations of sampling

systems have been developed commercially and some have been based upon EPA concepts. The important factors in canister sampling for RMA purposes are cleanliness inside the canisters, a sample delivery system free of contamination, and a constant flow rate during the sampling period. The simplest approach for 24-hour sampling involves the use of a 15L evacuated canister and a stainless steel vacuum regulator. Some users of this configuration have reported successful results in constant flow control over a 24-hour sampling period. An alternative configuration could involve the use of a pump in the sampling system to pressurize the canister.

A variety of methods have been developed for canister sample analysis. Typically, these have included GCMS-Scan, GCMS-Selective Ion Mode (SIM), or GCMS-Multidetector with a front end water management system to avoid cryotrap freeze-up. The water management systems usually include a Nafion dryer, which removes water from the sample stream, but while removing water with a Nafion dryer, polar compounds may also be removed. Such problems have been reduced or eliminated with the use of other apparatus recently developed by various manufacturers.

Disadvantages and advantages of sampling with passivated canisters are further discussed in Section 4.6.9.

4.6.8.5 Summary

The following is a summary of conclusions and recommendations pertaining to the improvement of the CAQMMP VOC sampling program:

- The existing VOC sampling and analytical method has certain limitations; additional tests and comparisons are needed to optimize the existing method or to select an alternative method.
- Reducing the volume flow rate from 200 sccm to 150 sccm has effectively reduced the front-to-back cartridge breakthrough percentage; several analytes, however, still exceed

the recommended breakthrough limit of 20 percent. In addition, several analytes now exhibit increased "greater than" levels.

- If the laboratory or the CM04 Method can accommodate a wider certified range, some of the negative effects of reduced flow might be eliminated. This could be investigated, but there are technical difficulties with this approach.
- Distributive sampling (multiple flow rates) is an EPA recognized procedure for VOC sampling where multiple analytes of different volatility are involved. The fact that the reduced flow rate produced more "greater than" and some decreased "greater than" CRL values, as well as more frequent "less than" and some decreased "less than" CRL values, attest to the unique physical characteristic of each of the RMA target compounds. Distributive sampling would address the physical qualities of each analyte and optimize the CM04 method. This approach would require increased development, and operational and analytical costs. It should be noted that the current data reviewed in this evaluation apply to samples collected during lower temperature winter time conditions, which is also a factor in performance.
- The canister approach is a recognized and viable alternative method for VOC sampling. The method has gained acceptance and may be effectively incorporated into the CAQMMP. An increased development effort would be required to ensure reliability and acceptability of the methods for RMA purposes, but operation costs would probably be comparable or less, once procedures were established.
- A final concern is the impact of any VOC methods change on the CMP/CAQMMP database. Improved methods may influence conclusions drawn concerning potential RMA toxic impacts and remedial progress to date. It should be noted, however, that the limitations of the VOC database have been consistently documented in previous

CMP/CAQMMP annual reports. Consequently, improvement of this database remains a desirable goal of the program.

4.6.9 Canister Versus Tube VOC Sampling

4.6.9.1 Background

The sampling and analysis of VOCs has been an integral part of the CMP and the present CAQMMP at RMA. The general VOC sampling method used at RMA has always consisted of drawing air samples through solid sorbents packed in either glass tubes (during the CMP) or stainless steel tubes (during the CAQMMP). The front tube contains Tenax sorbent; the rear tube contains Tenax followed by charcoal. The samples are typically taken for 24 hours at flow rates of 200 standard cubic centimeters per minute (sccm) for the CMP to 150 sccm for the CAQMMP. As discussed in the previous section, this tube sampling method suffers from problems with potential sample breakthrough, artifact formation, and fixed upper CRLs (see the more detailed discussion in Section 4.6.8). For example, in FY93 there were more than 350 exceedances of the upper CRL, which have been reported as "greater than" (GT) CRL values in the IRDMIS data base. Advantages and disadvantages of the method are presented in Table 4.6-13.

An alternative VOC sampling method that is gaining wide acceptance and avoids the major problems associated with sorbent tube sampling uses passivated stainless steel canisters. Whole air samples can be collected with this method for either short or integrated time periods (such as 24 hours). The samples are cryogenically concentrated in the laboratory before analysis. Advantages of using of canisters include the elimination of problems associated with sorbent breakthrough and the need for optimizing flow rates or utilizing distributive sampling with multiple tubes. Another advantage is that in general only one sample will need to be analyzed using canisters whereas the sorbent tube method requires that both the front and back tubes be analyzed which typically doubles analytical costs. Additionally, the canister method allows multiple analytical runs to be made on a single sample, if necessary, and allows for high end

Table 4.6-13 Tenax® Sampling Medium

| Methods | Application | Advantages | Disadvantages |
|-------------------|---|--|--|
| TO-1 Ambient Vost | Volatile, nonpolar organics boiling point 80° - 200°C | <ol style="list-style-type: none"> 1) Compact sampling tubes, especially stainless steel. 2) Easy to handle and ship. 3) Analytical equipment commercially available. 4) May use standards in methanol solution or static dilution bulb. 5) Field-ready tubes can be purchased. 6) Low Detection Limits by concentrating large sample volumes. | <ol style="list-style-type: none"> 1) Polymer contaminates easily. 2) Breakthrough volumes largely unknown. 3) "one-Shot" analysis. 4) Analysis produces artifacts. 5) Distributed volumes backup tube(s) required. 6) Low breakthrough for very volatile compounds. 7) Samples contain water. 8) Cost due to the need for distributed volume sampling and analysis. |

Source: Merrill, R.G. Jr., et al (1990). "Practical Experience in Analysis of Organic Compounds in Ambient Air Using Canisters and Sorbents". AWMA 83rd Annual Meeting, Paper 90-170.4, Pittsburgh, PA., June 24-29.

Table 4.6-14 Summa® Polished Canister Sampling Medium

| Methods | Application | Advantages | Disadvantages |
|----------------|--|--|--|
| TO-12 TO-14 | Volatile and Semi-volatile organic compounds | <ol style="list-style-type: none"> 1) Canister allows multiple analyses. 2) Can be screened, diluted. 3) Rugged for field handling, shipping. 4) Field-ready canisters, sampling equipment can be rented. 5) Sampling can be done without power. 6) Samples as taken accurately represent field atmosphere. 7) No need to collect distributed volume samples (no breakthrough). | <ol style="list-style-type: none"> 1) High initial cost to purchase. 2) Much of the equipment required to support canister use is not commercially available. 3) Canisters are bulky to ship. 4) Canisters may produce analytical artifacts. 5) Samples contain a large amount of water. 6) Higher detection limits than most sorbent methods. |

Source: Merrill, R.G. Jr., et al (1990). "Practical Experience in Analysis of Organic Compounds in Ambient Air Using Canisters and Sorbents". AWMA 83rd Annual Meeting, Paper 90-170.4, Pittsburgh, PA., June 24-29.

quantification. Table 4.6-14 presents the major advantages and disadvantages of the canister sampling method.

4.6.9.2 Canister Versus Tube Test Results

During FY93, PMRMA authorized the testing and comparison of the CM04 method for VOC analysis using Tenax and Tenax/charcoal tubes versus Summa canisters. Intensive testing was conducted on three routine (6-day) monitoring periods, August 29, September 4, and September 10, 1993. Collocated pairs of tubes and canisters were placed at six RMA monitoring sites: AQ2, AQ3, AQ5, AQ6, SQ1, and SQ2; an additional collocated canister was placed at SQ2.

The results of the comparison for each analyte compound for each monitoring day and at each sampling site are shown sequentially in Table 4.6-15; it is difficult to provide a quantitative summary of these comparisons, although the percent difference of the two methods for each collocated analysis is provided in the table. As noted, there are many cases where the tube measurement was below the CRL or the canister result was below the laboratory detection limit; these cases are noted as "NA" in the percent difference column of Table 4.6-15, although a qualitative evaluation can also be discerned from these data.

Many results are comparable, while others show excessive differences. In particular, the data on September 4 show great variability. It is believed that the CM04 tube data on this date are suspect as they are internally inconsistent and are also much larger than the canister results (nevertheless, all results are provided in this summary). There are a number of cases where the tubes indicated low values and the canisters were "below detection", but at a higher level than the tubes. This might be considered a good match, however these cases are indicated as N/A on Table 4.6-15; it does indicate a present advantage of CM04 tube method as the tubes can generally be certified at lower levels than canisters.

On the other hand, except for the September 4, 1993 data, the canister results were on average higher than the tubes; this is believed to be a valid attribute of the canister method. Also, there

Table 4.6-15 Tenax and Tenax/Charcoal Tubes Versus Canister Volatile Organic Compound Results

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| Date | Site identification | Analyte | Tubes ($\mu\text{g}/\text{m}^3$) | Canister ($\mu\text{g}/\text{m}^3$) | Percent Difference |
|----------|---------------------|---------|---------------------------------------|--|-----------------------|
| 08/29/93 | AQ2 | 111TCE | 1.210 | 1.254 | -4 |
| 08/29/93 | AQ3 | 111TCE | 0.990 | 0.982 | 1 |
| 08/29/93 | AQ5 | 111TCE | 0.960 | 1.309 | -27 |
| 08/29/93 | AQ6 | 111TCE | 0.740 | 1.145 | -35 |
| 08/29/93 | SQ1 | 111TCE | 0.830 | 1.145 | -27 |
| 08/29/93 | SQ2 | 111TCE | 1.210 | 1.145 | 6 |
| 09/04/93 | AQ2 | 111TCE | 4.800 | 0.545 | NA |
| 09/04/93 | AQ3 | 111TCE | 4.000 | 1.091 | 267 |
| 09/10/93 | SQ1 | 111TCE | 1.240 | 1.091 | 14 |
| 09/12/93 | M526SE | 111TCE | 0.850 | 0.982 | -13 |
| 09/12/93 | M326E | 111TCE | 0.909 | 6.163 | -85 |
| 09/10/93 | AQ2 | 111TCE | 1.327 | 5.399 | -75 |
| 09/10/93 | AQ3 | 111TCE | 0.875 | 0.927 | -6 |
| 09/10/93 | SQ2 | 111TCE | 1.420 | 1.145 | 24 |
| 08/29/93 | AQ2 | C6H6 | 2.220 | 2.171 | 2 |
| 08/29/93 | AQ3 | C6H6 | 1.420 | 1.437 | -1 |
| 08/29/93 | AQ5 | C6H6 | 0.520 | 1.788 | -71 |
| 08/29/93 | AQ6 | C6H6 | 0.230 | 1.437 | -84 |
| 08/29/93 | SQ1 | C6H6 | 1.110 | 1.501 | -26 |
| 08/29/93 | SQ2 | C6H6 | 1.310 | 1.501 | -13 |
| 09/04/93 | AQ2 | C6H6 | 11.900 | 0.319 | NA |
| 09/04/93 | AQ3 | C6H6 | 9.800 | 1.181 | 730 |
| 09/10/93 | SQ1 | C6H6 | 1.080 | 0.990 | 9 |
| 09/12/93 | M526SE | C6H6 | 1.240 | 1.469 | -2 |
| 09/12/93 | M326E | C6H6 | 1.516 | 1.724 | -12 |
| 09/10/93 | AQ2 | C6H6 | 1.429 | 1.596 | -10 |
| 09/10/93 | AQ3 | C6H6 | 0.880 | 1.980 | -56 |
| 09/10/93 | SQ2 | C6H6 | 1.119 | 0.958 | 17 |
| 08/29/93 | AQ2 | CCL4 | 0.540 | 0.629 | -14 |
| 08/29/93 | AQ3 | CCL4 | 0.550 | 0.629 | -13 |
| 08/29/93 | AQ5 | CCL4 | 0.315 | 0.629 | NA |
| 08/29/93 | AQ6 | CCL4 | 0.340 | 0.629 | NA |
| 08/29/93 | SQ1 | CCL4 | 0.690 | 0.629 | 10 |
| 08/29/93 | SQ2 | CCL4 | 0.550 | 0.629 | -13 |
| 09/04/93 | AQ2 | CCL4 | 5.400 | 0.629 | NA |
| 09/04/93 | AQ3 | CCL4 | 1.500 | 0.755 | 99 |
| 09/10/93 | SQ1 | CCL4 | 0.700 | 0.755 | -7 |
| 09/12/93 | M526SE | CCL4 | 0.450 | 0.692 | -35 |
| 09/12/93 | M326E | CCL4 | 0.633 | 0.692 | -9 |
| 09/10/93 | AQ2 | CCL4 | 0.664 | 0.755 | -12 |
| 09/10/93 | AQ3 | CCL4 | 0.539 | 0.692 | -22 |
| 09/10/93 | SQ2 | CCL4 | 0.833 | 0.692 | 20 |
| 08/29/93 | AQ2 | CH2CL2 | LT 0.463 | 2.673 | NA |
| 08/29/93 | AQ3 | CH2CL2 | LT 0.463 | 2.846 | NA |
| 08/29/93 | AQ5 | CH2CL2 | LT 0.463 | 2.291 | NA |
| 08/29/93 | AQ6 | CH2CL2 | LT 0.463 | 2.326 | NA |

LT ### - Less than certified reporting limit
 NA - Not applicable
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter
 CHCL3 - Chloroform
 CH2CL2 - Methylene chloride
 ETC6H5 - Ethylbenzene
 MECGH5 - Toluene
 TLLEE - Tetrachloroethene
 Xylenes - Total xylenes

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Table 4.6-15 Tenax and Tenax/Charcoal Tubes Versus Canister Volatile Organic Compound Results
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| Date | Site identification | Analyte | | Tubes ($\mu\text{g}/\text{m}^3$) | | Canister ($\mu\text{g}/\text{m}^3$) | Percent Difference |
|----------|---------------------|---------|----|---------------------------------------|----|--|-----------------------|
| 08/29/93 | SQ1 | CH2CL2 | | 0.480 | | 1.909 | -75 |
| 08/29/93 | SQ2 | CH2CL2 | LT | 0.463 | | 2.083 | NA |
| 09/04/93 | AQ2 | CH2CL2 | | 2.700 | LT | 0.347 | NA |
| 09/12/93 | M326E | CH2CL2 | LT | 0.463 | | 0.555 | NA |
| 09/10/93 | AQ2 | CH2CL2 | | 1.067 | | 1.111 | -4 |
| 09/10/93 | AQ3 | CH2CL2 | | 2.043 | | 0.417 | 390 |
| 09/10/93 | SQ2 | CH2CL2 | | 2.643 | | 0.486 | 444 |
| 08/29/93 | AQ2 | CHCL3 | LT | 0.056 | LT | 0.488 | NA |
| 08/29/93 | AQ3 | CHCL3 | | 0.130 | LT | 0.488 | NA |
| 08/29/93 | AQ5 | CHCL3 | | 0.090 | LT | 0.488 | NA |
| 08/29/93 | AQ6 | CHCL3 | LT | 0.056 | LT | 0.488 | NA |
| 08/29/93 | SQ1 | CHCL3 | | 0.760 | | 0.488 | 56 |
| 08/29/93 | SQ2 | CHCL3 | | 0.508 | | 0.683 | -26 |
| 09/04/93 | AQ2 | CHCL3 | | 13.570 | LT | 0.488 | NA |
| 09/04/93 | AQ3 | CHCL3 | | 1.070 | LT | 0.488 | NA |
| 09/10/93 | SQ1 | CHCL3 | | 0.350 | LT | 0.488 | NA |
| 09/12/93 | M526SE | CHCL3 | | 0.250 | LT | 0.488 | NA |
| 09/12/93 | M326E | CHCL3 | LT | 0.056 | LT | 0.488 | NA |
| 09/10/93 | AQ2 | CHCL3 | | 0.220 | LT | 0.488 | NA |
| 09/10/93 | AQ3 | CHCL3 | | 0.240 | LT | 0.488 | NA |
| 09/10/93 | SQ2 | CHCL3 | | 0.437 | LT | 0.488 | NA |
| 08/29/93 | AQ2 | ETC6H5 | | 0.810 | | 0.695 | 17 |
| 08/29/93 | AQ3 | ETC6H5 | | 0.640 | LT | 0.434 | NA |
| 08/29/93 | AQ5 | ETC6H5 | LT | 0.148 | | 0.608 | NA |
| 08/29/93 | AQ6 | ETC6H5 | LT | 0.148 | | 0.478 | NA |
| 08/29/93 | SQ1 | ETC6H5 | LT | 0.148 | | 0.478 | NA |
| 08/29/93 | SQ2 | ETC6H5 | LT | 0.148 | | 0.478 | NA |
| 09/04/93 | AQ2 | ETC6H5 | | 2.800 | LT | 0.434 | NA |
| 09/04/93 | AQ3 | ETC6H5 | | 5.000 | | 1.563 | 220 |
| 09/10/93 | SQ1 | ETC6H5 | | 0.330 | LT | 0.434 | NA |
| 09/12/93 | M526SE | ETC6H5 | | 0.560 | | 0.608 | -8 |
| 09/12/93 | M326E | ETC6H5 | | 0.339 | | 1.303 | -74 |
| 09/10/93 | AQ2 | ETC6H5 | | 0.493 | | 1.042 | -53 |
| 09/10/93 | AQ3 | ETC6H5 | | 0.318 | LT | 0.434 | NA |
| 09/10/93 | SQ2 | ETC6H5 | | 0.423 | LT | 0.434 | NA |
| 08/29/93 | AQ2 | MEC6H5 | | 5.700 | | 7.681 | -26 |
| 08/29/93 | AQ3 | MEC6H5 | | 3.900 | | 3.577 | 9 |
| 08/29/93 | AQ5 | MEC6H5 | | 0.170 | | 5.460 | -97 |
| 08/29/93 | AQ6 | MEC6H5 | | 0.150 | | 6.853 | -98 |
| 08/29/93 | SQ1 | MEC6H5 | | 0.310 | | 4.255 | -93 |
| 08/29/93 | SQ2 | MEC6H5 | | 0.410 | | 3.878 | -89 |
| 09/10/93 | SQ1 | MEC6H5 | | 2.600 | | 3.163 | -18 |
| 09/12/93 | M526SE | MEC6H5 | | 3.400 | | 4.255 | -20 |
| 09/12/93 | M326E | MEC6H5 | | 2.000 | | 14.459 | -86 |
| 08/29/93 | AQ2 | TCLEE | | 0.470 | | 0.949 | -50 |
| 08/29/93 | AQ3 | TCLEE | | 0.390 | LT | 0.678 | NA |

LT ### - Less than certified reporting limit
 NA - Not applicable
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter
 CHCL3 - Chloroform
 CH2CL2 - Methylene chloride
 ETC6H5 - Ethylbenzene
 MEC6H5 - Toluene
 TCLEE - Tetrachloroethene

Table 4.6-15 Tenax and Tenax/Charcoal Tubes Versus Canister Volatile Organic Compound Results

Page 3 of 3

| Date | Site identification | Analyte | | Tubes ($\mu\text{g}/\text{m}^3$) | Canister ($\mu\text{g}/\text{m}^3$) | Percent Difference |
|----------|---------------------|---------|----|---------------------------------------|--|-----------------------|
| 08/29/93 | AQ5 | TCLEE | LT | 0.060 | LT 0.678 | NA |
| 08/29/93 | AQ6 | TCLEE | LT | 0.060 | LT 0.678 | NA |
| 08/29/93 | SQ1 | TCLEE | LT | 0.060 | LT 0.678 | NA |
| 08/29/93 | SQ2 | TCLEE | LT | 0.060 | LT 0.678 | NA |
| 09/10/93 | SQ1 | TCLEE | | 0.290 | LT 0.678 | NA |
| 09/10/93 | AQ2 | TCLEE | | 0.314 | 0.746 | -58 |
| 09/10/93 | AQ3 | TCLEE | | 0.220 | LT 0.678 | NA |
| 09/10/93 | SQ2 | TCLEE | | 0.268 | LT 0.678 | NA |
| 08/29/93 | AQ2 | TRCLE | LT | 0.056 | LT 0.537 | NA |
| 08/29/93 | AQ3 | TRCLE | LT | 0.056 | 0.860 | NA |
| 08/29/93 | AQ5 | TRCLE | LT | 0.056 | LT 0.537 | NA |
| 08/29/93 | AQ6 | TRCLE | LT | 0.056 | 0.732 | NA |
| 08/29/93 | SQ1 | TRCLE | LT | 0.056 | LT 0.537 | NA |
| 08/29/93 | SQ2 | TRCLE | LT | 0.056 | 0.591 | NA |
| 09/04/93 | AQ2 | TRCLE | | 0.380 | LT 0.537 | NA |
| 09/04/93 | AQ3 | TRCLE | | 0.150 | 1.733 | -91 |
| 09/10/93 | SQ1 | TRCLE | LT | 0.056 | LT 0.537 | NA |
| 09/12/93 | M526SE | TRCLE | LT | 0.056 | 1.655 | NA |
| 09/12/93 | M326E | TRCLE | LT | 0.056 | LT 0.537 | NA |
| 09/10/93 | AQ2 | TRCLE | LT | 0.056 | LT 0.537 | NA |
| 09/10/93 | AQ3 | TRCLE | LT | 0.056 | 0.645 | NA |
| 09/10/93 | SQ2 | TRCLE | LT | 0.056 | 0.806 | NA |
| 08/29/93 | AQ2 | XYLENES | | 3.600 | 3.170 | 14 |
| 08/29/93 | AQ3 | XYLENES | | 3.200 | 1.867 | 71 |
| 08/29/93 | AQ5 | XYLENES | LT | 0.690 | 2.779 | NA |
| 08/29/93 | AQ6 | XYLENES | LT | 0.690 | 2.214 | NA |
| 08/29/93 | SQ1 | XYLENES | LT | 0.690 | 2.214 | NA |
| 08/29/93 | SQ2 | XYLENES | LT | 0.690 | 2.214 | NA |
| 09/04/93 | AQ2 | XYLENES | | 12.000 | LT 0.434 | NA |
| 09/04/93 | AQ3 | XYLENES | | 23.000 | 2.084 | 1004 |
| 09/10/93 | SQ1 | XYLENES | | 1.500 | 1.824 | -18 |
| 09/12/93 | M526SE | XYLENES | | 2.700 | 3.170 | -15 |
| 09/12/93 | M326E | XYLENES | | 1.842 | 6.686 | -72 |
| 09/10/93 | AQ2 | XYLENES | | 2.407 | 4.038 | -40 |
| 09/10/93 | AQ3 | XYLENES | | 1.519 | 1.824 | -17 |
| 09/10/93 | SQ2 | XYLENES | | 2.276 | 1.997 | 14 |

LT ### - Less than certified reporting limit
 NA - Not applicable
 $\mu\text{g}/\text{m}^3$ - Micrograms per cubic meter
 TCLEE - Tetrachloroethene
 TRCLE - Trichloroethene
 Xylenes - Total xylenes

is good internal consistency in the canister data results. This can be ascertained from a cursory examination of the canister and tube data columns in Table 4.6-15. More significantly, Table 4.6-16 shows canister precision comparisons for the two collocated canister samples collected at station SQ2. Except for several outliers (benzene on August 29; toluene on August 29; and methylene chloride on September 10), canister collocated sample results are more consistent than those routinely obtained for the CM04 tube method over the past several years.

Additional refinements in the canister collection method and laboratory analyses are currently being investigated by the CAQMMP. It is possible that some VOC compounds may respond more favorably than others to the standard canister analysis procedures; again, this will be investigated. However, it is anticipated that the use of passivated canisters will be employed on a partial or full basis at RMA in FY95.

4.6.10 VOC Nontarget Analyte Results

When a sample is analyzed for VOCs with the GC/MS, results are only reported for target analytes with certified concentration ranges. Additional nontarget analyte peaks appearing on the chromatogram at greater than 10 percent of the internal standard area can be tentatively identified matching the spectra using a computer. IRDMIS refers to these tentatively identified compounds (TICs) as unknowns (UNKs) and labels them according to their relationship with the retention time of the internal standard as follows:

$$\text{UNK No.} = \frac{\text{Retention time of unknown}}{\text{Retention time of Internal Standard}} * 100$$

While the unknown numbers or units do not correlate to specific TICs, the laboratory provided a listing of most probable matches. These data are summarized in Table 4.6-17.

Table 4.6-16 Canister Precision Comparisons

Page 1 of 1

| Analyte | August 29, 1993 | | | September 4, 1993 | | | September 10, 1993 | | |
|---------|------------------------------|--------------------------------|-----------------------------------|------------------------------|--------------------------------|-----------------------------------|------------------------------|--------------------------------|-----------------------------------|
| | Primary µg/m ³ | Duplicate µg/m ³ | Relative Percent Difference | Primary µg/m ³ | Duplicate µg/m ³ | Relative Percent Difference | Primary µg/m ³ | Duplicate µg/m ³ | Relative Percent Difference |
| 111TCE | 1.145 | 1.200 | -5 | 1.691 | 1.691 | 0 | 1.145 | 1.145 | 0 |
| 112TCE | <0.545 | <0.545 | 0 | <0.545 | <0.545 | 0 | <0.545 | <0.545 | 0 |
| 111DCLE | <0.405 | <0.405 | 0 | <0.405 | <0.405 | 0 | <0.405 | <0.405 | 0 |
| 12DCLE | <0.405 | <0.405 | 0 | <0.405 | <0.405 | 0 | <0.405 | <0.405 | 0 |
| 12DCE | <0.396 | <0.396 | 0 | <0.396 | <0.396 | 0 | <0.396 | <0.396 | 0 |
| C6H6 | 1.501 | 0.607 | 85 | 3.448 | 3.608 | -5 | 0.958 | 0.926 | 3 |
| CCL4 | 0.629 | 0.629 | 0 | 0.817 | 0.817 | 0 | 0.692 | 0.755 | -9 |
| CH2CL2 | 2.083 | 1.770 | 16 | 0.937 | 1.007 | -7 | 0.486 | 0.764 | -44 |
| CHCL3 | 0.683 | <0.488 | NA | 0.635 | 0.683 | -7 | <0.488 | <0.488 | 0 |
| CLC6H5 | <0.460 | <0.460 | 0 | <0.460 | <0.460 | 0 | <0.460 | <0.460 | 0 |
| ETC6H5 | 0.478 | <0.434 | NA | 1.606 | 1.693 | -5 | <0.434 | 1.129 | NA |
| MEC6H5 | 3.878 | 1.093 | 112 | 11.823 | 10.807 | 9 | 3.088 | 2.447 | 23 |
| TCLEE | <0.678 | <0.678 | 0 | 1.085 | 1.017 | 0 | <0.678 | <0.678 | 0 |
| TRCLE | 0.591 | <0.537 | NA | 1.558 | <0.537 | NA | 0.806 | <0.537 | NA |
| XYLENES | 2.214 | <0.434 | NA | 7.815 | 8.770 | -12 | 1.997 | 0.612 | 106 |

< ## - Indicates the concentration is less than the lower certified reporting limit

NA - Not Applicable

µg/m³ - Micrograms per cubic meter

Table 4.6-17 Summary of FY93 Tentatively Identified Volatile
Organic Compound Data

Page 1 of 2

| Tentatively Identified Compound | Number of Detections |
|--|-------------------------|
| Cyclohexane, methyl- | 226 |
| Benzene, 1-ethyl-2-methyl- | 226 |
| Hexane | 119 |
| Octane | 92 |
| Benzene, 1-ethyl-4-methyl- | 92 |
| Benzene, 1-ethyl-3-methyl- | 66 |
| .alpha.-Pinene | 52 |
| Decane | 43 |
| Benzene, 1,2,4-trimethyl- | 30 |
| Pentane | 27 |
| Benzene, 1,2,3-trimethyl- | 25 |
| Cyclopentane, methyl- | 19 |
| Butylated Hydroxytoluene | 18 |
| Heptane | 17 |
| Butane, 2-methyl- | 16 |
| Benzene, 1,3,5-trimethyl- | 13 |
| Acetophenone | 10 |
| Pentane, 2-methyl- | 8 |
| Naphthalene, 2-methyl- | 8 |
| Methane, trichlorofluoro- | 8 |
| Benzene, 1,3-dimethyl- | 8 |
| Benzene, 1,2-dimethyl- | 8 |
| 1,2-Benzenedicarboxylic acid, 3-nitro- | 8 |
| Methane, dichloro- | 7 |
| 1-Hexanol, 2-ethyl- | 7 |
| Naphthalene, 1,7-dimethyl- | 6 |
| Naphthalene, 1,4-dimethyl- | 6 |
| Naphthalene, 1,2-dimethyl- | 6 |
| Heptane, 2-methyl- | 6 |
| Benzene, (1-methylethyl)- | 6 |
| 2-Propenoic acid, 3-phenyl-, ethyl ester | 6 |
| Naphthalene, 1-methyl- | 5 |
| Hexadecane | 5 |
| Dodecane, 2-methyl-6-propyl- | 5 |
| Bicyclo[3.1.1]hept-2-ene, 3,6,6-trimethyl- | 5 |
| Benzene, 2,4-diisocyanato-1-methyl- | 5 |
| 4-Carene, (1S,3S,6R)-(-)- | 5 |
| Nonane | 4 |
| Heptane, 3-methylene- | 4 |
| Ethene, 1,2-dichloro-, (Z)- | 4 |
| Dodecane | 4 |
| Cyclotrisiloxane, hexamethyl- | 4 |
| Camphene | 4 |
| Bis(2-ethylhexyl) phthalate | 4 |
| Phenothiazine | 3 |
| Octacosane | 3 |
| Naphthalene, 2,3-dimethyl- | 3 |
| Naphthalene, 1,8-dimethyl- | 3 |
| Hexanedioic acid, bis(2-ethylhexyl) ester | 3 |
| Heptadecane, 2,6,10,15-tetramethyl- | 3 |
| Ethene, 1,2-dichloro-, (E)- | 3 |
| Dibutyl phthalate | 3 |
| Decane, 2,3,5-trimethyl- | 3 |
| 1,3-Benzenediamine, 2-methyl- | 3 |
| Phenothiazine | 2 |
| Pentacosane | 2 |
| Naphthalene, 1,6-dimethyl- | 2 |
| Limonene | 2 |

Table 4.6-17 Summary of FY93 Tentatively Identified Volatile
Organic Compound Data

Page 2 of 2

| Tentatively Identified Compound | Number of Detections |
|--|-------------------------|
| Hexatriacontane | 2 |
| Hexanoic acid, 2-ethyl- | 2 |
| Hexane, 3-methyl- | 2 |
| Eicosane | 2 |
| Diethyl Phthalate | 2 |
| Benzene, 1-ethenyl-4-ethyl- | 2 |
| Benzene, 1,4-dimethyl- | 2 |
| 1,3,6-Octatriene, 3,7-dimethyl-, (E)- | 2 |
| 1,2-Benzenedicarboxylic acid, diisooctyl ester | 2 |
| .alpha.-L-Galactofuranose, 6-deoxy-1,2,3,5-tetraki | 2 |
| Triethylenediamine | 1 |
| Tridecane | 1 |
| Propanoic acid, 2-methyl-, 3-hydroxy-2,4,4-trimeth | 1 |
| Pentadecane | 1 |
| Oleic Acid | 1 |
| Octadecane, 5,14-dibutyl- | 1 |
| Nonadecane | 1 |
| Naphthalene, 2-ethenyl- | 1 |
| Naphthalene, 1,5-dimethyl- | 1 |
| Methane, dimethoxy- | 1 |
| Methane, chloro- | 1 |
| Hexane, 2,4-dimethyl- | 1 |
| Heptane, 3-methyl- | 1 |
| Heptadecane | 1 |
| Heneicosane | 1 |
| Furan, 2,5-dihydro-2,5-dimethyl- | 1 |
| Ethene, 1,1-dichloro- | 1 |
| Eicosane, 7-hexyl- | 1 |
| Docosane | 1 |
| D-Limonene | 1 |
| Cyclopentasiloxane, decamethyl- | 1 |
| Cyclododecane | 1 |
| Bis(2-methoxyethyl) phthalate | 1 |
| Bicyclo[4.2.0]octa-1,3,5-triene | 1 |
| Bicyclo[2.2.1]hept-2-ene, 1,7,7-trimethyl- | 1 |
| Benzoic acid, ammonium salt | 1 |
| Benzene, 4-pentenyl- | 1 |
| Benzene, 1-methyl-3-(1-methylethyl)- | 1 |
| Benzene, 1-methyl-2-(1-methylethyl)- | 1 |
| Benzene, 1,4-dichloro- | 1 |
| Benzene, 1,3-dimethyl- | 1 |
| Benzene, 1,3-dichloro- | 1 |
| Benzene, 1,3,5-tris(3-methyl-3-butenyl)- | 1 |
| Benzene, 1,2-dichloro- | 1 |
| Benzene, (2-methyl-1-butenyl)- | 1 |
| 4-Carene, (1S,3R,6R)-(-)- | 1 |
| 2-Undecanone, 6,10-dimethyl- | 1 |
| 2-Propenoic acid, 3-(4-methoxyphenyl)-, 2-ethylhex | 1 |
| 2-Pentadecanone, 6,10,14-trimethyl- | 1 |
| 2,5-Cyclohexadiene-1,4-dione, 2,6-bis(1,1-dimethyl | 1 |
| 1-Octadecanol | 1 |
| 1,2-Pentadiene | 1 |
| 1,2-Benzenedicarboxylic acid, bis(2-methoxyethyl) | 1 |
| 1,1'-Biphenyl | 1 |

4.7 SEMIVOLATILE ORGANIC COMPOUNDS AND ORGANOCHLORINE PESTICIDES

4.7.1 Monitoring, Analysis and Reporting Strategies

SVOC and OCP monitoring, which provide measurements of pesticide levels at the RMA interior and perimeter sites, are described in this section. During FY93, routine OCP monitoring was employed to measure pesticide levels at RMA. In FY91, OCP monitoring replaced SVOC monitoring as the principal method for pesticide monitoring. The change was the result of experience and analysis of several different methods to determine the most effective means of collecting and measuring pesticides. The F-7 GC/MS method, an MRI certified method, was initially used during Basin F remediation for analyses of SVOCs, and the H-7 GC/ECD method was used for the analyses of OCPs. During the FY88 Basin F remedial period, the H-7 GC/ECD method was also used to confirm low levels of SVOC pesticides. The H-7 method was approximately 100 times more sensitive to pesticides than the F-7 method, and in FY89 and FY90, there were no detections under the F-7 method; consequently, the more sensitive H-7 method identified pesticides for both of these periods for the Basin F and IRA-F monitoring program.

In FY90, when the ESE laboratory was selected for CMP analyses, method CM03 (GC/MS) was certified for SVOC analyses, and method CH01 (GC/ECD) for OCP analyses. These methods, however, displayed the same characteristics as before. OCPs were readily detected, while SVOCs were generally below detection. In FY93, the OCP method was used exclusively for routine monitoring; the SVOC method was used for high-event monitoring to present a broader suite of potential compounds and to ensure that SVOCs did not occur at levels of concern (as they did on occasion during Basin F remediation).

Routine OCP monitoring consisted of weekly sampling at 12 sites (including collocated stations) and monthly sampling at two sites. SVOCs were monitored monthly at two sites, seasonally at 8 sites, and for 12 high-event episodes. Table 3.2-2 provides details of the sampling schedule. A summary of OCP monitoring, including sampling sites, number of samples, and recoveries, is provided in Table 4.7-1. The high-event schedule (for SVOCs) is provided in Table 4.7-2.

Table 4.7-1 Summary of Organochlorine Pesticide Monitoring for FY93

Page 1 of 1

| Station | Samples Scheduled | Samples Collected | Valid Samples | Field % Recovery | Overall % Recovery |
|-----------|----------------------|----------------------|------------------|---------------------|-----------------------|
| AQ2 | 61 | 58 | 57 | 95 | 93 |
| AQ3 | 61 | 57 | 56 | 93 | 92 |
| AQ5 | 61 | 60 | 57 | 98 | 93 |
| AQ6 | 61 | 56 | 55 | 92 | 90 |
| AQ8 | 61 | 57 | 57 | 93 | 93 |
| AQ9 | 61 | 55 | 55 | 90 | 90 |
| FC1 | 61 | 56 | 56 | 92 | 92 |
| FC2 | 61 | 58 | 57 | 95 | 93 |
| FC3 | 12 | 11 | 11 | 92 | 92 |
| FC4 | 12 | 12 | 12 | 100 | 100 |
| FC5 | 61 | 57 | 57 | 93 | 93 |
| SQ1 | 61 | 55 | 55 | 90 | 90 |
| SQ2 | 61 | 56 | 56 | 92 | 92 |
| Duplicate | 61 | 57 | 57 | 93 | 93 |
| Overall | 756 | 705 | 698 | 93 | 92 |

Table 4.7-2 FY93 Semivolatile Organic Compound High-Event Sampling Locations

| Date | Target Area | Duration (Hours) | Sampling Locations |
|-----------|--|--------------------------|---|
| 12/11/92* | Basin F | 1100 - 1100 | AQ2, AQ3, AQ5, AQ6, FC1, FC2 |
| 1/14/93 | Seasonal | 1000 - 1000 | AQ1, AQ2, AQ3, AQ4, AQ5, AQ6, AQ8 |
| 1/22/93* | Basin F | 1300 - 1300 | AQ2, AQ3, AQ5, AQ6, FC2, FC4 |
| 2/24/93 | Seasonal | 1200 - 1200 | AQ1, AQ2, AQ3, AQ4, AQ5, AQ6, AQ8 |
| 3/4/93 | Submerged Quench Incinerator (baseline) | 1200 - 1200 | AQ2, AQ3, AQ5, AQ6, SQ1, SQ2, FC2, M1E |
| 4/10/93 | Basin A | 1500 - 1500 | AQ2, AQ3, AQ5, AQ6, M136E, M236W |
| 5/13/93 | Hydrazine Plant | 1100 - 1100 | AQ2, AQ5, AQ6, AQ9, M101E, M201W |
| 6/3/94 | Seasonal | 1100 - 1100 | AQ1, AQ2, AQ3, AQ4, AQ5, AQ6, AQ8 |
| 6/12/94 | Submerged Quench Incinerator (trial burn) | 0700 - 0700 | AQ3, AQ6, SQ1, SQ2, FC2, M125W |
| 6/24/93 | Basin A Excavation Activities | 1300 - 1300 | FC1, AQ8, AQ9, M336E, M436W |
| 6/30/93 | Basin F | 1100 - 1100 | FC3, FC4, AQ11, M226E |
| 7/15/93 | Basin A Soil Sifting | 0700 - 1500 (8 hours) | AQ1, AQ4, AQ5, AQ8, AQ9, AQ10, M601N, M701N |
| 7/30/93 | Section 27 Water Treatment Plant | 1000 - 1000 | AQ1 AQ4, AQ5, AQ8, AQ9, AQ10, M601N, M701 |
| 8/11/93 | South Plants Area | 1200 - 1200 | AQ5, AQ6, AQ8, AQ9, M801N, M102E |
| 8/29/93 | Submerged Quench Incinerator | 1300 - 1300 | AQ3, AQ6, FC2, SQ1, SQ2, M225W |
| 9/12/93 | Basin F Waste Pile | 1200 - 1200 | FC2, FC3, FC4, M326E, M426S, M526SE |
| 9/28/93 | Pond A Liner Cleaning | 1200 - 1200 | AQ2, AQ11, AQ12, FC1, FC4, FC5 |

* Laboratory performed OCP analysis instead of SVOC analysis
 OCP Organochlorine Pesticides
 SVOC Semivolatile Organic Compounds

RMA/0984 10/19/94 3:52 pm ap

4.7.2 The CAQMMP OCP Monitoring Results

Tables 4.7-3 and 4.7-4 present summaries of FY93 OCP monitoring results for all CAQMMP sampling sites at RMA. Table 4.7-3 shows the annual average, and Table 4.7-4 presents the 24-hour maximum concentrations of each OCP at each monitoring site. As in the case of VOCs, those OCPs below the CRL were assigned a value of one-half the detection limit for averaging purposes (shown in parentheses in Table 4.7-3).

Measured concentrations during FY93 at Basin F and at RMA perimeter sites were generally comparable to levels observed during the Basin F post-remedial period. The one exception was the AQ8 site, located just to the south of Basin A and north of the South Plants. Various construction and remedial activities were accomplished in the Basin A area during FY93, and several pesticides were measured at higher levels than in the recent past. These results will be discussed in further detail below. It is interesting that several of the OCPs measured at AQ8 were an order of magnitude higher than post-remedial OCPs at Basin F, but were nevertheless well below peak levels observed during the active Basin F remediation period.

The highest maximum 24-hour concentration of aldrin during FY93, $0.0023 \mu\text{g}/\text{m}^3$, occurred at FC2D. This was well below the FY92 aldrin 24-hour maximum value of $0.0065 \mu\text{g}/\text{m}^3$ (at FC2) and significantly below the peak value of $0.8366 \mu\text{g}/\text{m}^3$ during active Basin F remediation. The FY93 annual average aldrin level was $0.00023 \mu\text{g}/\text{m}^3$ at FC2, compared to $0.00021 \mu\text{g}/\text{m}^3$ in FY92. Aldrin levels at the Basin A site, AQ8 were slightly higher than the recent past with a 24-hour maximum level of $0.0018 \mu\text{g}/\text{m}^3$, and an annual average of $0.0003 \mu\text{g}/\text{m}^3$, which was the highest annual average of any of the sites.

The 24-hour maximum level of chlordane was $0.0360 \mu\text{g}/\text{m}^3$ at AQ8; the second highest 24-hour level was $0.0150 \mu\text{g}/\text{m}^3$ at AQ9 at the northern extremity of Section 36, northeast of Basin A. Chlordane appeared to be confined to the Basin A area. The maximum 24-hour chlordane concentration at Basin F was $0.0021 \mu\text{g}/\text{m}^3$ at FC2. Chlordane did not appear to be a

Table 4.7-3 Summary of FY93 Average Organochlorine Pesticide Concentrations (µg/m³) Page 1 of 1

| Station | Aldrin | Chlordane | Dieldrin | Endrin | Isodrin | PPDDE | PPDDT |
|---------|----------------------|----------------------|----------------------|----------------------|----------------------|----------------------|----------------------|
| AQ2 | <0.0003 (0.00015) | <0.0003 (0.00019) | <0.0003 (0.00019) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| AQ3 | <0.0003 (0.00015) | <0.0003 (0.00021) | <0.0003 (0.00021) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| AQ5 | <0.0003 (0.00015) | <0.0003 (0.00019) | <0.0003 (0.00018) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| AQ6 | <0.0003 (0.00017) | 0.0006 (0.00056) | 0.0007 (0.00069) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| AQ8 | 0.0003 | 0.0041 | 0.0046 | 0.0008 | <0.0003 (0.00019) | <0.0003 (0.00016) | 0.0003 |
| AQ9 | <0.0003 (0.00018) | 0.0014 | 0.0008 | <0.0003 (0.00018) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| FC1 | <0.0003 (0.00018) | 0.0003 | 0.0007 | <0.0003 (0.00016) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| FC2 | <0.0003 (0.00023) | 0.0003 | 0.0020 | <0.0003 (0.00024) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| FC3 | <0.0003 (0.00021) | 0.0004 | 0.0021 | <0.0003 (0.00020) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| FC4 | <0.0003 (0.00015) | <0.0003 (0.00020) | 0.0005 | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| FC5 | <0.0003 (0.00018) | 0.0003 | 0.0008 | <0.0003 (0.00016) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| SQ1 | <0.0003 (0.00015) | 0.0003 | 0.0006 | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |
| SQ2 | <0.0003 (0.00016) | 0.0004 | 0.0009 | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) | <0.0003 (0.00015) |

Note: < ### indicates average concentration was less than the lower certified reporting limit
 () indicates computed average, when [average] is less than the lower certified reporting limit.
 µg/m³ Micrograms per cubic meter

RMA/0985 10/11/94 10:13 am bpw

Table 4.7-4 Summary of FY93 24-Hour Maximum Organochlorine Pesticide Concentrations (µg/m³)

Page 1 of 1

| Station | Aldrin | Chlordane | Dieldrin | Endrin | Isodrin | PPDDE | PPDDT |
|---------|---------|-----------|----------|---------|---------|---------|---------|
| AQ2 | <0.0003 | 0.0013 | 0.0010 | <0.0003 | <0.0003 | <0.0003 | <0.0003 |
| AQ3 | <0.0003 | 0.0012 | 0.0012 | <0.0003 | <0.0003 | <0.0003 | 0.0005 |
| AQ5 | <0.0003 | 0.0007 | 0.0007 | <0.0003 | <0.0003 | <0.0003 | <0.0003 |
| AQ6 | 0.0009 | 0.0038 | 0.0048 | <0.0003 | <0.0003 | 0.0006 | <0.0003 |
| AQ8 | 0.0018 | 0.0360 | 0.0430 | 0.0052 | 0.0012 | <0.0003 | 0.0024 |
| AQ9 | 0.0013 | 0.0150 | 0.0068 | 0.0008 | <0.0003 | <0.0003 | <0.0003 |
| FC1 | 0.0017 | 0.0016 | 0.0068 | 0.0004 | <0.0003 | <0.0003 | <0.0003 |
| FC2 | 0.0023 | 0.0021 | 0.0190 | 0.0013 | <0.0003 | <0.0003 | 0.0004 |
| FC3 | 0.0006 | 0.0016 | 0.0120 | 0.0007 | <0.0003 | <0.0003 | <0.0003 |
| FC4 | <0.0003 | 0.0007 | 0.0015 | <0.0003 | <0.0003 | <0.0003 | <0.0003 |
| FC5 | 0.0015 | 0.0017 | 0.0059 | 0.0004 | <0.0003 | <0.0003 | <0.0003 |
| SQ1 | <0.0003 | 0.0017 | 0.0051 | 0.0004 | <0.0003 | <0.0003 | <0.0003 |
| SQ2 | 0.0006 | 0.0029 | 0.0070 | <0.0003 | <0.0003 | <0.0003 | <0.0003 |

Note: < ### indicates concentration was less than the lower certified reporting limit
µg/m³ Micrograms per cubic meter

contaminant of concern at Basin F during the remediation activities. The maximum annual chlordane average was $0.00407 \mu\text{g}/\text{m}^3$ at AQ8; the second maximum average was $0.00143 \mu\text{g}/\text{m}^3$ at AQ9. Other chlordane levels were quite low and primarily below detection.

Dieldrin was again highest at AQ8 where the 24-hour maximum concentration was $0.0430 \mu\text{g}/\text{m}^3$ and the annual average was $0.00451 \mu\text{g}/\text{m}^3$. Within the Basin F area, dieldrin reached a maximum 24-hour level of $0.0190 \mu\text{g}/\text{m}^3$ at FC2 and a maximum average of $0.00214 \mu\text{g}/\text{m}^3$. These concentrations were below FY92 levels, and considerably below active remediation levels when station BF2 reported a maximum 24-hour dieldrin level of $2.260 \mu\text{g}/\text{m}^3$.

Endrin was also highest at AQ8 during FY93, with a maximum 24-hour concentration of $0.0052 \mu\text{g}/\text{m}^3$ and an annual average of $0.00082 \mu\text{g}/\text{m}^3$. The next highest levels were at FC2 with a 24-hour maximum value of $0.0013 \mu\text{g}/\text{m}^3$ and an annual average of $0.00024 \mu\text{g}/\text{m}^3$. Again they were comparable to Basin F post-remediated levels and well below Basin F remedial levels when a maximum endrin concentration of $1.0954 \mu\text{g}/\text{m}^3$ was measured at station BF2C.

Isodrin was measured at the 24-hour maximum concentration of $0.0012 \mu\text{g}/\text{m}^3$ and a maximum annual concentration of $0.00019 \mu\text{g}/\text{m}^3$, both at AQ8. All other 24-hour concentrations of isodrin were below detection. During Basin F remediation, isodrin occasionally reached high levels with a maximum 24-hour concentration of $0.9450 \mu\text{g}/\text{m}^3$ at station BF2C. Isodrin has been measured infrequently during the post-remedial period.

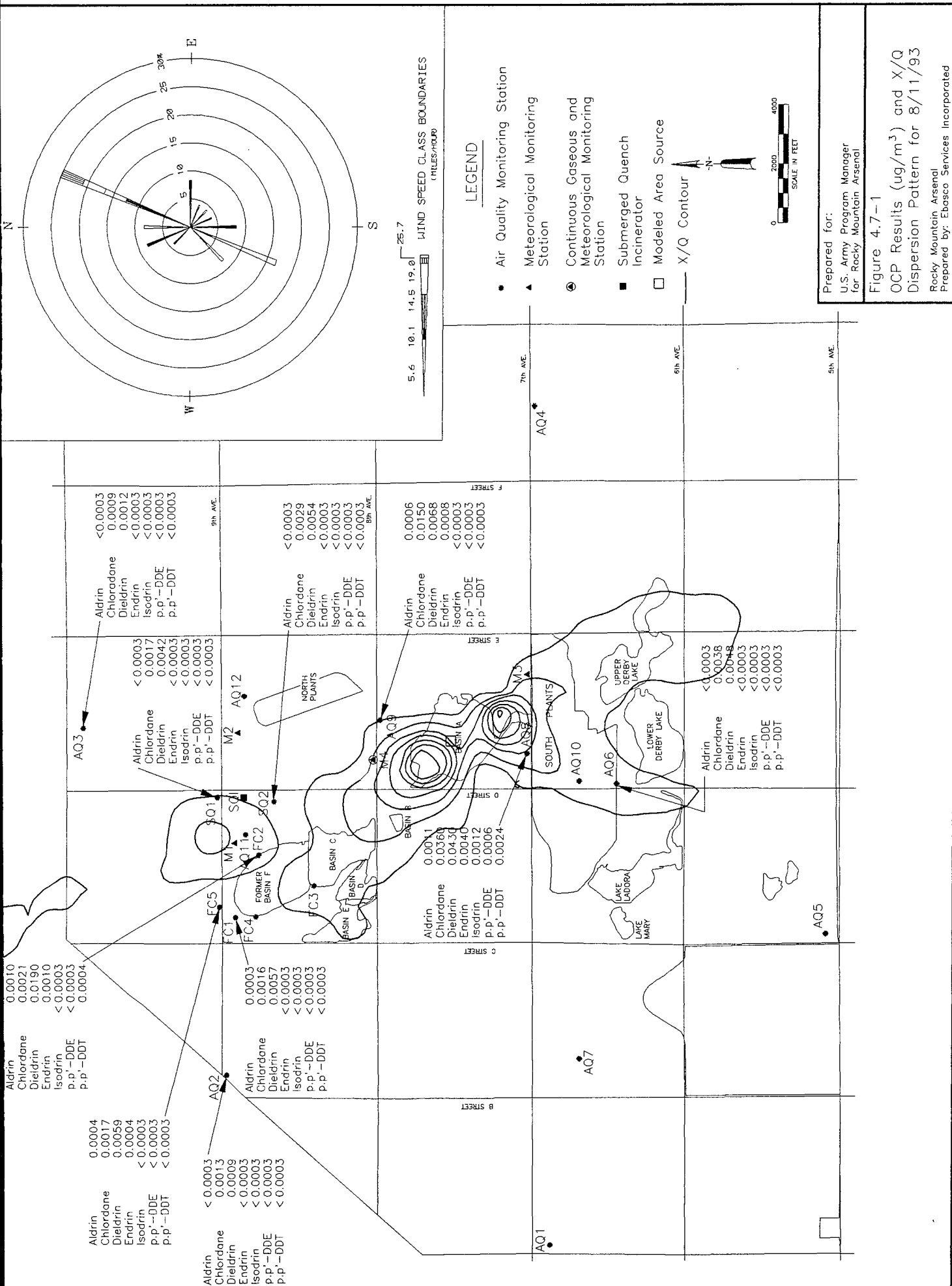
Both PPDE and PPDDT were measured infrequently during FY93. The highest 24-hour maximum values were $0.0024 \mu\text{g}/\text{m}^3$ for PPDDT, and $0.0006 \mu\text{g}/\text{m}^3$ for PPDE, both at AQ8. These concentrations appear to have been related to construction activities in the Basin A area and on the pipeline leading from Basin F to the SQI. The annual maximum averages at AQ8 were $0.0003 \mu\text{g}/\text{m}^3$ for PPDDT and $0.0002 \mu\text{g}/\text{m}^3$ for PPDE. Almost all measurements at all sites (including AQ8) were below detection. These compounds did not appear to have been a

Basin F remediation concern but were occasionally measured at low levels during the Basin F remedial period.

4.7.3 High-Event and Special Day Monitoring

In accordance with the CAQMMP Technical Plan, high-event monitoring for SVOCs was conducted on 12 days in FY93 at locations and dates shown in Table 4.7-2. As previously noted, the detection level for SVOCs is considerably higher than for OCPs. Nevertheless the target list for SVOCs contains some 22 potential contaminants and an effort was made to screen these analytes for possible impacts from RMA activities. No SVOCs were detected, however, except for one measurement of dieldrin at $0.0540 \mu\text{g}/\text{m}^3$ at a manhole cover near Building 511 on September 12, 1993. This level was slightly above the maximum dieldrin concentration measured at AQ8 using the OCP method. It is anticipated that future pesticide high events will be conducted for OCPs with additional SVOC screening during emergency events. It is important also to point out that in FY93, OCP monitoring was conducted as a backup on many of the SVOC high-event monitoring days.

It is of special interest that on those days identified with high VOC concentrations, especially the December 8, 1992 inversion episode when there was a brown cloud intrusion over RMA, OCP concentrations remained at low levels. This suggests that pesticides are confined to special areas of RMA where they were deposited during previous industrial or remediation activities. It appears that Basin A is one such area. During OCP monitoring on August 11, 1993, five pesticides were measured at their highest levels of the year and peak concentrations were centered around Basin A as illustrated in Figure 4.7-1. Ambient OCP concentrations and the X/Q dispersion pattern (with a potential source over Basin A) are shown in this figure. On this day, winds were primarily from the north-northeast, however, all directions were represented; wind speeds were strong and gusty at the beginning of the monitoring period, becoming light to moderate the remainder of the period. Temperatures reached the low 80s. Atmospheric stability was typical of a summer day—stable during the evening (classes 5 and 6), unstable at midday (classes 1 and 2), and neutral (class 4) during several strong wind speed periods. The dispersion



pattern shows maximum X/Q values centered over the Basin A potential source area and extending northwest and southeast; it is interesting that this pattern reflects the hours of low wind speeds and stable atmospheric stability rather than prevailing wind flow under stronger wind speeds. However, the dispersion pattern corresponds closely to actual OCP concentrations with the highest levels of all pesticides at AQ8 and decreasing proportional northwest and southeast. As Basin F is also a potential source of several of these compounds, it is not surprising that moderate levels were also measured at the Basin F monitoring sites. However, perimeter sites AQ2 and AQ3, farther downwind and outside the principal dispersion patterns, measured very low or non-detectable concentrations. The August 11 data are typical of other FY93 data during Basin A intrusive activity and indicate the sensitivity of the OCP method to detect potential impacts and also the capability of available modeling tools to predict potential impacts.

4.7.4 Basin F Pesticide Impacts

As previously discussed, a comprehensive evaluation of the ambient impacts from Basin F cleanup operations and other remedial activities is performed by considering all CAQMMP and Basin F Remedial Monitoring Program data collected during the remedial and post-remedial periods. Some SVOC/OCP data have been presented in previous sections; however, this section will provide a more comprehensive comparative analysis.

Tables 4.7-5 and 4.7-6 provide a comparison between Basin F SVOC/OCP data during remediation (Phase 1) and the most recent post-remediation period (Phase 7). Intermediate phases are provided in previous CMP/CAQMMP Annual Reports. Several perimeter sites and interior site AQ8 are also included in this comparison. It is apparent that measurements of pesticides in the remediated basin area have been significantly reduced during this post-remedial period (by several orders of magnitude); also see Figure 4.7-2, which shows bar graphs for several pesticides over the remedial and post-remedial periods. Perimeter site measurements have remained near constant levels, indicating that the contaminants are confined to selected interior locations (i.e., Basin F, Basin A). Of interest is the increase in pesticide concentrations at AQ8 during FY93 intrusive activities. Nevertheless, these levels are well below the high

Table 4.7-5 Summary of Average Organochlorine Pesticide Results for Phases 1, 2 Stage 1, and 7 ($\mu\text{g}/\text{m}^3$) Page 1 of 2

| Site | Phases | Aldrin | Chlordane | Dieldrin | Endrin | Isodrin | PPDDE | PPDDT |
|---------|-----------------|--------|-----------|----------|--------|---------|--------|--------|
| AQ3 | Phase 1 | 0.0053 | 0.0003 | 0.0022 | 0.0013 | 0.0005 | ND | 0.0004 |
| | Phase 2 Stage 1 | ND | 0.0003 | 0.0004 | 0.0003 | ND | ND | 0.0003 |
| | Phase 7 | ND | 0.0002 | 0.0002 | ND | ND | ND | 0.0002 |
| AQ5 | Phase 1 | 0.0013 | 0.0003 | 0.0013 | 0.0009 | 0.0003 | 0.0004 | 0.0006 |
| | Phase 2 Stage 1 | ND | 0.0003 | 0.0003 | ND | ND | ND | ND |
| | Phase 7 | ND | 0.0002 | 0.0002 | ND | ND | ND | ND |
| AQ8 | Phase 1 | ND | ND | ND | ND | ND | 0.0390 | ND |
| | Phase 2 Stage 1 | * | * | * | * | * | * | * |
| | Phase 7 | 0.0003 | 0.0041 | 0.0046 | 0.0008 | 0.0002 | 0.0002 | 0.0003 |
| BF1/FC1 | Phase 1 | 0.1913 | ND | 0.1408 | 0.0197 | 0.0076 | ND | ND |
| | Phase 2 Stage 1 | 0.0106 | ND | 0.0075 | 0.0041 | 0.0013 | ND | ND |
| | Phase 7 | 0.0002 | 0.0003 | 0.0007 | 0.0002 | ND | ND | ND |
| BF2/FC2 | Phase 1 | 0.4579 | ND | 0.2768 | 0.1209 | 0.0424 | ND | ND |
| | Phase 2 Stage 1 | 0.0154 | ND | 0.0107 | 0.0044 | 0.0010 | ND | ND |
| | Phase 7 | 0.0002 | 0.0003 | 0.0020 | 0.0002 | ND | ND | 0.0002 |
| BF3/FC3 | Phase 1 | 0.0662 | ND | 0.0808 | 0.0250 | 0.0030 | ND | ND |
| | Phase 2 Stage 1 | 0.0026 | ND | 0.0032 | 0.0017 | 0.0007 | ND | ND |
| | Phase 7 | 0.0002 | 0.0004 | 0.0021 | 0.0002 | ND | ND | ND |
| BF4/FC4 | Phase 1 | 0.0499 | ND | 0.0477 | 0.0234 | 0.0030 | ND | ND |
| | Phase 2 Stage 1 | 0.0034 | ND | 0.0027 | 0.0014 | 0.0007 | ND | ND |
| | Phase 7 | ND | 0.0002 | 0.0005 | 0.0002 | ND | ND | ND |
| BF5 | Phase 1 | 0.0221 | ND | 0.0193 | 0.0096 | 0.0011 | ND | ND |
| | Phase 2 Stage 1 | 0.0010 | ND | 0.0010 | 0.0008 | ND | ND | ND |
| | Phase 7 | * | * | * | * | * | * | * |

Table 4.7-5 Summary of Average Organochlorine Pesticide Results for Phases 1, 2 Stage 1, and 7 ($\mu\text{g}/\text{m}^3$) Page 2 of 2

| Site | Phases | Aldrin | Chlordane | Dieldrin | Endrin | Isodrin | PPDDE | PPDDT |
|------|-----------------|--------|-----------|----------|--------|---------|-------|-------|
| FC5 | Phase 1 | * | * | * | * | * | * | * |
| | Phase 2 Stage 1 | * | * | * | * | * | * | * |
| | Phase 7 | 0.0002 | 0.0003 | 0.0008 | 0.0002 | ND | ND | ND |

* Samples not collected at this site.

ND Not detected.

$\mu\text{g}/\text{m}^3$ Micrograms per cubic meter

Note: Phase 1 is from March 22 to December 12, 1988.

Phase 2, Stage 1 is from December 13, 1988 to February 15, 1989.

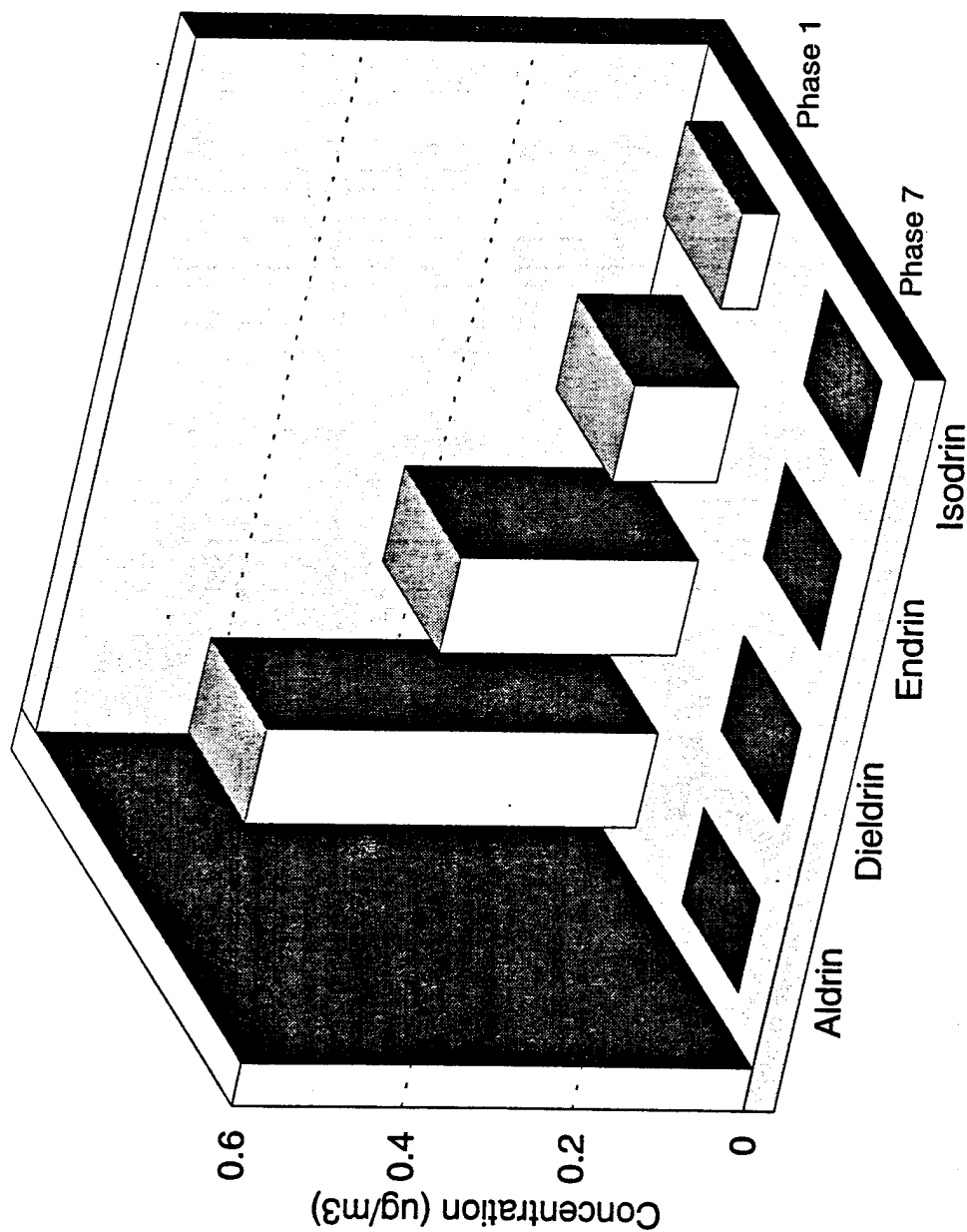
Phase 7 is from October 1, 1992 to September 30, 1993.

Table 4.7-6 Summary of 24-Hour Maximum Organochlorine Pesticide Results for Phases 1, 2 Stage 1, and 7 (µg/m³) Page 1 of 2

| Site | Phases | Aldrin | Chlordane | Dieldrin | Endrin | Isodrin | PPDDE | PPDDT |
|---------|-----------------|--------|-----------|----------|--------|---------|--------|--------|
| AQ3 | Phase 1 | 0.0290 | 0.0005 | 0.0061 | 0.0041 | 0.0024 | ND | 0.0012 |
| | Phase 2 Stage 1 | ND | 0.0004 | 0.0009 | 0.0004 | ND | ND | 0.0006 |
| | Phase 7 | ND | 0.0012 | 0.0012 | ND | ND | ND | 0.0005 |
| AQ5 | Phase 1 | 0.0095 | 0.0008 | 0.0074 | 0.0046 | 0.0008 | 0.0023 | 0.0050 |
| | Phase 2 Stage 1 | ND | 0.0004 | 0.0006 | ND | ND | ND | ND |
| | Phase 7 | ND | 0.0007 | 0.0007 | ND | ND | ND | ND |
| AQ8 | Phase 1 | ND | ND | ND | ND | ND | 0.0390 | ND |
| | Phase 2 Stage 1 | * | * | * | * | * | * | * |
| | Phase 7 | 0.0018 | 0.0360 | 0.0430 | 0.0052 | 0.0012 | 0.0006 | 0.0024 |
| BF1/FC1 | Phase 1 | 0.8366 | ND | 0.4904 | 0.2497 | 0.0604 | ND | ND |
| | Phase 2 Stage 1 | 0.0830 | ND | 0.0361 | 0.0223 | 0.0114 | ND | ND |
| | Phase 7 | 0.0017 | 0.0016 | 0.0068 | 0.0004 | ND | ND | ND |
| BF2/FC2 | Phase 1 | 2.8290 | ND | 2.2960 | 0.9020 | 0.8160 | ND | ND |
| | Phase 2 Stage 1 | 0.1244 | ND | 0.0355 | 0.0147 | 0.0100 | ND | ND |
| | Phase 7 | 0.0023 | 0.0021 | 0.0190 | 0.0013 | ND | ND | 0.0004 |
| BF3/FC3 | Phase 1 | 0.4288 | ND | 1.6568 | 0.5458 | 0.0226 | ND | ND |
| | Phase 2 Stage 1 | 0.0234 | ND | 0.0130 | 0.0061 | 0.0034 | ND | ND |
| | Phase 7 | 0.0006 | 0.0016 | 0.0120 | 0.0007 | ND | ND | ND |
| BF4/FC4 | Phase 1 | 0.3478 | ND | 0.3265 | 0.2591 | 0.1136 | ND | ND |
| | Phase 2 Stage 1 | 0.0321 | ND | 0.0099 | 0.0042 | 0.0078 | ND | ND |
| | Phase 7 | ND | 0.0007 | 0.0015 | ND | ND | ND | DN |
| BF5 | Phase 1 | 0.1723 | ND | 0.0499 | 0.0275 | 0.0080 | ND | ND |
| | Phase 2 Stage 1 | 0.0080 | ND | 0.0030 | 0.0010 | ND | ND | ND |
| | Phase 7 | * | * | * | * | * | * | * |

Table 4.7-6 Summary of 24-Hour Maximum Organochlorine Pesticide Results for Phases 1, 2 Stage 1, and 7 (µg/m³) Page 1 of 2

| Site | Phases | Aldrin | Chlordane | Dieldrin | Endrin | Isodrin | PPDDE | PPDDT |
|---------|-----------------|--------|-----------|----------|--------|---------|--------|--------|
| AQ3 | Phase 1 | 0.0290 | 0.0005 | 0.0061 | 0.0041 | 0.0024 | ND | 0.0012 |
| | Phase 2 Stage 1 | ND | 0.0004 | 0.0009 | 0.0004 | ND | ND | 0.0006 |
| | Phase 7 | ND | 0.0012 | 0.0012 | ND | ND | ND | 0.0005 |
| AQ5 | Phase 1 | 0.0095 | 0.0008 | 0.0074 | 0.0046 | 0.0008 | 0.0023 | 0.0050 |
| | Phase 2 Stage 1 | ND | 0.0004 | 0.0006 | ND | ND | ND | ND |
| | Phase 7 | ND | 0.0007 | 0.0007 | ND | ND | ND | ND |
| AQ8 | Phase 1 | ND | ND | ND | ND | ND | 0.0390 | ND |
| | Phase 2 Stage 1 | * | * | * | * | * | * | * |
| | Phase 7 | 0.0018 | 0.0360 | 0.0430 | 0.0052 | 0.0012 | 0.0006 | 0.0024 |
| BF1/FC1 | Phase 1 | 0.8366 | ND | 0.4904 | 0.2497 | 0.0604 | ND | ND |
| | Phase 2 Stage 1 | 0.0830 | ND | 0.0361 | 0.0223 | 0.0114 | ND | ND |
| | Phase 7 | 0.0017 | 0.0016 | 0.0068 | 0.0004 | ND | ND | ND |
| BF2/FC2 | Phase 1 | 2.8290 | ND | 2.2960 | 0.9020 | 0.8160 | ND | ND |
| | Phase 2 Stage 1 | 0.1244 | ND | 0.0355 | 0.0147 | 0.0100 | ND | ND |
| | Phase 7 | 0.0023 | 0.0021 | 0.0190 | 0.0013 | ND | ND | 0.0004 |
| BF3/FC3 | Phase 1 | 0.4288 | ND | 1.6568 | 0.5458 | 0.0226 | ND | ND |
| | Phase 2 Stage 1 | 0.0234 | ND | 0.0130 | 0.0061 | 0.0034 | ND | ND |
| | Phase 7 | 0.0006 | 0.0016 | 0.0120 | 0.0007 | ND | ND | ND |
| BF4/FC4 | Phase 1 | 0.3478 | ND | 0.3265 | 0.2591 | 0.1136 | ND | ND |
| | Phase 2 Stage 1 | 0.0321 | ND | 0.0099 | 0.0042 | 0.0078 | ND | ND |
| | Phase 7 | ND | 0.0007 | 0.0015 | ND | ND | ND | DN |
| BF5 | Phase 1 | 0.1723 | ND | 0.0499 | 0.0275 | 0.0080 | ND | ND |
| | Phase 2 Stage 1 | 0.0080 | ND | 0.0030 | 0.0010 | ND | ND | ND |
| | Phase 7 | * | * | * | * | * | * | * |



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.7-2

Average OCP Concentrations
at FC2

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

concentrations achieved during active Basin F remediation. One exception is that higher concentrations of chlordane have been measured during Phase 7 than in previous post-remedial phases, indicating the potential for pesticide releases during continuing remediation activities at certain specified areas.

4.7.5 Summary of SVOC/OCP Results

Table 4.7-7 shows the locations of both maximum long-term (annual) average and maximum short-term (24-hour) SVOC/OCP concentrations for the combined CMP/CAQMMP and Basin F data analyses results for the phases of remediation. It is noted that almost all of the highest concentrations occurred in the initial remediation phase; however, the 24-hour maximum concentration of chlordane at AQ8 in FY93 (Phase 7) matched the highest 24-hour level in Phase 2 (Step 1) at BF2. Also, highest concentrations occurred most often at station BF2, directly downwind from Basin F.

As in the case of VOCs, the CAQMMP awaits the promulgation of CDH air toxic standards in order to provide a more definitive evaluation of RMA OCP/SVOC monitored data. In summary, Basin F appeared to be a potential source of several SVOC and OCP compounds measured under the CAQMMP and Basin F monitoring programs. In particular, these included aldrin, endrin, and dieldrin. Highest levels were clustered around Basin F. Monitoring station BF2 on the northeast perimeter of the basin measured the highest concentrations for most compounds. Pesticide levels at all RMA sites, and in particular Basin F sites, decreased significantly during the Basin F post-remedial period including FY93, although low levels of some pesticide compounds continued to be measured.

4.8 AIR STRIPPER OFF-GAS TESTING

This section provides the results of the CAQMMP air stripper off-gas testing at the A-Neck groundwater treatment facility. The air stripper system is located in the south room of the A-Neck groundwater treatment building. Water supplied to the air stripper for treatment is obtained from a groundwater extraction well located near the Basin F area. Off-gas samples were obtained

Table 4.7-7 Maximum Average Long-term and Short-term Semivolatile Organic Compound Concentrations

Page 1 of 1

| SVOC | Maximum Long-term* Average ($\mu\text{g}/\text{m}^3$) | Location | Phase | Maximum Short-term Concentration ($\mu\text{g}/\text{m}^3$) | Location | Phase |
|-----------|---|----------|-------|---|------------|-------------|
| Aldrin | 0.5263 | BF2 | P1 | 2.8290 | BF2 | P1 |
| Chlordane | 0.0071 | BF2 | P2-S1 | 0.0360 0.0360 | BF2 AQ8 | P2-S1 P7 |
| Dieldrin | 0.3971 | CMP/BF2 | P1 | 2.2960 | BF2 | P1 |
| Endrin | 0.1343 | CMP/BF2 | P1 | 1.0954 | BF2 | P1 |
| Isodrin | 0.1122 | CMP/BF2 | P1 | 0.9450 | BF2 | P1 |
| PPDDE | 0.0390 | AQ8 | P1 | 0.0390 | BF2 | P1 |
| PPDDT | 0.0350 | CMP/BF2 | P1 | 0.0585 | CMP/BF2 | P1 |
| Atrazine | ND | NA | NA | ND | NA | NA |
| Malathion | ND | NA | NA | ND | NA | NA |
| Parathion | 0.1311 | CMP/BF2 | P1 | 0.2099 | CMP/BF2 | P1 |
| Supona | ND | NA | NA | ND | NA | NA |

* Several maximum long-term averages provided in this table represent limited sampling from the high event programs.
 $\mu\text{g}/\text{m}^3$ Micrograms per cubic meter

Legend: PPDDE = Dichlorodiphenylethane
 PPDDT = Dichlorodiphenyltrichloroethane
 P1 = Phase 1
 P2-S1 = Phase 2-Stage 1
 P7 = Phase 7
 NA = Not applicable
 ND = Not detected

at three locations: upstream, prior to carbon adsorption; midstream, after treatment in the first carbon unit; and downstream, following the second carbon unit and exhaust blower. The carbon adsorption units are filled with granular activated carbon and are designed for vapor phase contaminant reduction in air flow streams. Sorbent tubes and Tedlar bags were used for sample gas collection at the three off-gas sample ports. Background concentrations were determined by collection of ambient air in Tedlar bags and sorbent tubes at a position outside the building. All samples collected in Tedlar bags were immediately analyzed on site using OVA and OVM direct-reading instruments. Sorbent tube samples were shipped to the laboratory for analysis.

In each sampling event, sorbent tube samples were aerated with air stripper off gas for approximately 80 minutes. A volume of 10 liters of off gas was typically sampled using the tube sets. The samples obtained in Tedlar bags were collected in a period of approximately 3 minutes. A series of three bag samples was collected during the tube aeration periods. During weekly events, the normal testing routine entailed the collection of bag samples at the three ports and exterior position, plus sorbent tube samples at the midstream port (port B) using front and middle Tenax tubes followed by an XAD (adsorbent resin for SVOC sampling) tube in series. During monthly events, the normal testing routine entailed collection of bag samples at the same locations as in weekly events plus sorbent tube samples outside the air stripper building and at the upstream (port A), midstream (port B), and downstream (port C) ports using Tenax, Tenax back-up, and XAD tubes in series. Monthly events also included a collocated tube sample set at the midstream port and a set of Tenax and XAD field blanks.

4.8.1 Real-time Sampling Results

A synopsis of sampling efforts is shown in Table 4.8-1, including the calendar week, test date, and type of test performed. A listing of the real-time OVA and OVM readings for each event is shown in Table 4.8-2. For each port position data point, a set of three consecutive off-gas samples were collected using the Tedlar bags to obtain the ambient air background compensated average. A summary of the readings is presented in Table 4.8-3.

Table 4.8-1 Synopsis of Air Stripper Off-Gas Testing Program

Page 1 of 2

| Week of | Test Date | Testing Performed |
|----------|-----------|-----------------------------|
| 10/04/92 | 10/08/92 | Real-time only ¹ |
| 10/11/92 | 10/15/92 | Real-time only ¹ |
| 10/18/92 | 10/21/92 | Real-time only ¹ |
| 10/25/92 | 10/29/92 | Real-time only ¹ |
| 11/01/92 | 11/05/92 | Real-time only ¹ |
| 11/08/92 | 11/12/92 | Real-time only ¹ |
| 11/15/92 | 11/17/92 | Real-time only ¹ |
| 11/22/92 | 11/27/92 | Real-time only ¹ |
| 11/30/92 | 12/02/92 | Real-time only ¹ |
| 12/06/92 | 12/11/92 | Real-time only ¹ |
| 12/13/92 | 12/17/92 | Weekly |
| 12/20/92 | 12/23/92 | Monthly |
| 12/29/92 | 12/29/92 | Weekly |
| 01/03/93 | 01/06/93 | Weekly |
| 01/10/93 | 01/15/93 | Weekly |
| 01/17/93 | 01/21/93 | Monthly |
| 01/24/93 | 01/28/93 | Weekly |
| 01/31/93 | 02/03/93 | Weekly |
| 02/07/93 | 02/09/93 | Weekly |
| 02/14/93 | 02/19/93 | Monthly |
| 02/21/93 | 02/26/93 | Weekly |
| 02/28/93 | 03/05/93 | Weekly |
| 03/07/93 | 03/10/93 | Weekly |
| 03/14/93 | 03/16/93 | Weekly |
| 03/21/93 | 03/23/93 | Monthly |
| 03/28/93 | 04/02/93 | Weekly |
| 04/04/93 | 04/08/93 | Weekly |
| 04/11/93 | 04/15/93 | Weekly |
| 04/18/93 | 04/23/93 | Monthly |

Table 4.8-1 Synopsis of Air Stripper Off-Gas Testing Program

Page 2 of 2

| Week of | Test Date | Testing Performed |
|----------------------|-----------|-------------------|
| 04/25/93 | 04/28/93 | Weekly |
| 05/02/93 | 05/04/93 | Weekly |
| 05/09/93 | 05/11/93 | Weekly |
| 05/16/93 | 05/19/93 | Weekly |
| 05/23/93 | 05/27/93 | Monthly |
| 05/30/93 | 06/03/93 | Weekly |
| 06/06/93 | 06/08/93 | Weekly |
| 06/13/93 to 07/11/93 | NA | None ² |
| 07/18/93 | 07/20/93 | Monthly |
| 07/25/93 to 09/29/93 | NA | None ² |

NA - Not Applicable

1 - Sampling media from laboratory not available

2 - Air stripper inoperative during these periods

Table 4.8-2 FY93 Air Stripper Real-Time Testing Results

Page 1 of 2

| Sample Date | OVA (ppm) | | | | OVM (ppm) | | | |
|-------------|-----------|-------------|------------|-----------|-----------|-------------|------------|-----------|
| | Ambient | Downstream* | Midstream* | Upstream* | Ambient | Downstream* | Midstream* | Upstream* |
| 10-08-92 | 1.4 | 0.2 | 1.2 | NS | 0.1 | 0.0 | 0.0 | NS |
| 10-15-92 | 0.3 | 0.1 | 0.9 | 2.7 | 0.2 | 0.0 | 0.1 | 0.1 |
| 10-21-92 | 1.2 | 1.9 | 3.7 | 3.7 | 0.2 | 0.0 | 0.1 | 0.1 |
| 10-29-92 | 1.1 | 4.1 | 6.1 | 6.2 | 0.0 | 0.1 | 0.0 | 0.0 |
| 11-05-92 | 1.1 | 1.2 | 2.8 | 4.2 | 0.0 | 0.0 | 0.0 | 0.1 |
| 11-12-92 | 0.9 | 2.0 | 2.7 | 5.9 | 0.0 | 0.0 | 0.0 | 0.0 |
| 11-17-92 | 1.3 | 2.7 | 5.1 | 12.3 | 0.0 | 0.2 | 0.2 | 0.3 |
| 11-27-92 | 0.5 | 0.7 | 1.2 | 2.0 | 0.3 | 0.0 | 0.1 | 0.2 |
| 12-02-92 | 3.9 | 3.3 | 29.1 | 33.8 | 0.0 | 0.0 | 0.0 | 0.0 |
| 12-11-92 | 1.6 | 3.1 | 5.9 | 6.1 | 0.2 | 0.0 | 0.1 | 0.1 |
| 12-17-92 | 0.3 | 3.3 | 5.5 | 5.6 | 0.0 | 0.6 | 0.8 | 1.2 |
| 12-23-92 | 1.5 | 3.9 | 5.2 | 5.1 | 0.0 | 0.2 | 0.3 | 0.4 |
| 12-29-92 | 2.2 | 2.9 | 2.5 | 3.2 | 0.0 | 0.1 | 0.1 | 0.2 |
| 01-06-93 | 0.8 | 4.7 | 4.9 | 5.2 | 0.0 | 0.2 | 0.3 | 0.5 |
| 01-15-93 | 2.0 | 4.8 | 5.4 | 5.2 | 0.0 | 0.0 | 0.4 | 0.3 |
| 01-21-93 | 1.9 | 4.3 | 4.8 | 4.7 | 0.0 | 0.0 | 0.0 | 0.0 |
| 01-28-93 | 1.4 | 4.3 | 4.7 | 0.6 | 0.0 | 0.2 | 0.2 | 0.3 |
| 02-03-93 | 1.8 | 4.1 | 4.3 | 5.0 | 0.0 | 0.0 | 0.0 | 0.1 |
| 02-09-93 | 2.2 | 4.0 | 4.1 | 4.9 | 0.0 | 0.0 | 0.0 | 0.3 |
| 02-19-93 | 1.8 | 4.3 | 3.6 | 5.6 | 0.0 | 0.0 | 0.0 | 0.0 |

* - Note: these results are background compensated (i.e., Result-Ambient reading)

OVA - Organic Vapor Analyser
 OVM - Organic Vapor Meter
 ppm - Parts per million

RMA/0893 10/20/94 8:53 am ap

Table 4.8-2 FY93 Air Stripper Real-Time Testing Results

Page 2 of 2

| Sample Date | OVA (ppm) | | | | OVM (ppm) | | | |
|-------------|-----------|-------------|------------|-----------|-----------|-------------|------------|-----------|
| | Ambient | Downstream* | Midstream* | Upstream* | Ambient | Downstream* | Midstream* | Upstream* |
| 02-26-93 | 0.0 | 3.2 | 3.4 | 5.7 | 0.0 | 0.0 | 0.0 | 0.1 |
| 03-05-93 | 1.9 | 4.0 | 4.6 | 7.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 03-10-93 | 1.6 | 0.3 | 0.3 | 6.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 03-16-93 | 1.7 | 0.2 | 0.2 | 6.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| 03-23-93 | 1.9 | 0.3 | 0.3 | 5.8 | 0.0 | 0.0 | 0.0 | 0.0 |
| 04-02-93 | 1.8 | 0.4 | 0.9 | 5.4 | 0.0 | 0.0 | 0.0 | 0.0 |
| 04-08-93 | 2.0 | 0.3 | 1.6 | 4.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 04-15-93 | 2.0 | 0.3 | 3.1 | 5.4 | 0.0 | 0.0 | 0.0 | 0.0 |
| 04-23-93 | 0.9 | 0.1 | 3.1 | 4.7 | 0.0 | 0.0 | 0.0 | 0.0 |
| 04-28-93 | 1.8 | 0.0 | 3.6 | 4.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 05-05-93 | 1.7 | 0.2 | 3.5 | 2.9 | 0.0 | 0.0 | 0.0 | 0.0 |
| 05-11-93 | 1.9 | 0.1 | 2.6 | 10.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| 05-19-93 | 1.8 | 0.4 | 2.9 | 10.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 05-27-93 | 2.0 | 1.0 | 9.1 | 13.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 06-03-93 | 0.9 | 1.6 | 11.1 | 8.7 | 0.0 | 0.0 | 0.0 | 0.0 |
| 06-08-93 | 1.1 | 1.7 | 11.9 | 15.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 07-20-93 | 2.0 | 1.0 | 9.3 | 9.8 | 0.0 | 0.0 | 0.0 | 0.0 |

* - Note: these results are background compensated (i.e., Result-Ambient reading)

OVA - Organic Vapor Analyser
 OVM - Organic Vapor Meter
 ppm - Parts per million

RMA/0893 10/20/94 8:53 am ap

Table 4.8-3 Summary of FY93 Air Stripper Real-Time Testing Results

Page 1 of 1

| Location | OVA Average (ppm) | OVA Maximum (ppm) | OVM Average (ppm) | OVM Maximum (ppm) |
|-------------|----------------------|----------------------|----------------------|----------------------|
| Ambient | 1.5 | 3.9 | 0.03 | 0.3 |
| Upstream* | 6.9 | 33.8 | 0.12 | 1.2 |
| Midstream* | 4.7 | 29.1 | 0.07 | 0.8 |
| Downstream* | 2.0 | 4.8 | 0.04 | 0.6 |

* Note: These readings are background compensated (i.e. Result-Ambient)

OVA - Organic Vapor Analyzer
OVM - Organic Vapor Meter
ppm - Parts per million

As indicated in Table 4.8-3, the overall direct reading instrument averages indicate that concentrations measured at the midstream and downstream ports are consistently lower than those measured at the upstream port. The trend of decreasing concentrations measured from upstream, to midstream, and then downstream ports is consistent for both OVA and OVM average and maximum values. The average VOC reductions from upstream to downstream port positions as indicated by the OVA and OVM, were 71 percent and 67 percent, respectively. The maximum upstream and midstream OVA readings were 33.8 and 29.1 ppm, respectively and were observed on December 2, 1992. The high OVA readings on that date may be associated with the use of air stripper descaling agent in the raw feed water. On December 2, 1992, the sampling technicians observed a citrus type odor present near the exterior doors of the air stripper room in the area where the descaling solution tank is located.

4.8.2 Air Stripper Sorbent Tube Sampling Results

A synopsis of the FY93 air stripper sampling events using sorbent tubes is presented in Table 4.8-4, and results from analysis of the Tenax and XAD tubes are summarized in Table 4.8-5. In general, the air stripper sample analyses were performed using the same methods as in the CAQMMP ambient air VOC and SVOC sampling. The complete data set and results for field blanks are provided in Appendix E. As seen in the data listing, results for many compounds were less than the CRL at all sampling ports. Those compounds detected most frequently include methylene chloride, chloroform, toluene, and benzene. Average values for the summarized analytical data indicate that 1,2-dichloroethene, methylene chloride, chloroform, and toluene were detected at the upstream, midstream, and downstream ports at concentrations greater than 5 $\mu\text{g}/\text{m}^3$. Analysis of the field blanks indicate results which were generally less than the CRLs for the various target compounds.

Methylene chloride, a non-certified compound, and chloroform were both frequently detected at concentrations above the CRLs and were significantly higher than the other target compounds. Many reported values for these two compounds were based upon post-analytical laboratory estimates as discussed in Section 4.6. The highest average methylene chloride concentration

Table 4.8-4 Synopsis of FY93 Air Stripper Sorbent Tube Sampling

Page 1 of 1

| Location | VOC Samples Collected | SVOC Samples Collected |
|---------------------|-----------------------|------------------------|
| Upstream | 7 | 7 |
| Midstream | 26 | 21 |
| Downstream | 7 | 7 |
| Ambient | 5 | 5 |
| Midstream duplicate | 7 | 7 |

VOC - Volatile Organic Compound

SVOC - Semivolatile Organic Compound

Table 4.8-5 Summary of FY93 Air Stripper Volatile Organic Compound Concentrations (µg/m³)

| Location | 111TCE | 112TCE | 11DCLE | 12DCLE | 12DCE | BCHPD | C6H6 | CCL4 | CH2CL2 | CHCL3 |
|---------------------|---------|--------|--------|--------|-------|-------|-------|-------|--------|---------|
| Upstream | Average | < 2.3 | < 3.6 | 2.6 | 7.6 | 6.3 | 5.1 | < 1.8 | 656.7 | 3226.0 |
| | Maximum | 3.6 | < 3.6 | 6.0 | 15.0 | 14.0 | 9.2 | < 1.8 | 2410.0 | 5280.0 |
| Midstream | Average | 2.4 | < 3.6 | 2.6 | 8.1 | 2.7 | 2.6 | 3.8 | 282.3 | 2321.0 |
| | Maximum | 9.6 | < 3.6 | 8.9 | 46.0 | 52.0 | 20.0 | 73.0 | 1368.0 | 6942.0 |
| Midstream Duplicate | Average | 1.8 | < 3.6 | 2.0 | 3.9 | 2.6 | 2.5 | < 1.8 | 351.7 | 1729.0 |
| | Maximum | 4.1 | < 3.6 | 5.0 | 12.0 | 8.8 | 6.4 | < 1.8 | 545.0 | 5020.0 |
| Downstream | Average | 3.0 | < 3.6 | 3.1 | 6.1 | < 1.2 | < 1.6 | < 1.8 | 1088.7 | 3119.0 |
| | Maximum | 11.0 | < 3.6 | 13.0 | 39.0 | < 1.2 | 3.2 | < 1.8 | 5010.0 | 15830.0 |
| Ambient | Average | 2.3 | < 3.6 | < 1.2 | < 1.2 | < 1.2 | 4.5 | < 1.8 | 477.4 | < 1.2 |
| | Maximum | 7.1 | < 3.6 | < 1.2 | < 1.2 | < 1.2 | 10.8 | < 1.8 | 2300.0 | 3.2 |

| | CLC6H5 | DBCP | DCPD | DMDS | ETC6H5 | MEC6H5 | MIBK | TCLEE | TRCLE | XYLENES |
|---------------------|---------|-------|--------|-------|--------|--------|--------|-------|-------|---------|
| Upstream | Average | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | 5.6 | 4.2 | < 15.0 |
| | Maximum | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | 20.0 | 8.4 | < 15.0 |
| Midstream | Average | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | 2.8 | < 1.3 | < 15.0 |
| | Maximum | < 1.3 | < 10.0 | < 8.9 | < 4.7 | 6.0 | < 16.0 | 44.0 | < 1.3 | 26.0 |
| Midstream Duplicate | Average | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | < 1.3 | < 1.3 | < 15.0 |
| | Maximum | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | < 1.3 | 3.1 | < 15.0 |
| Downstream | Average | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | < 1.3 | < 1.3 | < 15.0 |
| | Maximum | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | < 1.3 | < 1.3 | < 15.0 |
| Ambient | Average | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | < 1.3 | < 1.3 | < 15.0 |
| | Maximum | < 1.3 | < 10.0 | < 8.9 | < 4.7 | < 3.2 | < 16.0 | < 1.3 | < 1.3 | < 15.0 |

| | | | | | | | | | | | |
|--------|---|--------------------------|--------|---|----------------------|--------|---|----------------------|---------|---|------------------------|
| 111TCE | - | 1,1,1-Trichloroethane | BCHPD | - | Bicycloheptadiene | CLC6H5 | - | Chlorobenzene | MEC6H5 | - | Toluene |
| 112TCE | - | 1,1,2-Trichloroethane | C6H6 | - | Benzene | DBCP | - | Dibromochloropropane | MIBK | - | Methyl Isobutyl Ketone |
| 11DCLE | - | 1,1-Dichloroethane | CCL4 | - | Carbon Tetrachloride | DCPD | - | Dicyclopentadiene | TCLEE | - | Tetrachloroethene |
| 12DCLE | - | 1,2-Dichloroethane | CH2CL2 | - | Methylene Chloride | DMDS | - | Dimethyldisulfide | TRCLE | - | Trichloroethene |
| 12DCE | - | trans-1,2-Dichloroethene | CHCL3 | - | Chloroform | ETC6H5 | - | Ethylbenzene | XYLENES | - | Total Xylenes |

< ##
µg/m³
Indicates concentration less than the lower certified reporting limit
Micrograms per meter

was 1088.7 $\mu\text{g}/\text{m}^3$ at the downstream port. Upstream and midstream methylene chloride average concentrations were 656.7 and 282.3 $\mu\text{g}/\text{m}^3$, respectively. The average ambient air background concentration for methylene chloride was 477.4 $\mu\text{g}/\text{m}^3$. However, this was based on a 1-day (December 23, 1992) very high measurement; all other measurements were less than 3 $\mu\text{g}/\text{m}^3$; in addition, the field blank on this day indicated a high level of methylene chloride, making the sample suspect.

Average chloroform concentrations were the highest overall among all compounds at each sampling port location. The upstream, midstream, and downstream chloroform average concentrations were 3226.0, 2321.0, and 3119.0 $\mu\text{g}/\text{m}^3$, respectively. Results of the blank sample analyses indicate values less than the lower CRL. Average ambient air background concentrations for chloroform were also less than the lower CRL. It should also be noted that the concentrations measured for all stripper analytes represent a very small sample volume for a short period (i.e., 80 minutes), and cannot be directly compared to the 24-hour measurements collected under the routine CAQMMP VOC sampling program.

Evaluation of the analytical results with respect to performance of the two carbon adsorption units indicate progressive reduction of off-gas concentrations from the first to the second unit in the case of 1,2-dichloroethene, bicycloheptadiene, benzene, carbon tetrachloride, toluene, and tetrachloroethene; exceptions to this sequence were methylene chloride and chloroform, suggesting possible concerns in system performance for these compounds, although laboratory analysis may also be a factor; CAQMMP will further investigate these results. Average results for 1,1,2-trichloroethane, trans-1,2-dichloroethene, chlorobenzene, dibromochloropropane, dicyclopentadiene, dimethyldisulfide, ethylbenzene, methyl isobutyl ketone, and total xylenes were less than the lower CRL for samples collected at all sampling locations. Analytical results for duplicate samples indicated general agreement with primary sample values with the exception of 1,2-dichloroethane and toluene. Possible reasons for the nonconformance of the paired sample results for these two compounds are not readily apparent.

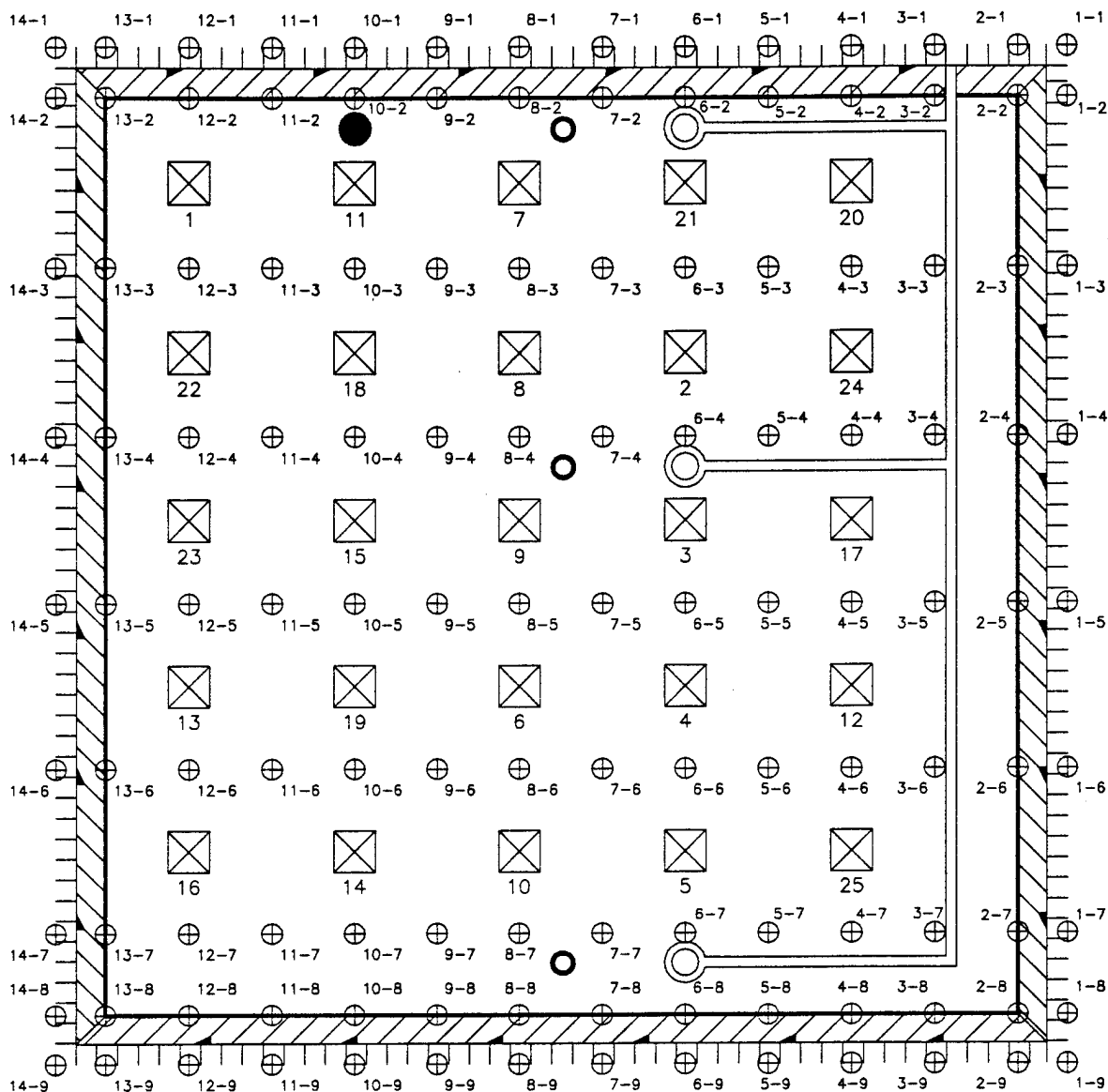
Results of the XAD tube analyses indicated that concentrations for all SVOC target compounds were less than the lower CRLs, or detection limits, for noncertified compounds. Results of field blank analyses indicated no detections.

4.9 REAL-TIME CAP AND VENT MONITORING

4.9.1 Program Description

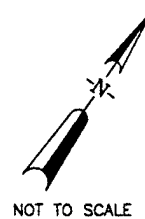
The FY93 CAQMMP included periodic monitoring of the restored Basin F floor, Basin F waste pile cap surface, Basin F waste pile vents, and vents associated with Basin F liquid storage at Pond A and the three holding tanks. Both real-time monitoring and VOC sampling with sorbent tubes were performed. The stated testing schedule entailed seasonal real-time monitoring and quarterly VOC sampling with the Tenax and Tenax/Charcoal tubes. The VOC sampling was actually performed in conjunction with real-time monitoring during periods when prevailing atmospheric conditions were conducive to the development of positive vent outflows. The real-time monitoring and sampling were conducted to characterize emissions of the sources and aid in assessing their possible impact on ambient air quality. Real-time monitoring was performed using portable instrumentation which included an OVA, an OVM, an ammonia meter, and a hydrogen sulfide meter (see Sections 3.3.4, 3.3.8, and 3.3.9 for descriptions of these instruments).

During construction of the Basin F waste pile, 25 passive off-gas release vents were installed through the liner to relieve potential air pressure gradients from recurring atmospheric pressure changes and from internal vaporization of contaminants. The vents are spaced in an irregular grid across the waste pile, as indicated in Figure 4.9-1, and the vents are numbered randomly. The height of each vent varies with the depth of the waste pile cap and ranges from 1- to 3-ft above ground. The vent pipes are 6 inches in diameter and terminate in an elbow facing the ground. The waste pile vents penetrate an average of 1.5 feet through the soil cover to the plastic liner cap. The vents allow passive emission of soil gases that have passed through the pile material and accumulated beneath the upper liner. Vent emission flow rates are typically negligible due to the shallow depths of the vent pipes and the relative absence of void spaces or active offgassing conditions within the compacted waste pile. Offgassing rates are also



LEGEND

- 5 Vent with Existing Vent Number
- Sand Bags
- Baseplate Standpipe
- Detection Sump and Baseplate Standpipe
- Leachate Collection Sump and Baseplate Standpipe
- Transfer Pipe



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.9-1

Waste Pile Vent Locations

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

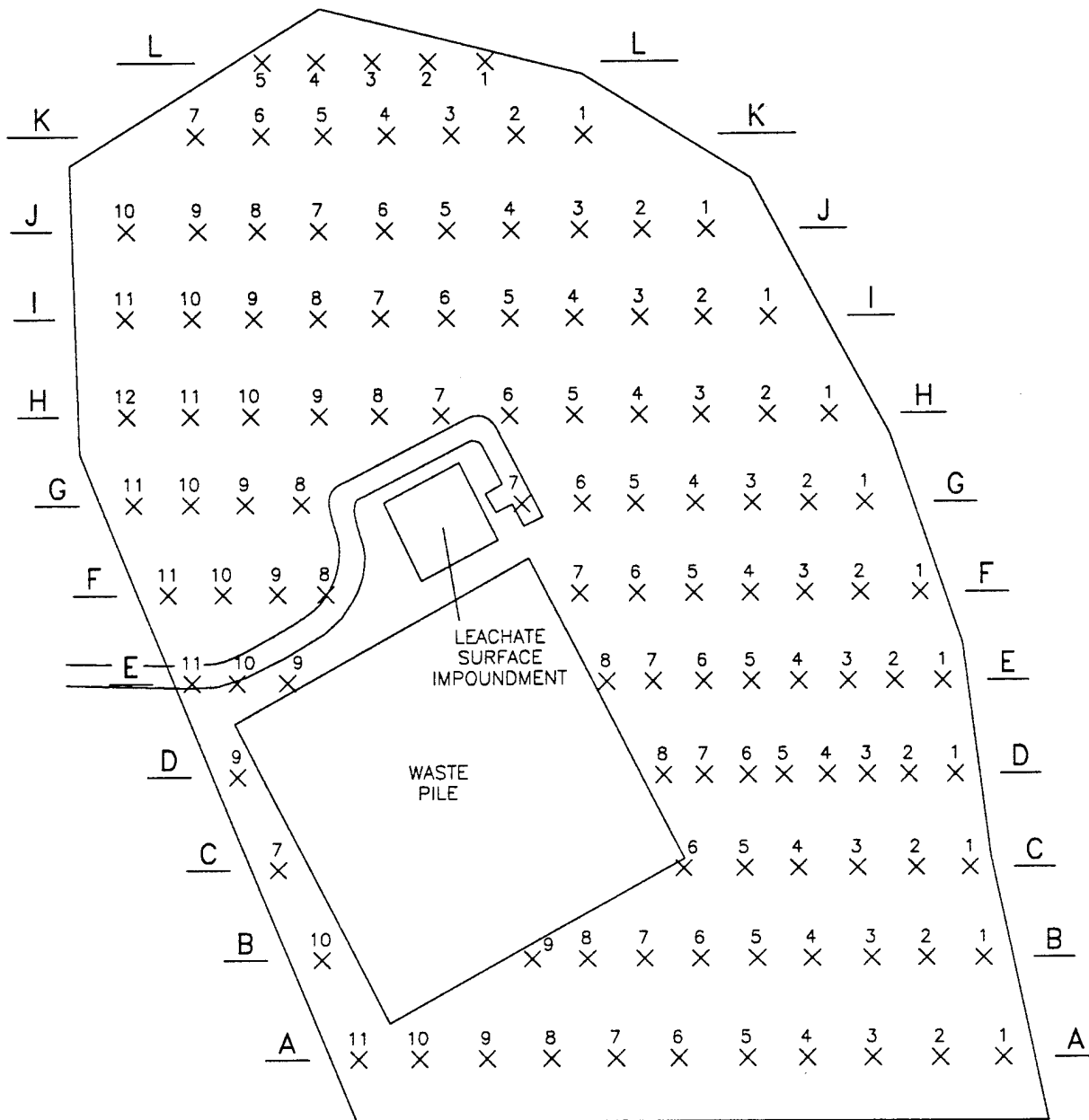
influenced by barometric pressure trends, with prolonged periods of high pressure being conducive for the accumulation of ambient air beneath the vents within the upper portions of the waste pile. A subsequent decrease in barometric pressure will enhance the brief outflow of accumulated ambient air and soil gas. Vent flow rate tests have confirmed a typically neutral condition of vent flow. Therefore, emissions of volatiles and semivolatiles from the waste pile vents occur mainly as a result of passive diffusion from each vent to the atmosphere.

Real-time monitoring of the waste pile vents took place four times during the FY93 period. The readings were taken about 1 inch below the vent opening on the downwind side, during periods when winds were light (less than 10 mph) and atmospheric pressure was dropping. Readings were taken with the four real-time instruments over an approximate 30-second interval.

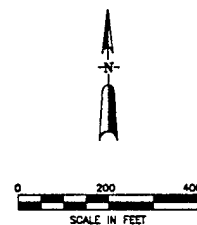
Real-time readings of the waste pile cap surface were also taken. The four real-time instruments were fitted with an extension tube ending in a small funnel. The technician traversed the waste pile cap in a predetermined pattern and collected representative readings at a total of 126 marked locations regularly spaced across the waste pile. These locations are shown in Figure 4.9-1. The sample was drawn from approximately 1 inch above ground at each location, and real-time readings were recorded on field data sheets.

Similar readings were taken from the restored Basin F floor, which is covered with a clay and topsoil cap. The Basin F floor readings were taken at 115 regularly spaced locations marked by sandbags, as depicted in Figure 4.9-2. Weather conditions and the physical appearance of the floor during each sampling episode were recorded prior to commencement of sampling.

Pond A is a double-lined liquid storage pond located just north of Basin F. The pond is covered with a high density polyethylene cover and has four vents, one in the middle of each side of the pond. The vents are equipped with trip valves that open to allow emission of gases and close to prevent entrainment of air beneath the pond cover. As a result, the pond gases are emitted in short pulses. Sampling the pond vents required a sample duration that spanned several pulses



LEGEND
 11
 X Sand Bags



Prepared for:
 U.S. Army Program Manager
 for Rocky Mountain Arsenal
 Figure 4.9-2
 Basin F Floor Sampling
 Locations
 Rocky Mountain Arsenal
 Prepared by: Ebasco Services Incorporated

of the trip valve. Pond A vents were tested with the real-time instrument probe inlets held near the vent openings.

The three storage tanks for the Basin F liquid are located northeast of the basin floor and are covered by metal roofs with passive vents atop each tank. Real-time tank vapor readings were taken with the four real-time instruments. Extension tubes were utilized to allow the instruments to draw vent gas from the vent opening to the platform location.

VOCs were also sampled at the waste pile vents, Pond A, and the three storage tanks. As noted, the VOC sampling was performed in conjunction with real-time monitoring activities. The same type of Tenax front and Tenax/charcoal backup tubes that were used routinely for ambient air sampling were used for vent sampling. A flow rate of 85 standard cubic centimeters per minute (sccm) was used for sampling over a 60-minute period. For samples collected at the selected pond, tank, and waste pile vents, the Tenax tube inlets were monitored in approximately the same positions used for real-time monitoring. Both trip blanks and field blanks were submitted with the VOC samples for analysis.

4.9.2 Basin F and Waste Pile Cap Monitoring Results

Results of real-time monitoring performed on the February 26, March 20, June 15, and October 4, 1993 at the Basin F floor locations indicate no detections were obtained with the OVA, OVM, and NH_3 instruments. Results of the restored Basin F floor and waste pile cap ground surface testing are summarized in Table 4.9-1. The basin floor and waste pile cap appear to be negligible sources of air emissions according to the levels detected by the real-time monitoring instruments utilized in this sampling program. In 1993, all OVA and OVM readings taken on the cap were 0.0 ppm. Furthermore, only 1 out of 920 readings taken at the basin floor in 1993 was not 0.0 (OVM 0.01 ppm). These results are comparable to previous years' data. During those monitoring events, there were numerous trace level recordings of H_2S . The averaged H_2S readings ranged from 0.0005 to 0.002 ppm and the maximum readings ranged from 0.001 to 0.005 ppm.

Table 4.9-1 Summary of Basin F Floor and Waste Pile Cap Monitoring Results

| Date | Location | OVA (ppm) | | | OVM (ppm) | | | NH ₃ (ppm) | | | H ₂ S (ppm) | | |
|---------|------------------------|-----------|---------|------|-----------|---|---------|-----------------------|------|---|------------------------|--------|-------|
| | | # | Detects | Avg. | Max. | # | Detects | Avg. | Max. | # | Detects | Avg. | Max. |
| 2/26/93 | Restored Basin F Floor | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0005 | 0.001 |
| 3/20/93 | Restored Basin F Floor | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0005 | 0.002 |
| 6/15/93 | Restored Basin F Floor | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.002 | 0.003 |
| 10/4/93 | Restored Basin F Floor | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.002 | 0.005 |
| 3/1/93 | Basin F Waste Pile Cap | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0009 | 0.004 |
| 3/20/93 | Basin F Waste Pile Cap | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0 | 0.0 | 4 | 4 | 0.0097 | 0.004 |
| 5/13/93 | Basin F Waste Pile Cap | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0011 | 0.002 |
| 10/1/93 | Basin F Waste Pile Cap | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0 | 0.0 | 0 | 0 | 0.0008 | 0.002 |

Note: Restored Basin F Floor - 115 sample locations
 Basin F Waste Pile Cap - 126 sample locations
 OVA - Organic Vapor Analyzer
 OVM - Organic Vapor Meter
 NH₃ - Ammonia
 H₂S - Hydrogen Sulfide
 ppm - Parts per million

Results of real-time waste pile cap monitoring that was performed March 1, March 20, May 13, and October 1, 1993 indicate no detections associated with the OVA and OVM instruments. Ammonia was detected at four locations on March 20, 1993 at an average concentration of 0.004 ppm. Hydrogen sulfide readings taken at various locations on the waste pile cap ranged from 0.007 to 0.011 ppm for averaged results and from 0.002 to 0.004 ppm for maximum readings.

4.9.3 Waste Pile Vent Monitoring Results

Results of real-time readings taken at the waste pile vents are presented in Table 4.9-2. During FY93, real-time readings of the vents were taken on February 1, March 23, August 4, and September 8, 1993. The highest annual average of the four readings, 21.8 ppm, was observed at Vent 1 during FY93. The maximum OVA reading for FY93, 54.0 ppm, was also observed at Vent 1 on August 4. Vents 1, 13, 16, 22, and 23 had the highest average OVA readings. The OVM, NH_3 , and H_2S readings were substantially lower than those obtained with the OVA. OVM readings averaged 0.14 ppm in 1993 while H_2S readings averaged 0.017 ppm; and NH_3 results indicated no detections.

Real-time readings were also taken at the three main sump covers. OVA readings at the sumps were not substantially different from vent readings. The OVM, NH_3 , and H_2S readings taken at the sumps were generally higher than those obtained at the vents.

During each quarterly monitoring event, the four waste pile vents with the highest OVA readings were sampled using VOC cartridges. In FY93, nine different vents were sampled, with vents 13 and 16 being sampled most frequently. Table 4.9-3 summarizes the results of this sampling. Chloroform was the most prevalent compound and was also measured at the highest levels. It was present in 5 of the 9 samples analyzed at levels that ranged from below detection at several vents to $5412 \mu\text{g}/\text{m}^3$ at Vent 23. Methylene chloride and tetrachloroethene were also detected frequently. Methylene chloride concentrations ranged up to $106.5 \mu\text{g}/\text{m}^3$ at Vent 8 and tetrachloroethene levels reached $250.0 \mu\text{g}/\text{m}^3$ at Vent 23. Other analytes regularly detected

Table 4.9-2 Basin F Waste Pile Vent Real-Time Monitoring Results

| Vent | OVA (ppm) | | | | OVM (ppm) | | | | NH ₃ (ppm) | | | | H ₂ S (ppm) | | | |
|--------|-----------|---------|--------|--------|-----------|---------|--------|--------|-----------------------|---------|--------|--------|------------------------|---------|--------|--------|
| | 2/1/93 | 3/23/93 | 8/4/93 | 9/8/93 | 2/1/93 | 3/23/93 | 8/4/93 | 9/8/93 | 2/1/93 | 3/23/93 | 8/4/93 | 9/8/93 | 2/1/93 | 3/23/93 | 8/4/93 | 9/8/93 |
| 1 | 4.8 | 4.2 | 54.0 | 24.0 | 0.5 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.012 | 0.001 | 0.0 | 0.0 |
| 2 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.053 | 0.001 | 0.0 | 0.0 |
| 3 | 3.4 | 6.0 | 0.5 | 0.0 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.045 | 0.002 | 0.0 | 0.0 |
| 4 | 0.7 | 1.2 | 2.2 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.024 | 0.003 | 0.0 | 0.0 |
| 5 | 0.8 | 1.8 | 2.4 | 3.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.036 | 0.002 | 0.0 | 0.0 |
| 6 | 4.8 | 3.2 | 4.6 | 3.2 | 0.5 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.022 | 0.003 | 0.0 | 0.0 |
| 7 | 6.6 | 0.8 | 0.0 | 1.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.039 | 0.001 | 0.0 | 0.0 |
| 8 | 6.2 | 0.8 | 0.3 | 4.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.028 | 0.001 | 0.0 | 0.0 |
| 9 | 5.2 | 3.2 | 1.8 | 7.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.024 | 0.003 | 0.0 | 0.0 |
| 10 | 4.4 | 0.4 | 1.4 | 2.6 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.022 | 0.003 | 0.0 | 0.0 |
| 11 | 0.0 | 0.0 | 0.1 | 0.0 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.013 | 0.000 | 0.0 | 0.0 |
| 12 | 0.2 | 0.6 | 0.6 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.006 | 0.005 | 0.0 | 0.0 |
| 13 | 5.6 | 17.0 | 6.8 | 18.0 | 1.0 | 0.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.017 | 0.002 | 0.0 | 0.0 |
| 14 | 2.8 | 1.8 | 2.8 | 2.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.020 | 0.004 | 0.0 | 0.0 |
| 15 | 0.0 | 0.6 | 0.6 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.013 | 0.002 | 0.0 | 0.0 |
| 16 | 10.8 | 26.0 | 19.0 | 12.0 | 0.5 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.016 | 0.003 | 0.0 | 0.0 |
| 17 | 1.0 | 0.4 | 0.2 | 1.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.005 | 0.062 | 0.0 | 0.0 |
| 18 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.012 | 0.001 | 0.0 | 0.0 |
| 19 | 0.2 | 8.4 | 4.6 | 1.4 | 0.5 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.016 | 0.002 | 0.0 | 0.0 |
| 20 | 0.6 | 0.4 | 0.0 | 1.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.004 | 0.160 | 0.0 | 0.0 |
| 21 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.042 | 0.001 | 0.0 | 0.0 |
| 22 | 1.6 | 2.8 | 48.0 | 18.0 | 0.5 | 0.0 | 0.0 | 2.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.013 | 0.001 | 0.0 | 0.0 |
| 23 | 2.0 | 2.6 | 5.0 | 12.2 | 0.5 | 0.6 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.015 | 0.002 | 0.0 | 0.0 |
| 24 | 1.2 | 1.0 | 0.2 | 0.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.004 | 0.066 | 0.0 | 0.0 |
| 25 | 2.8 | 9.6 | 1.0 | 3.8 | 0.5 | 0.6 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.005 | 0.000 | 0.0 | 0.0 |
| Sump 1 | 0.6 | 10.2 | 0.5 | 2.0 | 0.0 | 14.0 | 0.0 | 0.0 | 4.0 | 46.1 | 12.0 | 4.0 | 0.119 | 0.420 | 0.0 | 0.0 |
| Sump 2 | 7.2 | 18.0 | 7.0 | 2.0 | 7.1 | 18.2 | 2.2 | 0.0 | 16.0 | 44.0 | >150.0 | 0.0 | 0.330 | 1.300 | 0.0 | 0.0 |
| Sump 3 | 2.8 | 14.2 | 6.4 | 18.0 | 2.5 | 16.8 | 0.4 | 10.0 | 2.0 | > 60.0 | 10.0 | 9.0 | 0.046 | 2.100 | 0.0 | 0.0 |

> ### - Indicates concentration above instrument upper detection limit

OVA - Organic Vapor Analyzer

OVM - Organic Vapor Meter

NH₃ - AmmoniaH₂S - Hydrogen Sulfide

ppm - Parts per million

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Table 4.9-3 Summary of FY93 Waste Pile Vents Volatile Organic Compound Concentrations (µg/m³) Page 1 of 2

| | | 111TCE | 112TCE | 11DCE | 12DCE | BCHPD | C6H6 | CCL4 | CH2CL2 | CHCL3 |
|---------|---------|--------|--------|-------|-------|-------|-------|-------|--------|--------|
| Vent 1 | Average | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | < 3.1 | < 3.5 | 43.2 | 2199.0 |
| | Maximum | 6.3 | < 7.1 | 2.9 | < 2.4 | < 2.4 | < 3.1 | < 3.5 | 76.0 | 3618.0 |
| Vent 5 | Average | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 4.7 | < 3.5 | 19.6 | < 2.4 |
| | Maximum | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 4.7 | < 3.5 | 19.6 | < 2.4 |
| Vent 7 | Average | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 5.0 | < 3.5 | 76.7 | < 2.4 |
| | Maximum | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 5.0 | < 3.5 | 76.7 | < 2.4 |
| Vent 8 | Average | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | < 3.1 | < 3.5 | 106.5 | < 2.4 |
| | Maximum | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | < 3.1 | < 3.5 | 106.5 | < 2.4 |
| Vent 13 | Average | < 4.5 | < 7.1 | < 2.4 | < 2.4 | 13.0 | < 3.1 | < 3.5 | 41.8 | 2243.0 |
| | Maximum | < 4.5 | < 7.1 | < 2.4 | < 2.4 | 21.0 | 4.9 | < 3.5 | 98.8 | 4200.0 |
| Vent 16 | Average | < 4.5 | < 7.1 | 3.9 | < 2.4 | < 2.4 | 6.2 | < 3.5 | 64.3 | 2668.0 |
| | Maximum | 7.7 | < 7.1 | 9.4 | < 2.4 | 180.0 | 8.6 | < 3.5 | 98.0 | 5314.0 |
| Vent 19 | Average | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 13.9 | < 3.5 | 28.3 | < 2.4 |
| | Maximum | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 13.9 | < 3.5 | 28.3 | < 2.4 |
| Vent 22 | Average | 6.0 | < 7.1 | 5.6 | < 2.4 | < 2.4 | 3.6 | < 3.5 | 91.0 | 5150.0 |
| | Maximum | 9.8 | < 7.1 | 6.0 | < 2.4 | 2.7 | 5.6 | < 3.5 | 99.0 | 5300.0 |
| Vent 23 | Average | < 4.5 | < 7.1 | 6.9 | 4.0 | 21.0 | < 3.1 | < 3.5 | 50.0 | 5412.0 |
| | Maximum | < 4.5 | < 7.1 | 6.9 | 4.0 | 21.0 | < 3.1 | < 3.5 | 50.0 | 5412.0 |

< ### Indicates concentration less than the lower certified reporting level
 µg/m³ Micrograms per cubic meter

| | | | | | |
|--------|---|--------------------------|--------|---|----------------------|
| 111TCE | - | 1,1,1-Trichloroethane | BCHPD | - | Bicycloheptadiene |
| 112TCE | - | 1,1,2-Trichloroethane | C6H6 | - | Benzene |
| 11DCE | - | 1,1-Dichloroethane | CCL4 | - | Carbon Tetrachloride |
| 12DCE | - | trans-1,2-Dichloroethene | CH2CL2 | - | Methylene Chloride |
| | | | CHCL3 | - | Chloroform |

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Table 4.9-3 Summary of FY93 Waste Pile Vents Volatile Organic Compound Concentrations (µg/m³) Page 2 of 2

| | CLC6H5 | | | DBCP | DCPD | DMDS | ETC6H5 | MEC6H5 | MBK | TCLEE | TRCLE | XYLENES |
|---------|---------|-------|--------|--------|-------|-------|--------|--------|-------|-------|--------|---------|
| Vent 1 | Average | < 2.5 | < 20.0 | 128.5 | < 9.2 | < 6.3 | < 2.7 | < 31.0 | 7.3 | < 2.6 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | 230.0 | < 9.2 | < 6.3 | < 2.7 | < 31.0 | 11.0 | < 2.6 | < 29.2 | |
| Vent 5 | Average | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | < 2.7 | < 31.0 | < 2.5 | < 2.6 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | < 2.7 | < 31.0 | < 2.5 | < 2.6 | < 29.2 | |
| Vent 7 | Average | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | 8.1 | < 31.0 | < 2.5 | < 2.6 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | 8.1 | < 31.0 | < 2.5 | < 2.6 | < 29.2 | |
| Vent 8 | Average | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | 4.6 | < 31.0 | < 2.5 | < 2.6 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | 4.6 | < 31.0 | < 2.5 | < 2.6 | < 29.2 | |
| Vent 13 | Average | < 2.5 | < 20.0 | 24.6 | < 9.2 | < 6.3 | 3.1 | < 31.0 | 19.3 | < 2.6 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | 52.0 | < 9.2 | < 6.3 | 4.9 | < 31.0 | 37.0 | < 2.6 | < 29.2 | |
| Vent 16 | Average | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | 5.5 | < 31.0 | 86.6 | 6.5 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | 11.2 | < 31.0 | 240.0 | 17.0 | < 29.2 | |
| Vent 19 | Average | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | 4.7 | < 31.0 | < 2.5 | < 2.6 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | < 17.5 | < 9.2 | < 6.3 | 4.7 | < 31.0 | < 2.5 | < 2.6 | < 29.2 | |
| Vent 22 | Average | < 2.5 | < 20.0 | 92.5 | < 9.2 | < 6.3 | < 2.7 | < 31.0 | 34.5 | 4.8 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | 95.0 | < 9.2 | < 6.3 | 3.8 | < 31.0 | 41.0 | 5.8 | < 29.2 | |
| Vent 23 | Average | < 2.5 | < 20.0 | 230.0 | < 9.2 | < 6.3 | < 2.7 | < 31.0 | 250.0 | 7.1 | < 29.2 | |
| | Maximum | < 2.5 | < 20.0 | 230.0 | < 9.2 | < 6.3 | < 2.7 | < 31.0 | 250.0 | 7.1 | < 29.2 | |

< ### Indicates concentration less than the lower certified reporting level
 µg/m³ Micrograms per cubic meter

| | | | | | |
|--------|---|----------------------|---------|---|-----------------------|
| CLC6H5 | - | Chlorobenzene | MEC6H5 | - | Toluene |
| DBCP | - | Dibromochloropropane | MBK | - | Methyl isobutylketone |
| DCPD | - | Dicyclopentadiene | TCLEE | - | Tetrachloroethene |
| DMS | - | Dimethyldisulfide | TRCLE | - | Trichloroethene |
| ETC6H5 | - | Ethylbenzene | XYLENES | - | Total Xylenes |

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included benzene and dicyclopentadiene. All analytes showed wide variations in concentrations. Individual analytical results for all waste pile vents and liquid storage vent samples are given in Appendix H. Short-term variances in emission rates and airflow turbulence at the locations where direct readings and VOC samples were obtained possibly contribute to the observed instrumental and analytical differences. Still, the use of a nonspeciating, direct-reading instrument remains the most practical method of selecting vents to sample when the prohibitive cost of laboratory analysis for sampling at all vents is considered.

4.9.4 Pond A and Tank Farm Monitoring Results

Results of real-time monitoring at the Pond A vents are presented in Table 4.9-4. The NH_3 and H_2S results were typically 0.0 ppm with the exception of H_2S readings ranging from 0.007 to 0.011 ppm at the four vents on January 4, 1993. OVM results were all 0.0 ppm with the exception of consistent readings of 0.4 ppm at all four vents on August 23, 1993. Pond A vent measurements performed using the OVA ranged from 0.0 to 40.0 ppm. Averaged OVA readings for all four vents indicate concentrations of 7.1, 0.6, 6.4, and 13.5 ppm for the north, east, south, and west vents, respectively.

Analytical results for Pond A vent samples are presented in Table 4.9-5. Compounds detected included 1,1,1-trichloroethane, 1,1-dichloroethane, benzene, methylene chloride, chloroform, dimethylsulfide, toluene, tetrachloroethene, total xylenes, and trichloroethene.

Results of real-time monitoring for the Basin F tank farm are presented in Table 4.9-6. Measurements made with the real-time instruments ranged widely for each parameter and among the three tanks. Averaged OVA readings indicate values of 10.1, 10.3, and 12.0 ppm for the north, east, and south tanks, respectively. On September 8, 1993, a measurement of 22.0 ppm at the south side tank vent was observed as the maximum OVA measurement. The highest OVM measurement, 12.0 ppm, was observed on September 8, 1993, also at the south tank vent. The highest NH_3 measurement, 28.0 ppm, was measured on July 2, 1993 at the north tank vent.

Table 4.9-4 Pond A Real-Time Monitoring Results

Page 1 of 1

| Date | Location | OVA (ppm) | OVM (ppm) | NH ₃ (ppm) | H ₂ S (ppm) |
|---------|---------------------------------|--------------|--------------|--------------------------|---------------------------|
| 1/04/93 | Pond A Liner Vents - North Vent | 0.0 | 0.0 | 0.0 | 0.010 |
| | Pond A Liner Vents - East Vent | 0.0 | 0.0 | 0.0 | 0.010 |
| | Pond A Liner Vents - South Vent | 0.8 | 0.0 | 0.0 | 0.007 |
| | Pond A Liner Vents - West Vent | 0.0 | 0.0 | 0.0 | 0.011 |
| 3/29/93 | Pond A Liner Vents - North Vent | 11.0 | 0.0 | 0.0 | 0.000 |
| | Pond A Liner Vents - East Vent | 0.0 | 0.0 | 0.0 | 0.000 |
| | Pond A Liner Vents - South Vent | 0.0 | 0.0 | 0.0 | 0.000 |
| | Pond A Liner Vents - West Vent | 40.0 | 0.0 | 0.0 | 0.000 |
| 7/15/93 | Pond A Liner Vents - North Vent | 7.2 | 0.0 | 0.0 | 0.000 |
| | Pond A Liner Vents - East Vent | 2.2 | 0.0 | 0.0 | 0.000 |
| | Pond A Liner Vents - South Vent | 12.0 | 0.0 | 0.0 | 0.000 |
| | Pond A Liner Vents - West Vent | 1.8 | 0.0 | 0.0 | 0.000 |
| 8/23/93 | Pond A Liner Vents - North Vent | 10.1 | 0.4 | 0.0 | 0.000 |
| | Pond A Liner Vents - East Vent | 0.0 | 0.4 | 0.0 | 0.000 |
| | Pond A Liner Vents - South Vent | 12.6 | 0.4 | 0.0 | 0.000 |
| | Pond A Liner Vents - West Vent | 12.0 | 0.4 | 0.0 | 0.000 |

OVA - Organic Vapor Analyzer
 OVM - Organic Vapor Meter
 NH₃ - Ammonia
 H₂S - Hydrogen sulfide
 ppm - Parts per million

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Table 4.9-5 Summary of FY93 Pond A Volatile Organic Compound Concentrations (µg/m³) Page 1 of 1

| | 111TCE | 112TCE | 11DCLE | 12DCLE | 12DCE | BCHPD | C6H6 | CCL4 | CH2CL2 | CHCL3 |
|---------|--------|--------|--------|--------|--------|--------|--------|--------|--------|---------|
| Pond A | | | | | | | | | | |
| Average | 44.8 | < 7.1 | 2.6 | < 2.4 | < 2.4 | < 2.4 | 3.5 | < 3.5 | 37.8 | 3.4 |
| Maximum | 113.0 | < 7.1 | 6.9 | < 2.4 | < 2.4 | < 2.4 | 9.4 | < 3.5 | 102.0 | 6.9 |
| | CLC6H5 | DBCP | DCPD | DMDS | ETC6H5 | MEC6H5 | MIBK | TCL4EE | TRCLE | XYLENES |
| Pond A | | | | | | | | | | |
| Average | < 2.5 | < 20.0 | < 17.5 | 66.5 | < 6.3 | 9.6 | < 31.0 | 141.0 | 82.1 | < 29.2 |
| Maximum | < 2.5 | < 20.0 | < 17.5 | 160.0 | 9.6 | 26.5 | < 31.0 | 333.0 | 196.0 | 36.9 |

< ## indicates concentration less than the lower certified reporting limit
µg/m³ Micrograms per cubic meter

| | | | | | | | | | | | |
|--------|---|--------------------------|--------|---|----------------------|--------|---|----------------------|---------|---|------------------------|
| 111TCE | - | 1,1,1-Trichloroethane | BCHPD | - | Bicycloheptadiene | CLC6H5 | - | Chlorobenzene | MEC6H5 | - | Toluene |
| 112TCE | - | 1,1,2-Trichloroethane | C6H6 | - | Benzene | DBCP | - | Dibromochloropropane | MIBK | - | Methyl Isobutyl Ketone |
| 11DCLE | - | 1,1-Dichloroethane | CCL4 | - | Carbon Tetrachloride | DCPD | - | Dicyclopentadiene | TCL4EE | - | Tetrachloroethene |
| 12DCLE | - | 1,2-Dichloroethane | CH2CL2 | - | Methylene Chloride | DMDS | - | Dimethyldisulfide | TRCLE | - | Trichloroethene |
| 12DCE | - | trans-1,2-Dichloroethene | CHCL3 | - | Chloroform | ETC6H5 | - | Ethylbenzene | XYLENES | - | Total Xylenes |

Table 4.9-6 Basin F Tank Farm Monitoring Results

Page 1 of 1

| Date | Location | OVA (ppm) | OVM (ppm) | NH ₃ (ppm) | H ₂ S (ppm) |
|---------|--------------------------------|--------------|--------------|--------------------------|---------------------------|
| 1/04/93 | Basin F Tank Farm - North Tank | 9.6 | 0.0 | 0.0 | 0.100 |
| | Basin F Tank Farm - East Tank | 11.8 | 2.5 | 0.0 | 0.170 |
| | Basin F Tank Farm - South Tank | 6.0 | 0.0 | 0.0 | 0.130 |
| 3/23/93 | Basin F Tank Farm - North Tank | 0.6 | 0.0 | Trace | 0.018 |
| | Basin F Tank Farm - East Tank | 6.0 | 0.0 | Trace | 0.020 |
| | Basin F Tank Farm - South Tank | 11.0 | 0.5 | 2.0 | 0.016 |
| 7/2/93 | Basin F Tank Farm - North Tank | 12.2 | 7.6 | 28.0 | 0.000 |
| | Basin F Tank Farm - East Tank | 11.4 | 0.0 | 9.0 | 0.000 |
| | Basin F Tank Farm - South Tank | 9.1 | 0.0 | 12.0 | 0.000 |
| 9/8/93 | Basin F Tank Farm - North Tank | 18.0 | 1.4 | 5.0 | 0.30 |
| | Basin F Tank Farm - East Tank | 12.0 | 5.4 | 4.0 | 0.36 |
| | Basin F Tank Farm - South Tank | 22.0 | 12.0 | 19.0 | 0.55 |

OVA - Organic Vapor Analyzer
 OVM - Organic Vapor Meter
 NH₃ - Ammonia
 H₂S - Hydrogen sulfide
 ppm - Parts per million

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A reading of 0.55 ppm was measured as the highest H₂S concentration and was observed on September 8, 1993 at the south tank vent.

Results of the tank farm vent VOC sample analyses are presented in Table 4.9-7. Dimethyldisulfide was found to have the highest average and maximum 24 concentrations, 410.8 µg/m³ and 990.0 µg/m³, respectively was obtained at the east tank.

4.10 THE CAQMMP SQI SUPPORT PROGRAM

4.10.1 General

One of the major remedial activities at RMA during the FY93 period was the testing and operational implementation of the SQI facility to dispose of Basin F liquids held in three storage tanks and Pond A. The CAQMMP provided intensive air monitoring support to this activity. The air program consisted of the special use of four existing sampler locations from the ongoing CAQMMP network. These were located upwind and downwind from the SQI. Stations SQ1 and AQ3 were 1,000 feet north and 6,100 feet north-northeast of the SQI, respectively; and SQ2 and AQ6 were 500 feet and 12,000 feet south of the SQI, respectively (see Figure 4.10-1). As the wind flow over RMA is diurnally bimodal overall, these stations captured the prevailing winds and dispersion patterns that might influence pollutants emitted by the SQI facility.

Ambient air sampling was conducted every third day (twice the frequency of the routine CAQMMP sampling program), and every day during a special test program period (June 10 through June 12, 1993). The third-day sampling program began on April 16, 1993 and continued until July 12, 1993. There was a 9-day break in the sampling effort that occurred in early June. A total of 32 special SQI sampling events were conducted. In addition, the full CAQMMP network was used for sampling every sixth day. Finally, special high-event monitoring was conducted for the SQI assessment (see Section 4.6.4). Other features of the program included fast laboratory turnaround and prompt data reporting to PMRMA.

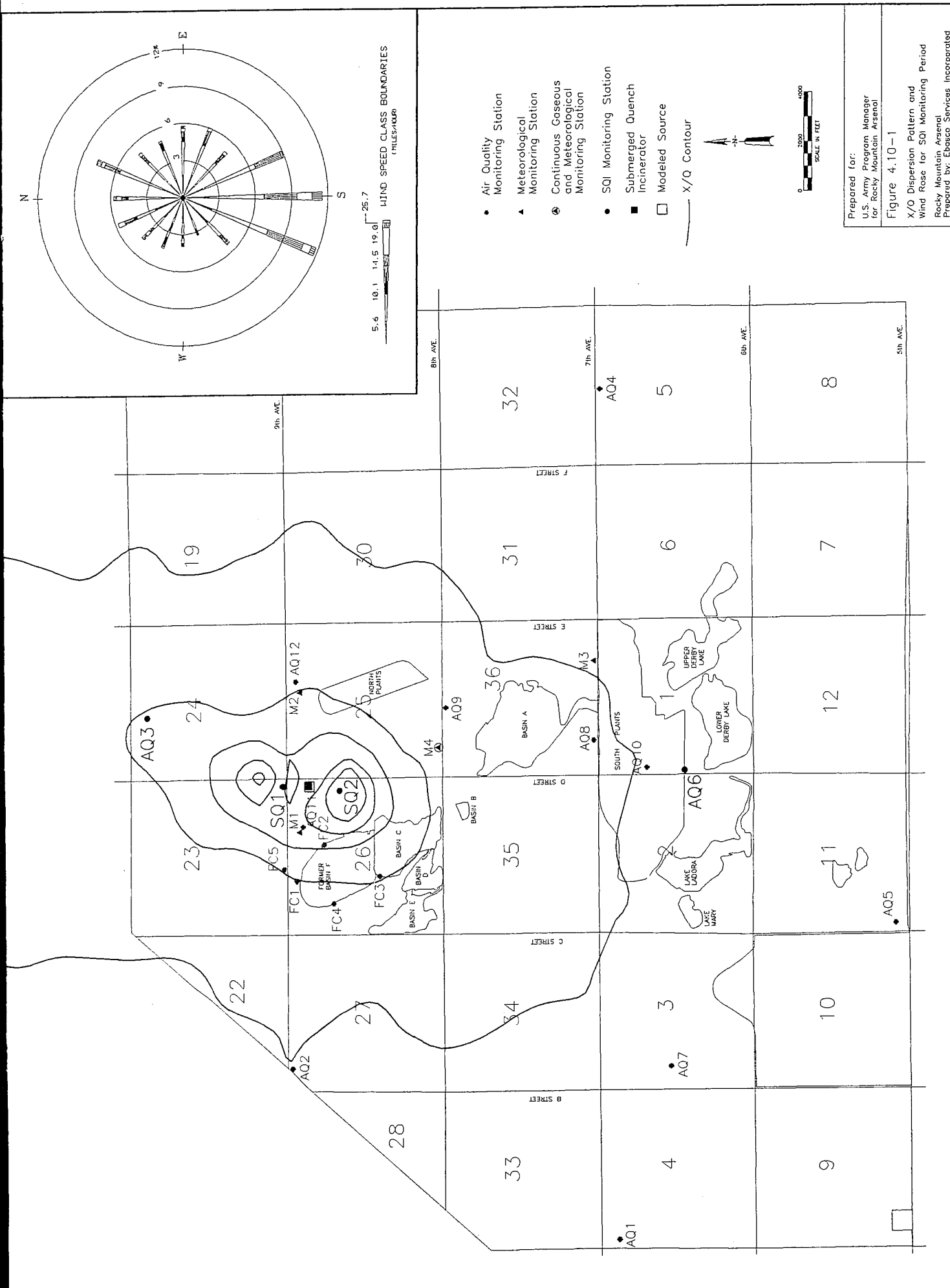
Table 4.9-7 Summary of FY93 Tank Farm Volatile Organic Compound Concentrations (µg/m³) Page 1 of 1

| | 111TCE | 112TCE | 11DCLE | 12DCLE | 12DCE | BCHPD | C6H6 | CCL4 | CH2CL2 | CHCL3 |
|------------|--------|--------|--------|--------|-------|-------|-------|-------|--------|-------|
| East Tank | | | | | | | | | | |
| Average | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | < 2.4 | 3.9 | < 3.5 | 25.7 | 6.8 |
| Maximum | 8.9 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 4.9 | 6.7 | < 3.5 | 88.1 | 19.0 |
| North Tank | | | | | | | | | | |
| Average | < 4.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | < 2.4 | < 3.1 | < 3.5 | < 15.0 | < 2.4 |
| Maximum | 7.1 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | < 2.4 | < 3.1 | < 3.5 | < 15.0 | 3.2 |
| South Tank | | | | | | | | | | |
| Average | 6.5 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 2.7 | 4.7 | < 3.5 | 49.1 | 23.1 |
| Maximum | 19.1 | < 7.1 | < 2.4 | < 2.4 | < 2.4 | 7.2 | 8.0 | < 3.5 | 102.0 | 46.0 |

| | CLC6H5 | DBCP | DCPD | DMDS | ETC6H5 | MEC6H5 | MIBK | TCLEE | TRCLE | XYLENES |
|------------|--------|--------|--------|-------|--------|--------|--------|-------|-------|---------|
| East Tank | | | | | | | | | | |
| Average | 6.3 | < 20.0 | 67.3 | 410.8 | < 6.3 | 38.5 | < 31.0 | 9.6 | < 2.6 | < 29.2 |
| Maximum | 19.0 | < 20.0 | 190.0 | 990.0 | < 6.3 | 92.0 | 90.0 | 25.0 | < 2.6 | < 29.2 |
| North Tank | | | | | | | | | | |
| Average | < 2.5 | < 20.0 | < 17.0 | 91.8 | < 6.3 | 4.2 | < 31.0 | < 2.5 | < 2.6 | < 29.2 |
| Maximum | 3.6 | < 20.0 | < 17.0 | 240.0 | < 6.3 | 9.9 | < 31.0 | < 2.5 | < 2.6 | < 29.2 |
| South Tank | | | | | | | | | | |
| Average | 9.2 | < 20.0 | 113.7 | 216.0 | < 6.3 | 65.5 | < 31.0 | 11.1 | < 2.6 | < 29.2 |
| Maximum | 17.0 | < 20.0 | 230.0 | 340.0 | < 6.3 | 110.0 | 44.0 | 19.0 | < 2.6 | < 29.2 |

| | | | | | | | | | | | |
|--------|---|--------------------------|--------|---|----------------------|--------|---|----------------------|---------|---|------------------------|
| 111TCE | - | 1,1,1-Trichloroethane | BCHPD | - | Bicycloheptadiene | CLC6H5 | - | Chlorobenzene | MEC6H5 | - | Toluene |
| 112TCE | - | 1,1,2-Trichloroethane | C6H6 | - | Benzene | DBCP | - | Dibromochloropropane | MIBK | - | Methyl Isobutyl Ketone |
| 11DCLE | - | 1,1-Dichloroethane | CCL4 | - | Carbon Tetrachloride | DCPD | - | Dicyclopentadiene | TCLEE | - | Tetrachloroethene |
| 12DCLE | - | 1,2-Dichloroethane | CH2CL2 | - | Methylene Chloride | DMDS | - | Dimethyldisulfide | TRCLE | - | Trichloroethene |
| 12DCE | - | trans-1,2-Dichloroethene | CHCL3 | - | Chloroform | ETC6H5 | - | Ethylbenzene | XYLENES | - | Total Xylenes |

< ### Indicates concentration less than the lower certified reporting level
µg/m³ Micrograms per cubic meter



Prepared for:
 U.S. Army Program Manager
 for Rocky Mountain Arsenal

Figure 4.10-1

X/Q Dispersion Pattern and
 Wind Rose for SQI Monitoring Period
 Rocky Mountain Arsenal
 Prepared by: Ebasco Services Incorporated

Parameters monitored included the following:

- TSP
- PM-10
- RMA target metals compounds
- RMA target VOC compounds
- RMA target OCP compounds
- Gaseous criteria pollutants (SO₂, CO, NO_x, O₃)
- Meteorological parameters (wind speed and direction, temperature, atmospheric stability, barometric pressure, humidity, and precipitation)

Results of the sampling were compared with long-term data at individual monitoring stations, pre-SQI data compiled under the CMP/CAQMMP during FY91 and FY92, and previous peak concentrations at RMA during Basin F remediation activities. Dispersion modeling was also conducted to assist in the assessment of results.

Figure 4.10-1 shows the wind rose for the composite period of the SQI special monitoring program and the atmospheric dispersion pattern associated with meteorological conditions for the same period. Although all wind directions were evident during this period, the results substantiate the placement of the four-station special network close to potential peak impacts, as well as along the axis of the primary dispersion. All of the CAQMMP monitoring stations used in the assessment are also shown in this figure.

4.10.2 Results

Table 4.10-1 provides composite results of the SQI special monitoring program for SQ1, SQ2, AQ3, and AQ6. In order to identify significant concentration levels for each of the sampled parameters, and for each site, the table format is structured to contain the following information:

Table 4.10-1 Submerged Quench Incinerator Monitoring Results

Page 1 of 8

| Station: AQ3 | | Pre-SQI Startup Results | | | | SQI 3rd Day Results 4/16/93 - 7/21/93 | |
|--------------------------|--|-------------------------|---------|---------|---------|--|--------------------|
| Analyte | FY88-FY92 24-hour Max. (All RMA Sites) | FY92 | | FY93 | | Long Term Average | 24-hour Maximum |
| | | AVG. | MAX. | AVG. | MAX. | | |
| 1,1,1-Trichloroethane | 72.54 | 1.105 | 2.460 | 1.716 | 4.854 | 1.344 | 2.530 |
| 1,1,2-Trichloroethane | 4.00 | < 0.170 | < 0.170 | < 0.170 | < 0.170 | < 0.170 | < 0.170 |
| 1,1-Dichloroethane | 0.27 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| 1,2-Dichloroethane | 3.80 | < 0.056 | 0.166 | < 0.056 | 0.301 | < 0.056 | 0.120 |
| trans-1,2-Dichloroethene | 0.06 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | 0.100 |
| Bicycloheptadiene | 39.46 | < 0.056 | < 0.056 | < 0.056 | 0.148 | < 0.056 | < 0.056 |
| Benzene | 44.29 | 1.234 | 3.570 | 2.943 | 8.758 | 1.032 | 2.130 |
| Carbon Tetrachloride | 9.76 | 0.511 | 1.001 | 0.699 | 1.046 | 0.747 | 1.343 |
| Methylene Chloride | 433.69 | 1.824 | 4.709 | 1.396 | 5.310 | 34.545 | 395. |
| Chloroform | 37.15 | 0.125 | 0.315 | 0.354 | 2.170 | 0.310 | 2.433 |
| Chlorobenzene | 2.53 | < 0.060 | < 0.060 | < 0.060 | 0.068 | < 0.060 | < 0.060 |
| Dibromochloropropane | 17.04 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.463 |
| Dicyclopentadiene | 29.12 | < 0.410 | < 0.410 | < 0.410 | < 0.410 | < 0.410 | < 0.410 |
| Dimethyldisulfide | 36.72 | < 0.220 | < 0.220 | < 0.220 | < 0.220 | < 0.220 | < 0.220 |
| Ethylbenzene | 13.04 | 0.295 | 1.042 | 0.689 | 2.330 | 0.187 | 0.530 |
| Toluene | 158.04 | 2.106 | 4.779 | 5.345 | 22.330 | 1.857 | 5.300 |
| Methyl Isobutyl Ketone | 1.13 | < 0.740 | < 0.740 | < 0.740 | < 0.740 | < 0.740 | < 0.740 |
| Tetrachloroethene | 24.02 | 0.255 | 1.067 | 0.534 | 2.450 | 0.199 | 0.560 |
| Trichloroethene | 40.43 | < 0.060 | < 0.060 | < 0.060 | 0.727 | < 0.060 | 0.091 |
| Total Xylenes | 49.91 | 1.340 | 5.320 | 3.250 | 10.600 | 0.856 | 2.600 |

Note: All results in micrograms per cubic meter

FY92 averages and maximums do not include estimates for values above the upper Certified Reporting Limit

< ### Indicates concentration less than certified reporting limit

SQI Submerged Quench Incinerator

Table 4.10-1 Submerged Quench Incinerator Monitoring Results

Page 2 of 8

| Station: AQ6 | | Pre-SQI Startup Results | | | | SQI 3rd Day Results 4/16/93 - 7/21/93 | |
|--------------------------|--|-------------------------|-----|---------|---------|--|--------------------|
| Analyte | FY88-FY92 24-hour Max. (All RMA Sites) | FY92 | | FY93 | | Long Term Average | 24-hour Maximum |
| | | AVG. | MAX | AVG. | MAX. | | |
| 1,1,1-Trichloroethane | 72.54 | N/A | N/A | 2.301 | 6.161 | 1.407 | 2.820 |
| 1,1,2-Trichloroethane | 4.00 | N/A | N/A | < 0.170 | < 0.170 | < 0.170 | < 0.170 |
| 1,1-Dichloroethane | 0.27 | N/A | N/A | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| 1,2-Dichloroethane | 3.80 | N/A | N/A | < 0.056 | 0.070 | < 0.056 | 0.130 |
| trans-1,2-Dichloroethene | 0.06 | N/A | N/A | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| Bicycloheptadiene | 39.46 | N/A | N/A | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| Benzene | 44.29 | N/A | N/A | 3.176 | 9.548 | 1.156 | 2.500 |
| Carbon Tetrachloride | 9.76 | N/A | N/A | 0.653 | 1.000 | 0.685 | 1.380 |
| Methylene Chloride | 433.69 | N/A | N/A | 1.912 | 11.510 | 36.405 | 441. |
| Chloroform | 37.15 | N/A | N/A | 0.309 | 0.777 | 0.305 | 1.898 |
| Chlorobenzene | 2.53 | N/A | N/A | < 0.060 | < 0.060 | < 0.060 | < 0.060 |
| Dibromochloropropane | 17.04 | N/A | N/A | < 0.463 | < 0.463 | < 0.463 | < 0.463 |
| Dicyclopentadiene | 29.12 | N/A | N/A | < 0.410 | < 0.410 | < 0.410 | < 0.410 |
| Dimethyldisulfide | 36.72 | N/A | N/A | < 0.220 | < 0.220 | < 0.220 | < 0.220 |
| Ethylbenzene | 13.04 | N/A | N/A | 0.757 | 2.190 | 0.270 | 0.730 |
| Toluene | 158.04 | N/A | N/A | 5.975 | 26.066 | 2.507 | 6.100 |
| Methyl Isobutyl Ketone | 1.13 | N/A | N/A | < 0.740 | < 0.740 | < 0.740 | < 0.740 |
| Tetrachloroethene | 24.02 | N/A | N/A | 0.646 | 3.610 | 0.286 | 0.642 |
| Trichloroethene | 40.43 | N/A | N/A | < 0.060 | 0.122 | < 0.060 | 0.100 |
| Total Xylenes | 49.91 | N/A | N/A | 3.588 | 10.210 | 1.325 | 3.700 |

Note: All results in micrograms per cubic meter
 N/A - No sampling results available for this time period

< ### Indicates concentration less than certified reporting limit
 SQI Submerged Quench Incinerator

Table 4.10-1 Submerged Quench Incinerator Monitoring Results

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| Station: SQ1 | | Pre-SQI Startup Results | | | | SQI 3rd Day Results 4/16/93 - 7/21/93 | |
|--------------------------|--|-------------------------|---------|---------|---------|--|--------------------|
| Analyte | FY88-FY92 24-hour Max. (All RMA Sites) | FY92 | | FY93 | | Long Term Average | 24-hour Maximum |
| | | AVG. | MAX. | AVG. | MAX. | | |
| 1,1,1-Trichloroethane | 72.54 | 1.230 | 3.469 | 2.310 | 13.130 | 1.250 | 2.830 |
| 1,1,2-Trichloroethane | 4.00 | < 0.170 | < 0.170 | < 0.170 | < 0.170 | < 0.170 | < 0.170 |
| 1,1-Dichloroethane | 0.27 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| 1,2-Dichloroethane | 3.80 | < 0.056 | 0.081 | < 0.056 | 0.058 | < 0.056 | 0.078 |
| trans-1,2-Dichloroethene | 0.06 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| Bicycloheptadiene | 39.46 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| Benzene | 44.29 | 1.334 | 3.944 | 4.173 | 33.830 | 1.007 | 2.090 |
| Carbon Tetrachloride | 9.76 | 0.487 | 0.871 | 0.776 | 1.717 | 0.692 | 2.050 |
| Methylene Chloride | 433.69 | 1.511 | 4.770 | 1.889 | 13.574 | 34.843 | 305. |
| Chloroform | 37.15 | 0.207 | 0.624 | 0.619 | 2.902 | 0.588 | 6.110 |
| Chlorobenzene | 2.53 | < 0.060 | < 0.060 | < 0.060 | < 0.060 | < 0.060 | < 0.060 |
| Dibromochloropropane | 17.04 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.463 |
| Dicyclopentadiene | 29.12 | < 0.410 | < 0.410 | < 0.410 | < 0.410 | < 0.410 | < 0.410 |
| Dimethyldisulfide | 36.72 | < 0.220 | < 0.220 | < 0.220 | < 0.220 | < 0.220 | < 0.220 |
| Ethylbenzene | 13.04 | 0.339 | 1.247 | 1.250 | 14.120 | 0.208 | 0.602 |
| Toluene | 158.04 | 2.279 | 5.090 | 8.561 | 87.560 | 1.905 | 5.200 |
| Methyl Isobutyl Ketone | 1.13 | < 0.740 | < 0.740 | < 0.740 | < 0.740 | < 0.740 | < 0.740 |
| Tetrachloroethene | 24.02 | 0.377 | 1.583 | 1.126 | 16.780 | 0.242 | 0.680 |
| Trichloroethene | 40.43 | < 0.056 | 0.086 | 0.065 | 0.657 | < 0.060 | 0.120 |
| Total Xylenes | 49.91 | 1.535 | 6.253 | 5.506 | 51.390 | 1.081 | 3.070 |

Note: All results in micrograms per cubic meter

FY 92 averages and maximums do not include estimates for values above the upper CRL

< ### Indicates concentration less than certified reporting limit

SQI Submerged Quench Incinerator

Table 4.10-1 Submerged Quench Incinerator Monitoring Results

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| Station: SQ2 | | Pre-SQI Startup Results | | | | SQI 3rd Day Results 4/16/93 - 7/21/93 | |
|--------------------------|--|-------------------------|---------|----------|---------|--|--------------------|
| Analyte | FY88-FY92 24-hour Max. (All RMA Sites) | FY92 | | FY93 | | Long Term Average | 24-hour Maximum |
| | | AVG. | MAX. | AVG. | MAX. | | |
| 1,1,1-Trichloroethane | 72.54 | 1.291 | 3.815 | 2.163 | 5.652 | 1.381 | 2.760 |
| 1,1,2-Trichloroethane | 4.00 | < 0.170 | < 0.170 | < 0.170 | < 0.170 | < 0.170 | < 0.170 |
| 1,1-Dichloroethane | 0.27 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| 1,2-Dichloroethane | 3.80 | < 0.056 | 0.785 | < 0.056 | < 0.056 | < 0.056 | 0.094 |
| trans-1,2-Dichloroethene | 0.06 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | 0.200 |
| Bicycloheptadiene | 39.46 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 | < 0.056 |
| Benzene | 44.29 | 1.359 | 3.734 | 3.544 | 11.304 | 1.181 | 2.500 |
| Carbon Tetrachloride | 9.76 | 0.518 | 1.104 | 0.711 | 2.100 | 0.757 | 1.760 |
| Methylene Chloride | 433.69 | 2.087 | 6.364 | 1.228 | 3.540 | 27.620 | 255. |
| Chloroform | 37.15 | 0.225 | 0.961 | 0.539 | 1.670 | 0.602 | 3.650 |
| Chlorobenzene | 2.53 | < 0.060 | < 0.060 | < 0.060 | < 0.060 | < 0.060 | < 0.060 |
| Dibromochloropropane | 17.04 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.463 | < 0.463 |
| Dicyclopentadiene | 29.12 | < 0.410 | < 0.410 | < 0.410 | < 0.410 | < 0.410 | < 0.410 |
| Dimethyldisulfide | 36.72 | < 0.220 | < 0.220 | < 0.0220 | < 0.220 | < 0.220 | < 0.220 |
| Ethylbenzene | 13.04 | 0.448 | 2.156 | 0.858 | 4.230 | 0.261 | 1.040 |
| Toluene | 158.04 | 2.591 | 6.186 | 6.186 | 29.800 | 2.254 | 6.600 |
| Methyl Isobutyl Ketone | 1.13 | < 0.740 | < 0.740 | < 0.740 | < 0.740 | < 0.740 | < 0.740 |
| Tetrachloroethene | 24.02 | 0.412 | 2.070 | 0.738 | 3.370 | 0.268 | 0.590 |
| Trichloroethene | 40.43 | < 0.060 | 0.050 | < 0.060 | 0.180 | < 0.060 | 0.092 |
| Total Xylenes | 49.91 | 1.957 | 10.814 | 3.995 | 19.810 | 1.182 | 3.200 |

Note: All results in micrograms per cubic meter

FY 92 averages and maximums do not include estimates for values above the upper CRL

< ### Indicates concentration less than certified reporting limit

SQI Submerged Quench Incinerator

Table 4.10-1 Submerged Quench Incinerator Monitoring Results

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| Station: AQ3 | | Pre-SQI Startup Results | | | | SQI 3rd Day Results 4/16/93 - 7/21/93 | |
|--------------|--|-------------------------|----------|----------|----------|--|--------------------|
| Analyte | FY88-FY92 24-hour Max. (All RMA Sites) | FY92 | | FY93 | | Long Term Average | 24-hour Maximum |
| | | AVG. | MAX. | AVG. | MAX. | | |
| Aldrin | 2.8290 | 0.0011 | 0.0011 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| Chlordane | 0.0360 | 0.0008 | 0.0012 | < 0.0003 | < 0.0003 | < 0.0003 | 0.0007 |
| Dieldrin | 2.2960 | 0.0006 | 0.0014 | < 0.0003 | < 0.0003 | < 0.0003 | 0.0012 |
| Endrin | 1.0954 | 0.0018 | 0.0018 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| Isodrin | 0.9450 | 0.0021 | 0.0021 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| PPDDE | 0.0390 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| PPDDT | 0.0585 | < 0.0003 | < 0.0003 | < 0.0003 | 0.0005 | < 0.0003 | < 0.0003 |
| Cadmium | 0.0281 | < 0.0025 | < 0.0025 | < 0.0025 | 0.0062 | < 0.0025 | < 0.0025 |
| Chromium | 0.2858 | < 0.0123 | 0.0158 | < 0.0123 | < 0.0123 | < 0.0123 | < 0.0123 |
| Copper | 1.4980 | 0.0872 | 0.1950 | 0.0770 | 0.1290 | 0.0858 | 0.1400 |
| Lead | 0.0984 | < 0.0245 | 0.0347 | < 0.0245 | 0.0379 | < 0.0245 | < 0.0245 |
| Zinc | 3.3576 | 0.0240 | 0.0500 | 0.0289 | 0.0853 | 0.0196 | 0.3660 |
| Arsenic | 0.0136 | < 0.0009 | 0.0013 | < 0.0009 | 0.0016 | < 0.0009 | 0.0011 |
| Mercury | 7.3000 | < 0.2315 | < 0.2315 | < 0.2315 | < 0.2315 | < 0.2315 | 0.2940 |
| TSP | 902 | 37 | 83 | 30 | 68 | 31 | 69 |
| PM-10 | 168 | 22 | 41 | 22 | 41 | 17 | 26 |

Note: All results in micrograms per cubic meter

< ### Indicates concentration less than certified reporting limit

SQI Submerged Quench Incinerator

TSP Total Suspended Particulates

PM-10 Particulate Matter less than 10 Micrometers

Table 4.10-1 Submerged Quench Incinerator Monitoring Results

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| Station: AQ6 | | Pre-SQI Startup Results | | | | SQI 3rd Day Results 4/16/93 - 7/21/93 | |
|--------------|--|-------------------------|------|----------|----------|--|--------------------|
| Analyte | FY88-FY92 24-hour Max. (All RMA Sites) | FY92 | | FY93 | | Long Term Average | 24-hour Maximum |
| | | AVG. | MAX. | AVG. | MAX. | | |
| Aldrin | 2.8290 | N/A | N/A | < 0.0003 | < 0.0003 | < 0.0003 | 0.0009 |
| Chlordane | 0.0360 | N/A | N/A | < 0.0003 | 0.0018 | 0.0007 | 0.0031 |
| Dieldrin | 2.2960 | N/A | N/A | 0.0003 | 0.0019 | 0.0009 | 0.0041 |
| Endrin | 1.0954 | N/A | N/A | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| Isodrin | 0.9450 | N/A | N/A | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| PPDDE | 0.0390 | N/A | N/A | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| PPDDT | 0.0585 | N/A | N/A | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| Cadmium | 0.0281 | N/A | N/A | < 0.0025 | 0.0073 | < 0.0025 | < 0.0025 |
| Chromium | 0.2858 | N/A | N/A | < 0.0123 | < 0.0123 | < 0.0123 | < 0.123 |
| Copper | 1.4980 | N/A | N/A | 0.0806 | 0.1710 | 0.0785 | 0.1550 |
| Lead | 0.0984 | N/A | N/A | < 0.0245 | 0.0478 | < 0.0245 | < 0.0245 |
| Zinc | 3.3576 | N/A | N/A | 0.0305 | 0.1000 | 0.0189 | 0.0346 |
| Arsenic | 0.0136 | N/A | N/A | < 0.0009 | 0.0016 | 0.0009 | 0.0139 |
| Mercury | 7.3000 | N/A | N/A | < 0.2315 | 0.2790 | < 0.2315 | 0.394 |
| TSP | 902 | 36 | 83 | 30 | 65 | 28 | 82 |
| PM-10 | 168 | N/A | N/A | 19 | 39 | 15 | 27 |

Note: All results in micrograms per cubic meter
N/A - No sampling results available for this time period

< ### Indicates concentration less than certified reporting limit

SQI Submerged Quench Incinerator

TSP Total Suspended Particulates

PM-10 Particulate Matter less than 10 Micrometers

Table 4.10-1 Submerged Quench Incinerator Monitoring Results

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| Station: SQ1 | | Pre-SQI Startup Results | | | | SQI 3rd Day Results 4/16/93 - 7/21/93 | |
|--------------|--|-------------------------|----------|----------|----------|--|--------------------|
| Analyte | FY88-FY92 24-hour Max. (All RMA Sites) | FY92 | | FY93 | | Long Term Average | 24-hour Maximum |
| | | AVG. | MAX. | AVG. | MAX. | | |
| Aldrin | 2.8290 | 0.0010 | 0.0021 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| Chlordane | 0.0360 | 0.0016 | 0.0067 | < 0.0003 | < 0.0003 | 0.0003 | 0.0014 |
| Dieldrin | 2.2960 | 0.0015 | 0.0054 | < 0.0003 | 0.0009 | 0.0009 | 0.0051 |
| Endrin | 1.0954 | 0.0005 | 0.0005 | < 0.0003 | < 0.0003 | < 0.0003 | 0.0004 |
| Isodrin | 0.9450 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| PPDDE | 0.0390 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| PPDDT | 0.0585 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| Cadmium | 0.0281 | < 0.0025 | 0.0030 | < 0.0025 | 0.0087 | < 0.0025 | < 0.0025 |
| Chromium | 0.2858 | < 0.0123 | 0.0196 | < 0.0123 | < 0.0123 | < 0.123 | < 0.0123 |
| Copper | 1.4980 | 0.1228 | 0.3166 | 0.0517 | 0.0948 | 0.0105 | 0.1190 |
| Lead | 0.0984 | < 0.0245 | 0.0346 | < 0.0245 | 0.0497 | < 0.0245 | < 0.0245 |
| Zinc | 3.3576 | 0.0247 | 0.0650 | 0.0298 | 0.1000 | 0.0209 | 0.0873 |
| Arsenic | 0.0136 | < 0.0009 | 0.0024 | < 0.0009 | 0.0019 | < 0.0009 | 0.0081 |
| Mercury | 7.3000 | N/A | N/A | < 0.2315 | < 0.2315 | < 0.2315 | 0.2620 |
| TSP | 902 | 46 | 129 | 34 | 97 | 32 | 76 |
| PM-10 | 168 | 26 | 56 | 25 | 50 | 18 | 30 |

Note: All results in micrograms per cubic meter
N/A - No sampling results available for this time period

< ### Indicates concentration less than certified reporting limit

SQI Submerged Quench Incinerator

TSP Total Suspended Particulates

PM-10 Particulate Matter less than 10 Micrometers

Table 4.10-1 Submerged Quench Incinerator Monitoring Results

Page 8 of 8

Station: SQ2

Pre-SQI Startup Results

SQI 3rd Day Results
4/16/93 - 7/21/93

| Analyte | FY88-FY92 24-hour Max. (All RMA Sites) | FY92 | | FY93 | | Long Term Average | 24-hour Maximum |
|-----------|--|----------|----------|----------|----------|----------------------|--------------------|
| | | AVG. | MAX. | AVG. | MAX. | | |
| Aldrin | 2.8290 | 0.0007 | 0.0009 | < 0.0003 | < 0.0003 | < 0.0003 | 0.0006 |
| Chlordane | 0.0360 | 0.0021 | 0.0084 | < 0.0003 | < 0.0003 | 0.0005 | 0.0022 |
| Dieldrin | 2.2960 | 0.0014 | 0.0056 | 0.0003 | 0.0022 | 0.0013 | 0.0070 |
| Endrin | 1.0954 | 0.0005 | 0.0006 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| Isodrin | 0.9450 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| PPDDE | 0.0390 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| PPDDT | 0.0585 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 | < 0.0003 |
| Cadmium | 0.0281 | < 0.0025 | 0.0029 | < 0.0025 | 0.0087 | < 0.0025 | 0.0110 |
| Chromium | 0.2858 | < 0.0123 | 0.0347 | < 0.0123 | < 0.0123 | < 0.0123 | < 0.0123 |
| Copper | 1.4980 | 0.0811 | 0.1442 | 0.0670 | 0.1120 | 0.0873 | 0.1320 |
| Lead | 0.0984 | < 0.0245 | 0.0324 | < 0.0245 | 0.0517 | < 0.0245 | 0.0518 |
| Zinc | 3.3576 | 0.0263 | 0.0618 | 0.0294 | 0.1010 | 0.0197 | 0.0398 |
| Arsenic | 0.0136 | < 0.0009 | 0.0010 | < 0.0009 | 0.0014 | < 0.0009 | 0.0047 |
| Mercury | 7.3000 | N/A | N/A | < 0.2315 | < 0.2315 | < 0.2315 | 0.2220 |
| TSP | 902 | 44 | 120 | 27 | 80 | 32 | 79 |
| PM-10 | 168 | 27 | 50 | 20 | 46 | 17 | 31 |

Note: All results in micrograms per cubic meter
N/A - No sampling results available for this time period

< ### Indicates concentration less than certified reporting limit

SQI Submerged Quench Incinerator

TSP Total Suspended Particulates

PM-10 Particulate Matter less than 10 Micrometers

- The maximum short-term (24-hour) concentration at RMA for each parameter or analyte measured during the CMP and CAQMMP (occurred mostly during the Basin F remediation period)
- The average long-term annual concentration at each station prior to startup (for FY92 and FY93)
- The maximum short-term concentration measured at the four monitoring stations sampling the SQI (SQ1, SQ2, AQ3, or AQ6) prior to SQI startup
- The average and maximum concentration at each station during every third-day SQI sampling program

This format allows for comparisons of SQI-sampled parameters and analytes with the highest short-term concentrations experienced at RMA during active remediation, as well as more recent baseline information prior to the SQI startup.

These data indicate that all TSP, PM-10, VOC, OCP, metals, arsenic, and mercury concentrations were at or below historical RMA levels with the exception of methylene chloride (discussed in Section 4.6.5.2 separately). With respect to PM-10 and TSP, there were no violations of short-term or longer-term standards. Maximum 24-hour concentrations of OCPs were smaller by a factor of at least 100 than the Basin F remediation levels. It appeared that dieldrin concentrations were slightly higher than recent pre-SQI startup levels; however, this could be a seasonal phenomenon. All metals levels were typical of RMA and urban background levels. There were several measurements of mercury during the SQI monitoring program; however, comparable measurements were sampled prior to SQI startup as reflected in Table 4.10-1.

VOCs were well below peak remediation concentrations (for those analytes identified as potential Basin F emissions) and were mostly typical of current RMA and regional levels. Carbon tetrachloride and chloroform levels were above recent FY93 concentrations, but well below peak RMA levels, and were typical of concentrations that have been experienced on occasion during the post-Basin F remediation monitoring programs.

There were no anomalous levels of gaseous criteria pollutants (SO₂, NO_x, CO, and O₃) measured at the CAQMMP monitoring site (approximately 1 mile south-southeast from the SQI) during the special SQI monitoring program.

4.11 THE SOIL VAPOR EXTRACTION SYSTEM SAMPLING PROGRAM

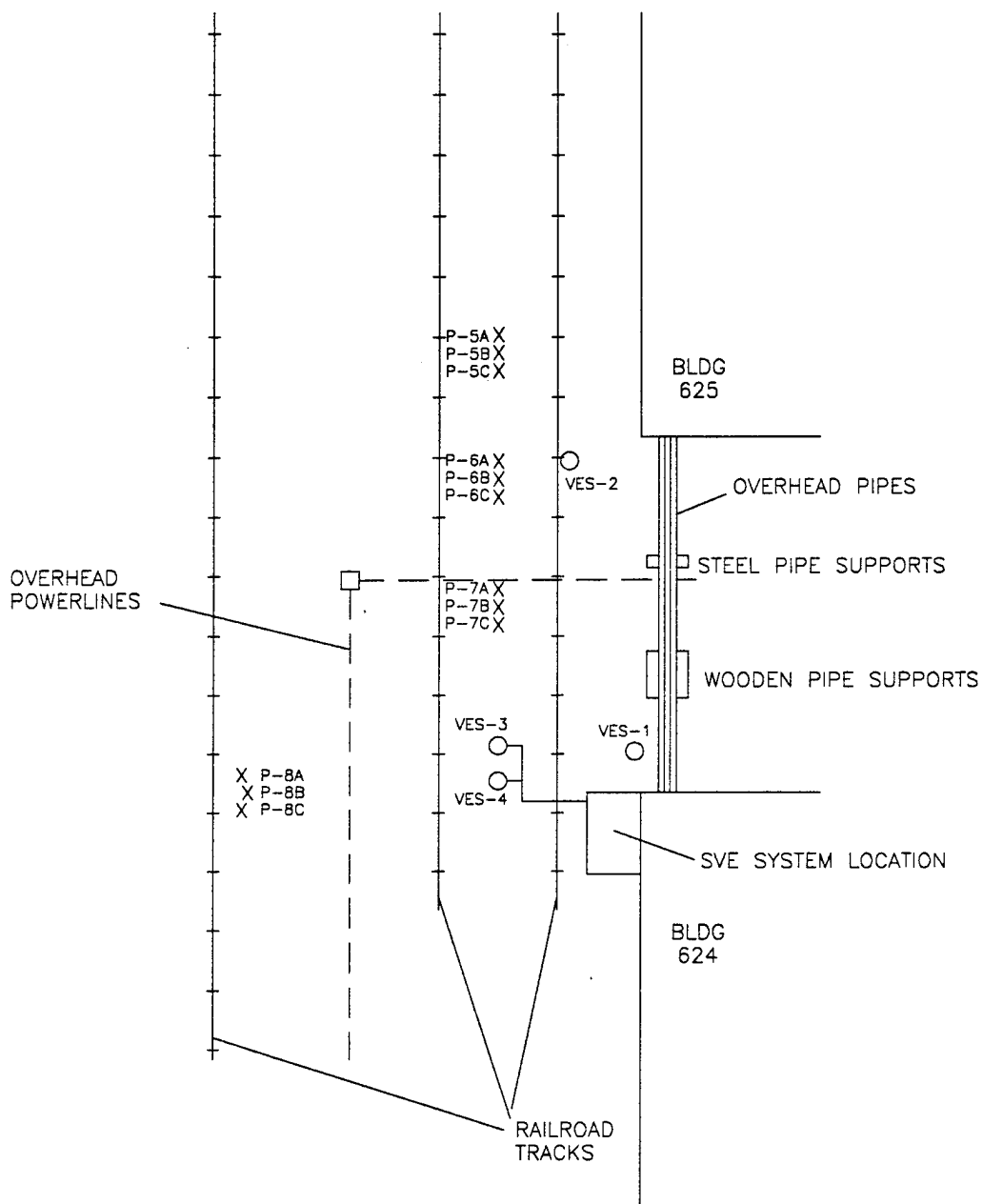
4.11.1 Overview

During FY93, PMRMA requested a special air monitoring task to be conducted in the area of the Motor Pool. This section describes the sampling procedures and results of the operational testing and air monitoring of the SVE system at the Motor Pool area from September 29 to October 1, 1993. This work was performed in accordance with the letter work plan submitted to PMRMA on September 10, 1993. The objective of this program was to furnish additional data to PMRMA for assessing the effectiveness of the remediation conducted under the Motor Pool Area IRA Soil Vapor Extraction Pilot Study in 1991.

A series of operational testing and three-phase air monitoring tasks were performed using the existing SVE system and soil gas monitoring wells at the Motor Pool area (Figure 4.11-1). Testing was conducted to verify the operability of the SVE system prior to the required 48-hour test run. During Phase I sampling, air samples were collected at the 12 soil gas monitoring wells and analyzed with an on-site Photovac 10S70 GC. A passivated canister sample (SUMMA) was then collected at the well site with the highest overall concentration of trichloroethene (TRCLE) and tetrachloroethene (TCLEE) and sent for analysis to the off-site laboratory. Phase II sampling was conducted during the 48-hour test run of the SVE. Canister samples were collected at the two vapor extraction well sites (VES-3 & VES-4) at hours 00, 16, 32, and 48 of the test period. Samples were also collected and analyzed with the on-site GC at hours 00 and 48. Phase II sampling replicated the sampling of the 12 soil gas monitoring wells that occurred during Phase I.

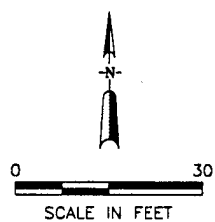
4.11.2 Operational Testing

The SVE system was inspected on September 9, 1993 prior to operational testing and deemed to be in adequate condition. The valves to the vapor extraction wells, VES-3 and VES-4, were



LEGEND

- VES-1 ○ Soil Vapor Extraction Well
P-8C X Soil Gas Monitoring Well
SVE Soil Vapor Extraction
BLDG Building



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 4.11-1

Motor Pool Area Soil Vapor
Extraction System Well Locations

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

then closed to isolate subsurface vapors from the vacuum generated during testing. Valve V-1, located between the separator tank and the blower, was opened to provide an ambient air source for the system (Figure 4.11-2). The SVE system was turned on and allowed to operate for 10 minutes. Operational pressure/vacuum, temperature, and flow readings were taken. Based upon these readings (Table 4.11-1) and a visual inspection, the SVE was judged functional and capable of performing the full-scale 48-hour test run.

4.11.3 Phase I Sampling

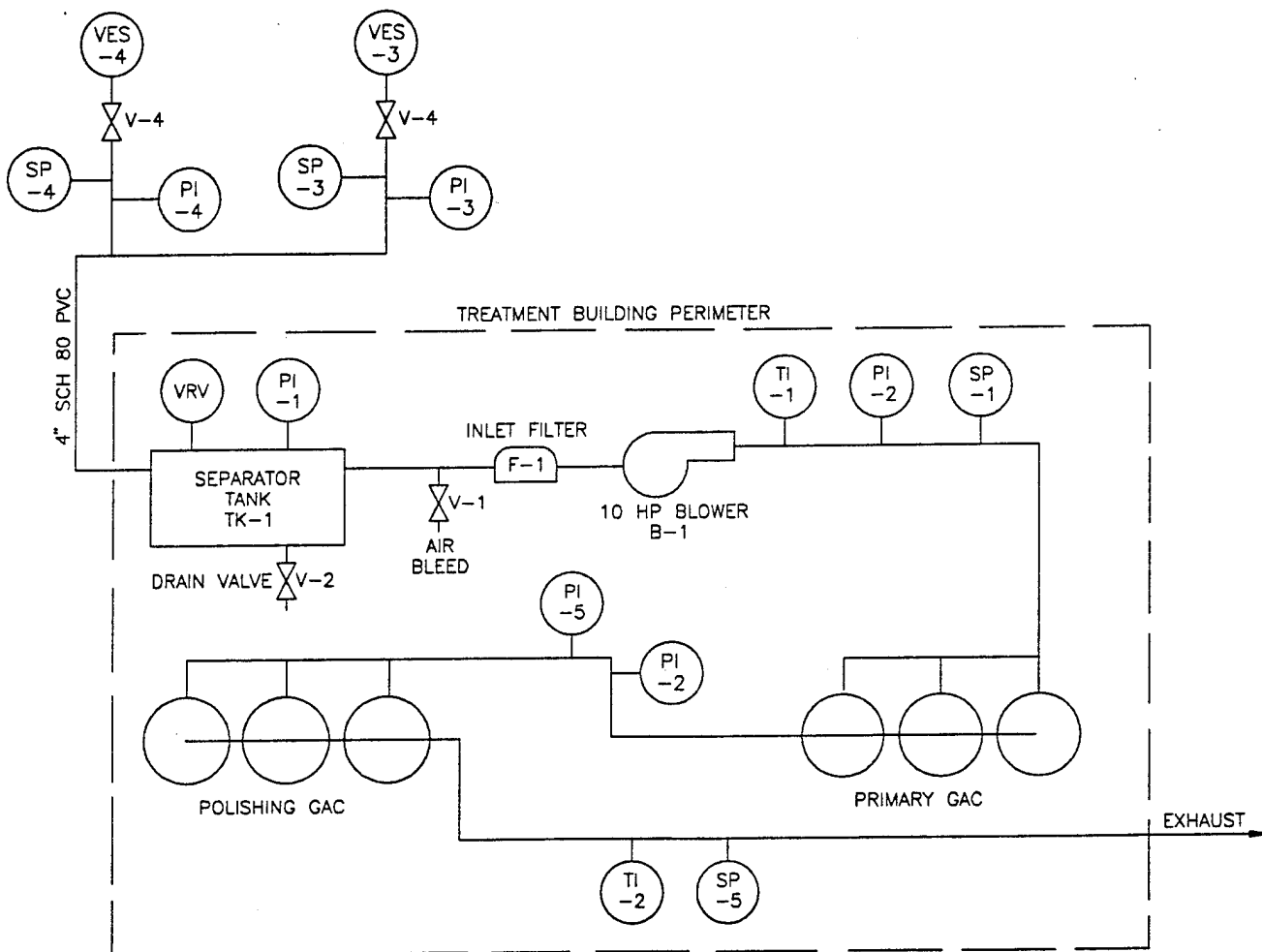
Prior to startup of the 48-hour test run, soil gas samples were collected at the 12 soil gas monitoring wells (Figure 4.11-1). The Photovac GC was prepared and calibrated according to the procedures described in the work plan. The soil gas monitoring well samples were then collected in the following manner:

- 1) A short length of polyethylene tubing was connected to the monitoring well tubing using a compression fitting. A digital manometer was connected in line using a T-connector and a ball valve to control flow to the instrument. Then, an initial pressure/vacuum reading was recorded.
- 2) A hand-operated pump was attached to the 1/4-inch polyethylene tubing. Prior to collecting the sample, a predetermined volume of air was purged out of each well to ensure that the sample was not diluted by ambient air (1.5 liters for "A" wells, 2.0 liters for "B" wells, and 3.0 liters for "C" wells).
- 3) After allowing the system to return to ambient pressure, a 1.0-milliliter aliquot of soil gas was collected using a syringe inserted through a septum in the sample tube upstream of the vacuum pump.
- 4) The 1.0 milliliter soil gas aliquot sample was immediately injected into the GC for analysis for the target compounds: TRCLE and TCLEE.
- 5) The sampling apparatus was purged with hydrocarbon-free air between well tests.

Table 4.11-2 summarizes the results of this sampling. Based upon these results, a SUMMA canister sample was collected at well P-7A according to the procedures described in the work plan. This canister sample had TRCLE and TCLEE concentrations of 3.070 ppm and 0.007 ppm, respectively.

LEGEND

VES - SOIL VAPOR EXTRACTION WELL
 PI - PRESSURE (VACUUM) INDICATOR
 TI - TEMPERATURE INDICATOR
 SP - SAMPLE PORT
 V - VALVE
 VRV - VACUUM RELIEF VALVE
 GAC - GRANULAR ACTIVATED CARBON (VAPOR PHASE)



Prepared for:
 U.S. Army Program Manager
 for Rocky Mountain Arsenal

Figure 4.11-2

Soil Vapor Extraction System
 Process Flow Diagram

Rocky Mountain Arsenal
 Prepared by: Ebasco Services Incorporated

Table 4.11-1 Soil Vapor Extraction System Operational Testing Results

Page 1 of 1

| | Before Startup | After Startup |
|---|-----------------------------|------------------------------|
| <u>Pressure readings</u> | | |
| Separator Tank (PI-1) | 1.5 inches H ₂ O | 9.5 inches H ₂ O |
| Blower Exhaust (PI-2) | 0.0 inches H ₂ O | 14.0 inches H ₂ O |
| Post-Primary Granular Activated Carbon (PI-5) | 0.0 inches H ₂ O | 8.5 inches H ₂ O |
| <u>Temperature readings</u> | | |
| Blower Exhaust (TI-1) | 11.0 °C | 41.0 °C |
| Post Granular Activated Carbon (TI-2) | 14.0 °C | 15.0 °C |
| <u>Gas flow rate</u> | | |
| After Blower (SP-1) | 0 feet per minute | 6,000 feet per minute |

°C - Degrees Celcius

Table 4.11-2 Phase I On-site GC Results for the Soil Gas Monitoring Wells

Page 1 of 1

| Soil Gas Monitoring Well | Trichloroethene (ppm) | Tetrachloroethene (ppm) |
|--------------------------|-----------------------|-------------------------|
| P5-A | 1.400 | 0.020 |
| P5-B | 0.680 | <0.015 |
| P5-C | 0.600 | <0.015 |
| P6-A | 3.600 | 0.030 |
| P6-B | 0.790 | 0.050 |
| P6-C | 1.300 | 0.020 |
| P7-A | 5.400 | 0.020 |
| P7-B | 0.670 | <0.015 |
| P7-C | 0.390 | 0.060 |
| P8-A | 0.300 | <0.015 |
| P8-B | 0.070 | <0.015 |
| P8-C | 0.360 | <0.015 |

< ### Indicates concentration less than the instrument detection limit
ppm Parts per million

4.11.4 Phase II Sampling

During the 48-hour test run of the SVE, air samples were collected at the two soil vapor extraction well sites (VES-3 & VES-4) at hours 00, 16, 32, and 48. These samples were collected as follows:

- 1) The air flow from the extraction well was temporarily interrupted by closing the valve (V-3 & VES-4) at the extraction well (Figure 4.11-2).
- 2) A short length of polyethylene tubing was connected to the available sample port using a compression fitting.
- 3) The "T" sampling apparatus used in Phase I sampling was again utilized. The sample port was opened and a vacuum applied with the hand pump.
- 4) A syringe sample was collected and analyzed with the Photovac GC, as described previously. Also, a canister sample was collected by closing the ball valve to the pump and opening the canister's needle valve. At hours 16 and 32, only canister samples were collected.
- 5) Upon completion of sampling, flow was restored from the vapor extraction well.

Table 4.11-3 lists the results of the Phase II canister sampling. Table 4.11-4 details the concentrations detected by the on-site GC.

4.11.5 Phase III Sampling

After completion of the 48-hour test run, the system was allowed to equilibrate. Within several minutes, pressure/vacuum readings at the 12 soil gas monitoring wells had reached pre-testing levels. Soil gas samples were then collected at each well. These samples were collected in the same manner as in Phase I. Table 4.11-5 lists the results of this sampling. Based upon these results, a canister sample was collected at well P-7B. TRCLE and TCLEE concentrations were 2.390 and < 0.010 ppm, respectively, for this sample.

Table 4.11-3 Phase II SUMMA Canister Results for the Vapor Extraction Wells during the
48-Hour Test Run Page 1 of 1

| Vapor Extraction Well | Time (hours into test run) | Trichloroethene (ppm) | Tetrachloroethene (ppm)* |
|-----------------------|----------------------------|-----------------------|--------------------------|
| VES-3 | 0.5 | 2.410 | 0.005 |
| VES-3 | 16.0 | 4.150 | 0.005 |
| VES-3 | 32.0 | 4.410 | <0.010 |
| VES-3 | 47.5 | 3.940 | <0.010 |
| VES-4 | 0.5 | 0.945 | 0.003 |
| VES-4 | 16.0 | 1.800 | <0.010 |
| VES-4 | 32.0 | 0.752 | <0.010 |
| VES-4 | 47.5 | 0.703 | <0.010 |

* Detection limit varies according to required sample dilution
 < ### indicates concentration less than lower certified reporting limit
 ppm Parts per million

Table 4.11-4 Phase II On-site GC Results for the Vapor Extraction Wells during the 48-Hour Test Run

Page 1 of 1

| Vapor Extraction Well | Time (hours into test run) | Trichloroethene (ppm) | Tetrachloroethene (ppm) |
|-----------------------|----------------------------|-----------------------|-------------------------|
| VES-3 | 0.5 | 4.300 | <0.015 |
| VES-3 | 47.5 | 1.400 | <0.015 |
| VES-4 | 0.5 | 2.000 | <0.015 |
| VES-4 | 47.5 | 0.380 | <0.015 |

< ### Indicates concentration less than the lower certified reporting limit
ppm Parts per million

Table 4.11-5 Phase III On-site GC Results for the Soil Gas Monitoring WellPage 1 of 1

| Soil Gas Monitoring Well | Trichloroethene (ppm) | Tetrachloroethene (ppm) |
|--------------------------|-----------------------|-------------------------|
| P5-A | 0.160 | <0.015 |
| P5-B | 0.180 | <0.015 |
| P5-C | 0.230 | <0.015 |
| P6-A | 0.230 | <0.015 |
| P6-B | 0.240 | <0.015 |
| P6-C | 0.090 | <0.015 |
| P7-A | 0.430 | <0.015 |
| P7-B | 0.760 | <0.015 |
| P7-C | 0.320 | <0.015 |
| P8-A | <0.015 | <0.015 |
| P8-B | 0.020 | <0.015 |
| P8-C | 0.310 | <0.015 |

< ### Indicates concentrations less than the lower certified reporting limit
 ppm Parts per million

4.11.6 Quality Assurance and Quality Control

Quality assurance and quality control procedures for the SVE sampling program consisted of collecting field duplicates and blank samples analyzed by the on-site GC and the off-site laboratory.

The on-site Photovac GC analyzed calibration standards, equipment blanks, instrument blanks, method blanks, and duplicate samples throughout the field program. Equipment blanks were taken with each sample set and involved the analysis of ambient air drawn through the decontaminated sampling apparatus when it was not connected to the well. Also, an aliquot of hydrocarbon-free air was drawn into the sample syringe and analyzed with the GC. Initial results indicated some minor carryover from the more concentrated samples; however, this was eliminated by flushing the sampling apparatus with hydrocarbon-free air between samples. Instrument blanks were performed daily and involved analysis with no injection. These illustrated no carryover contamination in the GC. The daily method blanks demonstrated the lack of background contamination of TRCLE and TCLEE in the ambient air. Table 4.11-6 highlights the satisfactory precision associated with the on-site GC sampling.

Table 4.11-7 lists the results of the duplicate canister sampling. These results demonstrate that the canister data were reliable. A canister sample of the hydrocarbon-free air used to purge the sampling apparatus between uses was also collected and analyzed. A small amount of TRCLE was detected (0.001 ppm) and no TCLEE was detected. Also, laboratory method blanks showed low levels of the TRCLE and methylene chloride. However, these quantities were insignificant when compared to the concentrations measured in the field samples (more than 100 times less). All data collected in the SVE air sampling program satisfied data quality objectives.

Non-target compound results for the canister samples are summarized in Table 4.11-8.

Table 4.11-6 On-site GC Duplicate Results

Page 1 of 1

| SITE | Trichloroethene (ppm) | Tetrachloroethene (ppm) |
|-----------------|-----------------------|-------------------------|
| P-7A | 5.400 | 0.020 |
| P-7A Duplicate | 3.200 | <0.015 |
| VES-3 | 4.300 | <0.015 |
| VES-3 Duplicate | 3.900 | <0.015 |
| P-7B | 0.760 | <0.015 |
| P-7B Duplicate | 0.670 | <0.015 |

< ### Indicates concentration less than the lower certified reporting limit
ppm Parts per million

Table 4.11-7 SUMMA Canister Duplicate Sample Results

Page 1 of 1

| SITE | Trichloroethene (ppm) | Tetrachloroethene (ppm) |
|-----------------|-----------------------|-------------------------|
| VES-3 | 5.400 | 0.020 |
| VES-3 Duplicate | 3.200 | < 0.015 |

< ### Indicates concentration less than the lower certified reporting limit
ppm Parts per million

Table 4.11-8 SUMMA Canister Results - Other Non-target Detections
(ppbv)

Page 1 of 3

P-7A (prior to 48 hour test run)

| | |
|--------------------------------|--------------------------------|
| Chloroform = 1.49 | Trans-1,2-Dichloroethene = 345 |
| Toluene = 1.24 | Benzene = 0.95 |
| Propylene = 1.65 | 1,1,1-Trichloroethane = 0.84 |
| 1,1,2 - Trichloroethane = 1.99 | |

VES-3 (hour 00 of test run)

| | |
|------------------------------|--------------------------------|
| Benzene = 0.93 | Carbon Tetrachloride = 1.74 |
| Chloroform = 1.89 | 1,1-Dichloroethylene = 1.59 |
| Ethylbenzene = 0.94 | m,p-Xylene = 0.78 |
| Toluene = 4.05 | 1,1,1-Trichloroethane = 3.01 |
| 1,1,2-Trichloroethane = 3.25 | Trans-1,2-Dichloroethene = 166 |
| m,p-Xylene = 0.78 | |

VES-4 (hour 00 of test run)

| | |
|------------------------------|---------------------------------|
| Benzene = 0.69 | Carbon Tetrachloride = 2.17 |
| Chloroform = 8.52 | 1,1-Dichloroethylene = 16.1 |
| Toluene = 26.6 | 1,1,1-Trichloroethane = 33.9 |
| 1,1,2-Trichloroethane = 1.62 | Trans-1,2-Dichloroethene = 33.5 |

VES-3 (hour 16 of test run)

| | |
|--------------------------------|------------------------------|
| Chloroform = 1.27 | Toluene = 0.73 |
| 1,1,1-Trichloroethane = 1.79 | 1,1,2-Trichloroethane = 5.15 |
| Trans-1,2-Dichloroethene = 157 | Benzene = 0.54 |

VES-4 (hour 16 of test run)

| | |
|------------------------------|-------------------------------|
| 1,4-Dichlorobenzene = 24.1 | 1,1-Dichloroethylene = 11.3 |
| Ethylbenzene = 15.5 | m,p-Xylene = 54.9 |
| Methylene Chloride = 588 | Toluene = 414 |
| 1,1,1-Trichloroethane = 26.1 | 1,2,4-Trimethylbenzene = 68.8 |
| Chlorodifluoromethane = 2080 | |

ppbv - Parts per billion volume

Table 4.11-8 SUMMA Canister Results - Other Non-target Detections
(ppbv)

Page 2 of 3

VES-3 (hour 32 of test run)

1,4-Dichlorobenzene = 27.2

m,p-Xylene = 64.8

o-Xylene = 30.7

Toluene = 493

1,3,5-Trimethylbenzene = 40.7

Hexane = 14.3

Ethylbenzene = 20.6

Methylene Chloride = 869

Styrene = 36.7

1,2,4-Trimethylbenzene = 75.9

Chlorodifluoromethane = 3580

Trans-1,2-Dichloroethene = 23.5

VES-4 (hour 32 of test run)

1,4-Dichlorobenzene = 29.2

m,p-Xylene = 69.5

o-Xylene = 33.3

Toluene = 473

1,2,4-Trimethylbenzene = 78.6

Chlorodifluoromethane = 743

Ethylbenzene = 21.6

Methylene Chloride = 616

Styrene = 39.3

1,1,1-Trichloroethane = 20.7

1,3,5-Trimethylbenzene = 21.4

VES-3 (hour 48 of test run)

1,4-Dichlorobenzene = 31.7

m,p-Xylene = 70.3

o-Xylene = 33.0

Toluene = 508

1,2,4-Trimethylbenzene = 75.7

Chlorodifluoromethane = 834

Hexane = 11.0

Ethylbenzene = 21.4

Methylene Chloride = 994

Styrene = 39.6

Trichlorofluoromethane = 12.0

1,3,5-Trimethylbenzene = 20.4

Pentane = 10.6

Trans-1,2-Dichloroethene = 20.4

Table 4.11-8 SUMMA Canister Results - Other Non-target Detections
(ppbv)

Page 3 of 3

VES-4 (hour 48 of test run)

1,4-Dichlorobenzene = 25.2

m,p-Xylene = 65.9

o-Xylene = 30.4

Toluene = 492

1,2,4-Trimethylbenzene = 77.2

Chlorodifluoromethane = 554

P7-B (after 48 hour test run)

1,4-Dichlorobenzene = 27.3

m,p-Xylene = 78.8

o-Xylene = 37.3

Toluene = 571

1,2,4-Trimethylbenzene = 80.8

Chlorodifluoromethane = 801

Ethylbenzene = 19.8

Methylene Chloride = 755

Styrene = 36.3

1,1,1-Trichloroethane = 21.3

1,3,5-Trimethylbenzene = 21.0

Ethylbenzene = 23.8

Methylene Chloride = 649

Styrene = 42.9

1,1,1-Trichloroethane = 10.9

1,3,5-Trimethylbenzene = 45.7

Trans-1,2-Dichloroethene = 14.4

5.0 CONTINUOUS AIR MONITORING PROGRAM

5.1 PROGRAM OVERVIEW

The Continuous Air Monitoring Program is described in Section 3.4. Measurements of criteria gaseous pollutants were taken continuously and recorded automatically on a data acquisition system. Summary tables of hourly averages and graphs showing daily average and maxima of the sampling data for carbon monoxide (CO), ozone (O₃), sulfur dioxide (SO₂), nitric oxide (NO), nitrogen dioxide (NO₂), and nitrogen oxides (NO_x) are presented in Appendices I and J for FY93 (October 1, 1992 through September 30, 1993) and are summarized in this section.

5.1.1 Objectives

The principal objective of the gaseous monitoring program is to identify background concentrations of pollutants. These background levels play a role in possible future remediation activities because they provide insight into the atmospheric characteristics in and around the RMA site. This data assessment provides a general overview of the gaseous concentrations and highlights any anomalous values, while helping to identify meteorological and dispersion conditions that may affect general air quality at RMA. For instance, diurnal drainage winds from the south, which frequently occur at night, and upslope winds from the north during the day affect all six gas concentrations to some extent, as well as suspended particulate matter and VOCs within the Denver airshed. Diurnal drainage winds are described in more detail in Section 2.2. In addition, daytime photochemical activity primarily influences O₃ and NO₂.

A second objective of the Continuous Air Monitoring Program is to compare the criteria pollutant concentrations at RMA with those detected at nearby regional continuous air monitoring sites. These adjacent air monitoring sites are administrated by the CDH Air Pollution Control Division. Locations of the CDH continuous air monitoring sampling sites are illustrated in Figure 5.1-1 and described in Table 5.1-1. Although there are several CDH monitoring sites located in and around the Denver metropolitan area, this report primarily compares the monitoring site at RMA to the two adjacent CDH monitoring sites (the CAMP station located at 2105 Broadway in downtown Denver, and the Welby station, located at 78th and Steele streets in Commerce City). These sites

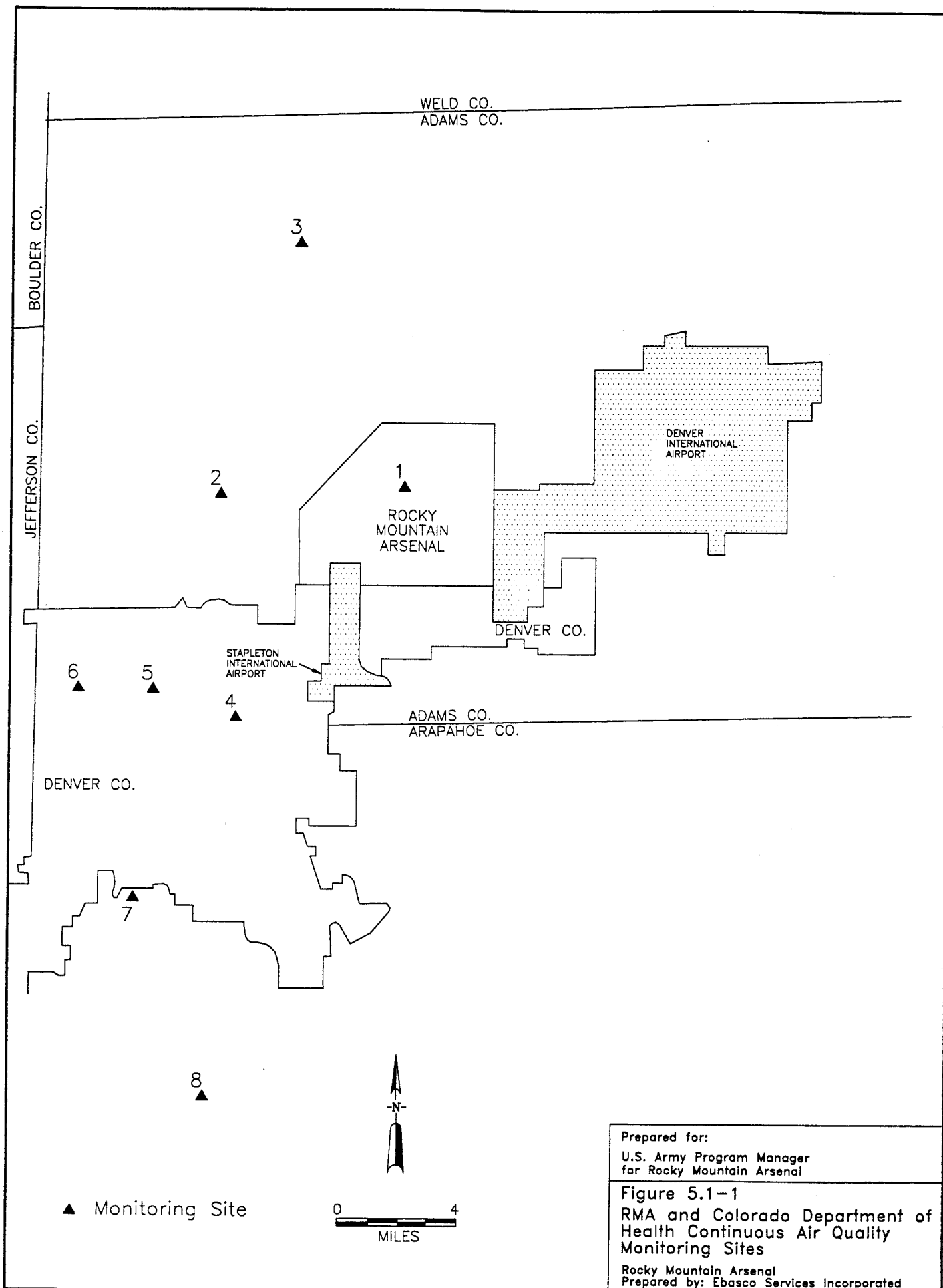


Table 5.1-1 Continuous Air Quality Monitoring Sites in the Vicinity of Rocky Mountain Arsenal

Page 1 of 1

| Map ID | Site Name | Site Address | Reported Parameters | | | | | | |
|--------|-----------|------------------------|---------------------|----------------|-----------------|----|-----------------|-----------------|----------------|
| | | | CO | O ₃ | SO ₂ | NO | NO ₂ | NO _x | PM-10 |
| 1 | RMA | 8th Avenue at D Street | X | X | X | X | X | X | X ⁺ |
| 2 | Welby | 78th and Steele | X | X | X | X | X | | X |
| 3 | Brighton | 140th and Quebec St | | X | | | | | |
| 4 | Albion | 14th at Albion | X | | | | | | |
| 5 | CAMP | 2105 Broadway | X | X | X | X | X | | X |
| 6 | Carriage | 23rd and Julian | X | X | | | | | |
| 7 | Englewood | Cinderella City | X | X | | X | X | | |
| 8 | Highland | 8100 South University | X | X | | | | | |

CO Carbon monoxide

O₃ Ozone

SO₂ Sulfur dioxide

NO Nitric oxide

NO₂ Nitrogen oxide

NO_x Nitrogen oxides

PM-10 Particulate matter less than 10 micrometers

+ 12 individual PM-10 sampling stations are installed at various locations across RMA

were chosen because they generally represent the Denver metropolitan area (CDH 1991) and because they are relatively close to RMA and are used for comparative representativeness to RMA.

The CAQMMP serves to establish background levels of ambient air quality for future assessments. Measured concentrations are compared with various meteorological data such as wind direction and atmospheric stability to identify migration patterns of gaseous pollutants from metropolitan Denver onto RMA. In addition, background levels are used to predict impacts of future remedial activity on the environment. The results shown in this section represent a complete year of data collection and provide an assessment of diurnal and annual cycles for each gas.

5.1.2 Data Recovery

Data recovery statistics for each of the monitored pollutants for FY93 are given in Table 5.1-2. Recoveries are based on the number of valid hourly measurements as compared to the total number of hours in the year. An observation was considered invalid or missing if there were less than 45 minutes of valid data recorded for the hour.

5.2 ANALYSIS OVERVIEW

A variety of tables and graphs were used to summarize continuous air quality data. Mean values, which refer to daily averages, and 1-hour maximum values, which refer to the highest 1-hour average values recorded each day, are provided. RMA data were compared to NAAQS and data from CDH stations. The analyses for carbon monoxide, ozone, and sulfur dioxide are presented individually in the following subsections. For nitrogen oxides, nitric oxide, and nitrogen dioxide, only total nitrogen oxides will be discussed because of the similarities in the chemical composition and concentration characteristics. Case studies were presented to examine the possible sources of some of the higher concentrations observed at RMA.

Table 5.1-2 Summary of Rocky Mountain Arsenal Continuous Monitoring
Data Recovery for FY93

Page 1 of 1

| Parameter | Annual Recovery % |
|------------------|-------------------|
| Ozone | 99.0 |
| Carbon Monoxide | 99.1 |
| Sulfur Dioxide | 98.6 |
| Nitric Oxide | 96.6 |
| Nitrogen Dioxide | 96.6 |
| Nitrogen Oxides | 96.6 |

% Percent

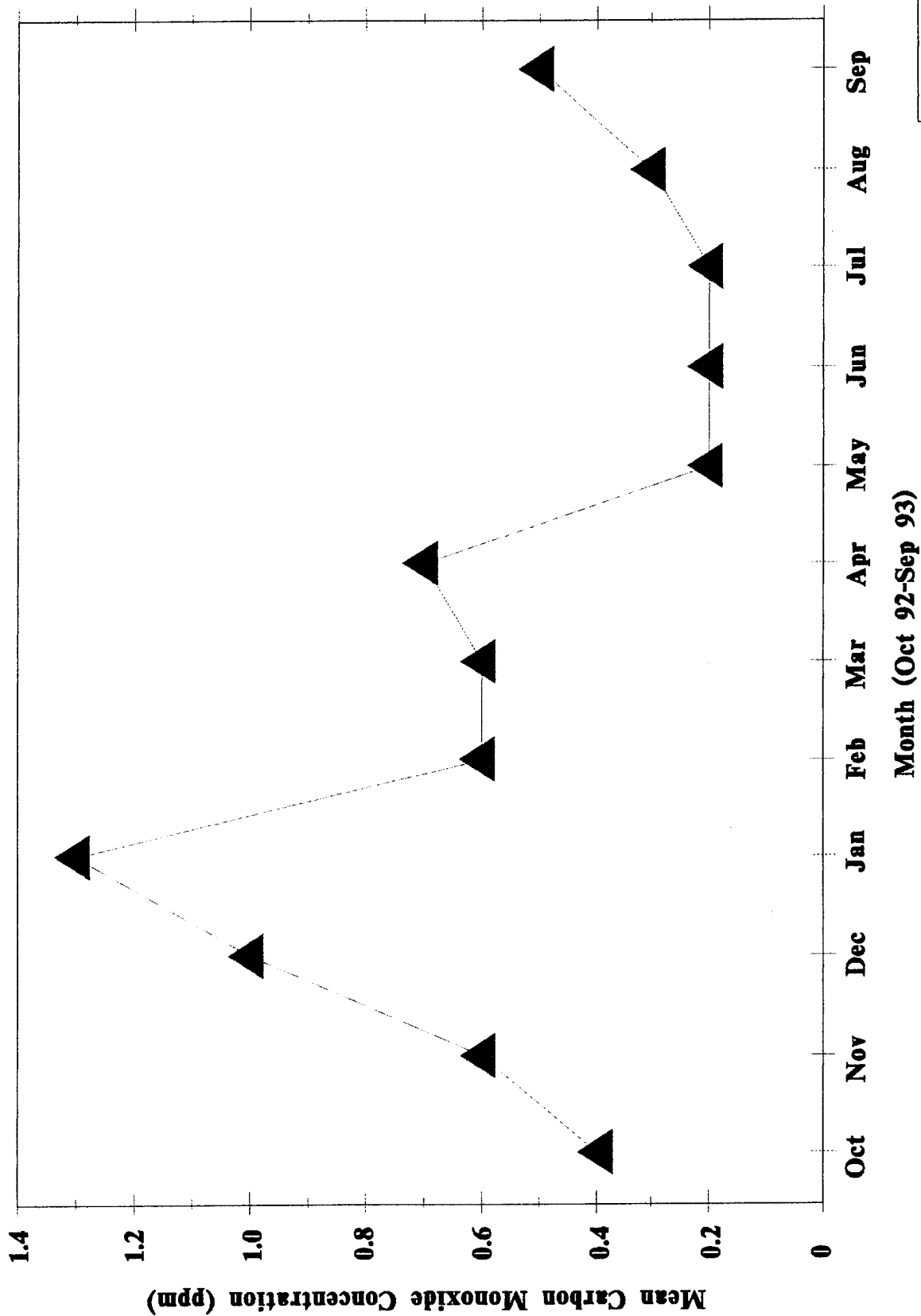
5.3 CARBON MONOXIDE

Average carbon monoxide concentrations are depicted on monthly graphs in Appendix J. During the sample collection period for FY93, there were a number of occasions when the daily maximum concentration was several times greater than the daily average. There were also several occasions during this monitoring period when the daily maximum and the daily mean concentration were nearly the same value; such instances usually occurred when persistent winds were blowing from either a northerly or a southeasterly direction. These flows allowed industrial pollutant matter from metropolitan Denver to migrate away from RMA and not be detected.

Figure 5.3-1 presents the monthly mean CO values for FY93. The gradual increase in the monthly mean concentrations during the fall and winter seasons (evident in this figure) was primarily due to shallow and intense night surface inversions and very light winds. The high concentrations began to decrease when spring brought warmer temperatures and frequently unstable atmospheric conditions.

Monthly summaries of the 1-hour and 8-hour averages are shown in Tables 5.3-1 and 5.3-2, respectively. The National and Colorado Ambient Air Quality Standards for the 1-hour and 8-hour averages, 35 ppm and 9 ppm, respectively, were never exceeded during the FY93 sample collection period at RMA. The maximum observed 1-hour concentration was 7.7 ppm on January 14, 1993, between the hours of 2300 and 2400 MST. The maximum 8-hour concentration was 4.3 ppm on January 14-15, 1993, between the hours of 2000 and 0300 MST. These maximum RMA concentrations represent 22.0 and 47.8 percent of the 1- and 8-hour standards, respectively. Figure 5.3-2 presents the diurnal cycle for CO for FY93.

As indicated in the Assessment Reports for previous fiscal years, there are no major stationary sources of carbon monoxide in the RMA area. The high CO levels represent an influx of CO from vehicles during the Denver metropolitan morning rush-hour (0600 to 1000 MST). There is also an increase in CO concentrations during the evening hours between 1700 and 2200 MST. This evening increase may be the result of several factors, including the onset of the evening



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for Rocky Mountain Arsenal

Figure 5.3-1

RMA FY93 Monthly Mean
Carbon Monoxide Concentration
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

Table 5.3-1 Summary of Carbon Monoxide 1-Hour Average Values in ppm¹: October 1, 1992 (0100 MST) through September 30, 1993 (2400 MST)

Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|------------------------|------|------|------|------|------|------|------|-----|------|------|------|-------|
| Mean | 0.4 | 0.6 | 1.0 | 1.3 | 0.6 | 0.6 | 0.7 | 0.2 | 0.2 | 0.2 | 0.3 | 0.5 |
| Maximum | 3.3 | 3.8 | 6.7 | 7.7 | 4.7 | 2.7 | 2.3 | 2.1 | 2.1 | 1.4 | 1.3 | 2.2 |
| 2nd Highest Maximum | 2.9 | 3.5 | 6.5 | 6.2 | 4.4 | 2.5 | 2.1 | 1.8 | 1.5 | 1.2 | 1.3 | 2.0 |
| Minimum | 0.1 | 0.1 | 0.1 | 0.5 | 0.1 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| Mean for Entire Period | 0.54 | | | | | | | | | | | |

¹ Federal and Colorado Ambient Air Quality Standard for maximum 1-hour average values is 35 ppm, not to be exceeded more than once a year.
MST Mountain Standard Time
ppm Parts per million

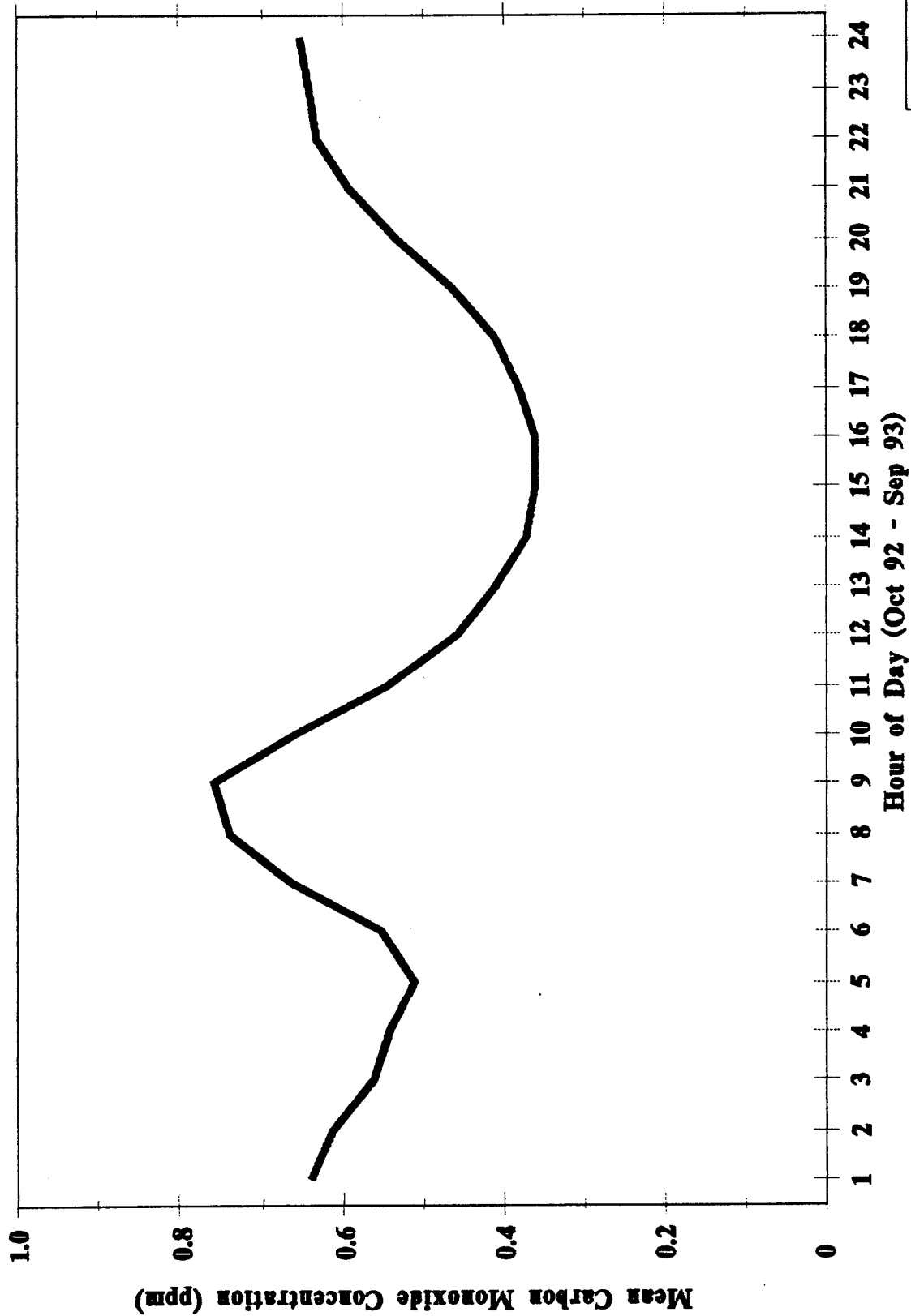
Table 5.3-2 Summary of Carbon Monoxide 8-Hour Average Values in ppm¹: October 1, 1992 (0100 MST) through September 30, 1993 (2400 MST)

Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|---------------------|------|------|------|------|------|------|------|-----|------|------|------|-------|
| Maximum | 2.0 | 2.0 | 4.1 | 4.3 | 3.0 | 1.6 | 1.4 | 1.3 | 0.8 | 0.8 | 0.9 | 1.2 |
| 2nd Highest Maximum | 2.0 | 2.0 | 4.0 | 4.3 | 3.0 | 1.6 | 1.4 | 1.3 | 0.8 | 0.7 | 0.8 | 1.2 |

¹ Federal and Colorado Ambient Air Quality Standard for maximum 8-hour average values is 9 ppm, not to be exceeded more than once a year.

MST Mountain Standard Time
ppm parts per million



Prepared for:
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for Rocky Mountain Arsenal

Figure 5.3-2

RMA FY93 Diurnal Cycle for
Carbon Monoxide Concentration
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

inversion, a shift in the wind direction, evening heating of homes (especially those with wood stoves), and the evening rush hour.

Figures 5.3-3 and 5.3-4 compare the highest 1-hour and 8-hour carbon monoxide concentrations recorded at RMA to the two CDH sites, Welby and CAMP, for FY93. These two graphs show that concentrations recorded at RMA were lower than both CDH locations and significantly lower than the CAMP site for the fall and winter months.

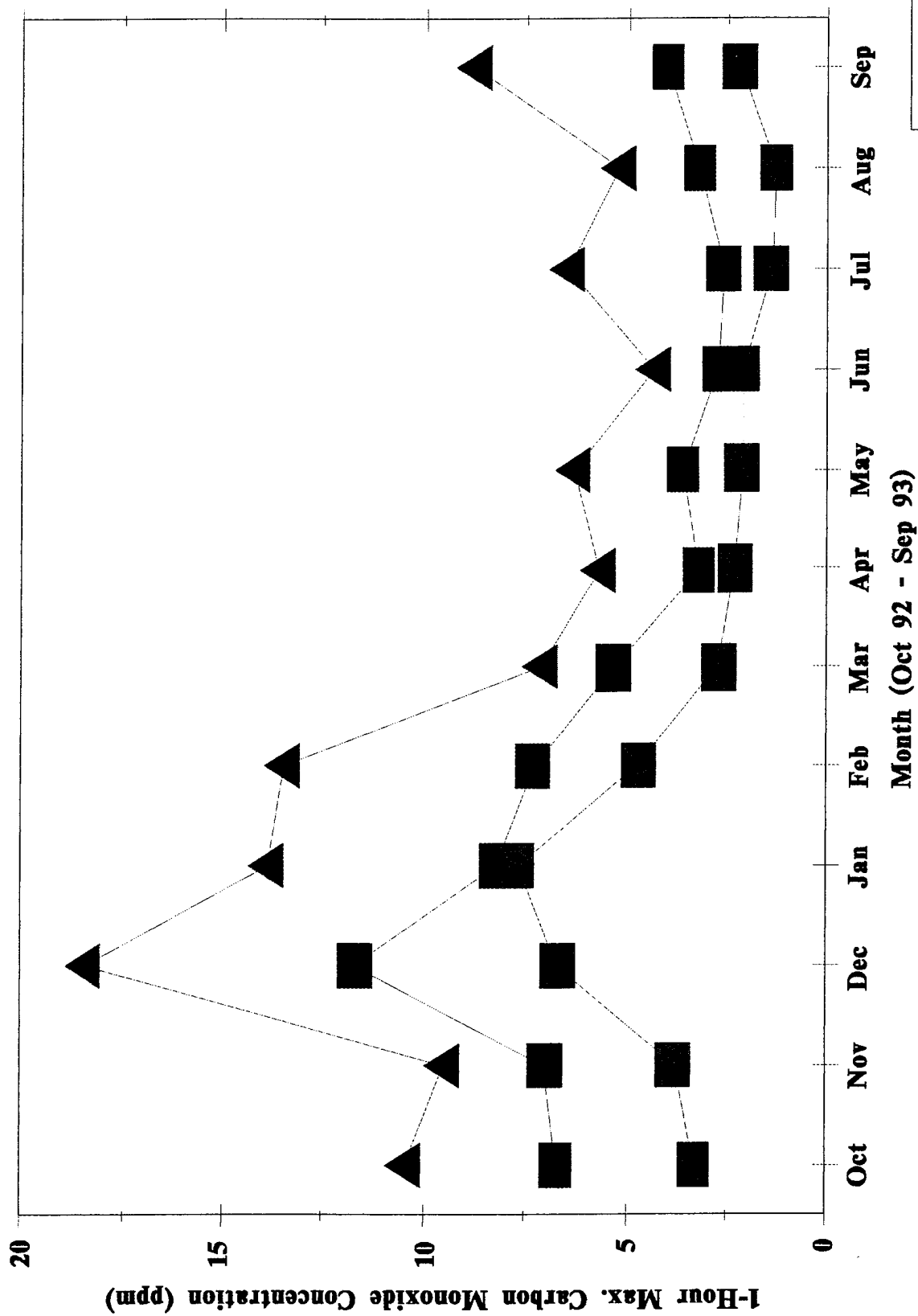
5.4 OZONE

Table 5.4-1 presents a summary of monthly mean and maximum ozone concentrations at the RMA site for FY93. Figure 5.4-1 displays the monthly mean ozone values for FY93. This graph also illustrates a very evident annual cycle for ozone. During the late spring and through the summer ozone increased, with peak concentrations observed during June and July. The figure also shows minimum concentrations for ozone observed during the winter months.

A very distinct diurnal pattern for ozone is illustrated in Figure 5.4-2. Peak ozone concentrations occurred between the hours of 1300 and 1600 MST, when solar radiation played a major role in the formation of ozone. Ozone concentrations decreased after the peak hours and stabilized at lower levels between 2400 and 0500 MST. Between the hours of 0500 and 0700 MST, ozone concentrations were further reduced. This reduction may be due to the increase in NO, NO₂, and NO_x concentrations associated with the beginning of the Denver metropolitan area rush-hour; or the consumption of ozone in the conversion of NO to NO₂.

The highest ozone concentration, 0.099 ppm, was recorded on August 20, 1993, between the hours of 1400 and 1500 MST. This concentration represents 83 percent of the primary 1-hour National and Colorado Ambient Air Quality Standard of 0.120 ppm for ozone.

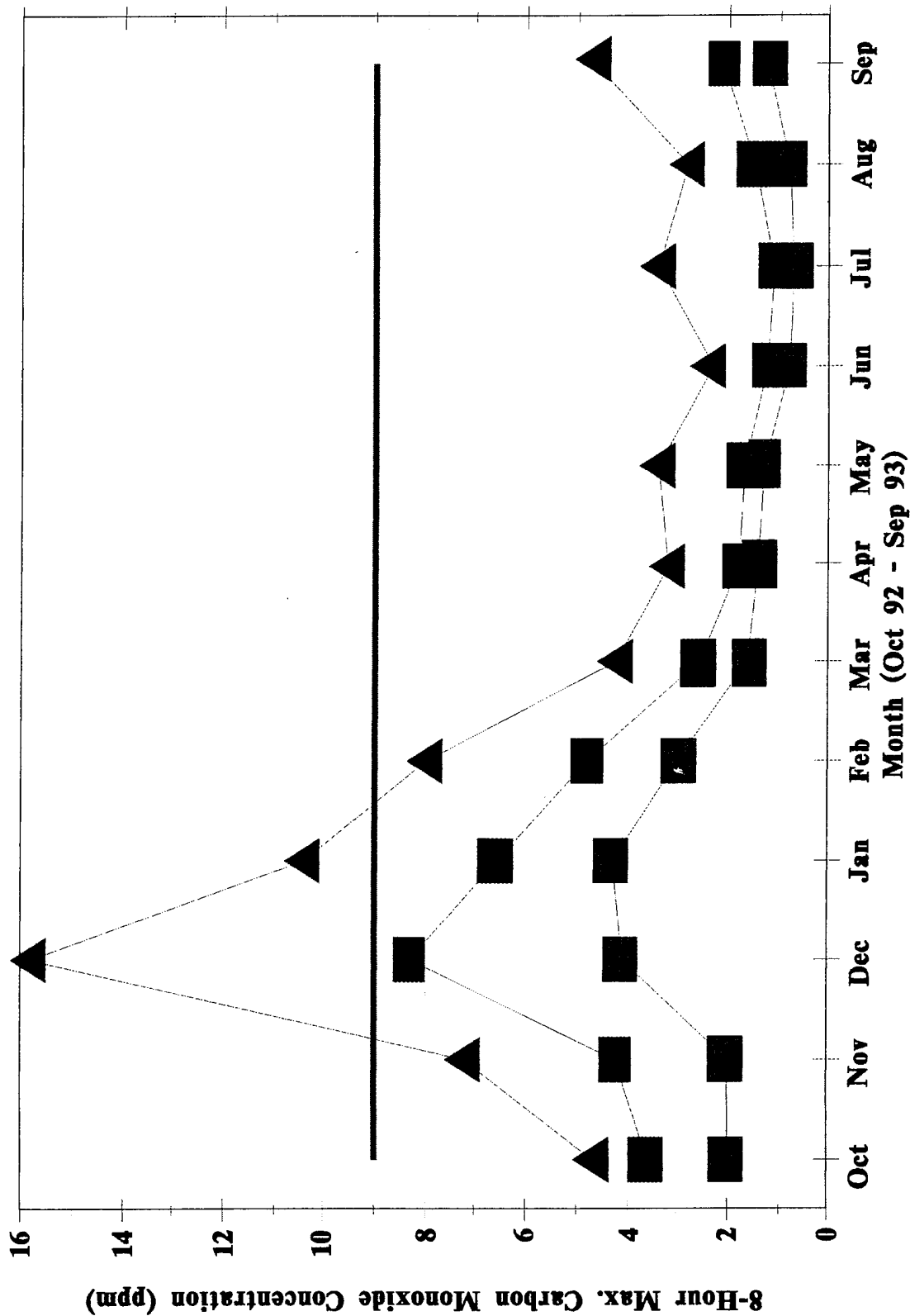
Figure 5.4-3 compares the highest 1-hour ozone concentrations recorded at RMA to the two CDH sites for the monitoring period. Concentrations at RMA were comparable but slightly higher than



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Figure 5.3-3

FY93 Comparison of 1-Hour
 Maximum Carbon Monoxide Con-
 centration for RMA and CDH Sites
 Rocky Mountain Arsenal
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Figure 5.3-4

FY93 Comparison of 8-Hour
Maximum Carbon Monoxide Con-
centration for RMA and CDH Sites
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

Legend:

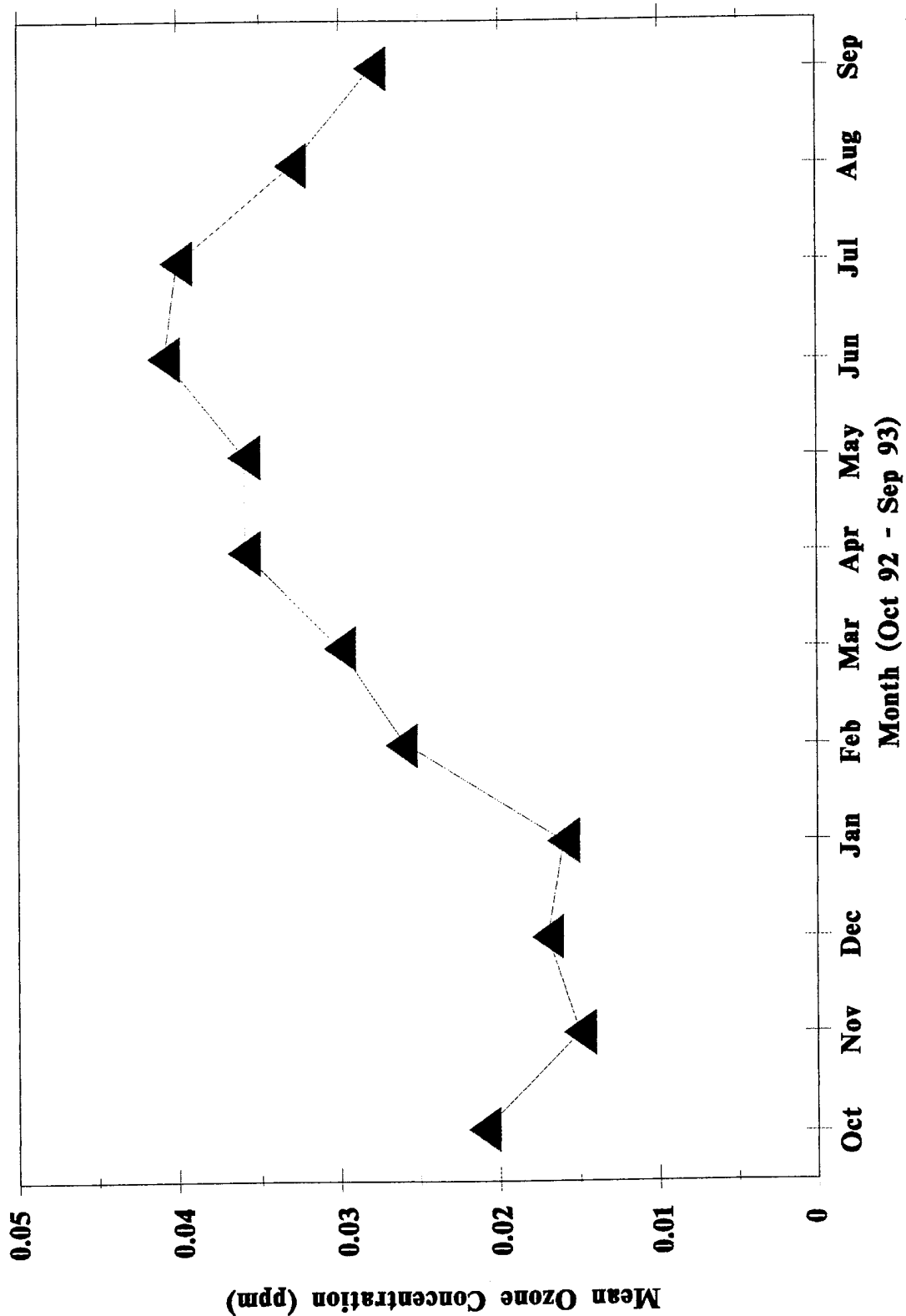
▲ CAMP ■ Welby ■ RMA — NAAQS

Table 5.4-1 Summary of Ozone 1-Hour Average Values in ppm¹: October 1, 1992 (0100 MST) through September 30, 1993
(2400 MST) Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|------------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Mean | 0.021 | 0.015 | 0.017 | 0.016 | 0.026 | 0.030 | 0.036 | 0.036 | 0.041 | 0.040 | 0.033 | 0.028 |
| Maximum | 0.073 | 0.063 | 0.044 | 0.045 | 0.062 | 0.063 | 0.075 | 0.081 | 0.088 | 0.085 | 0.099 | 0.067 |
| 2nd Highest Maximum | 0.071 | 0.053 | 0.041 | 0.044 | 0.060 | 0.063 | 0.068 | 0.079 | 0.080 | 0.082 | 0.096 | 0.066 |
| Minimum | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| Mean for Entire Period | 0.028 | | | | | | | | | | | |

¹ Federal and Colorado Ambient Air Quality Standard for maximum 1-hour average values is 0.120 ppm.

MST Mountain Standard Time
ppm parts per million



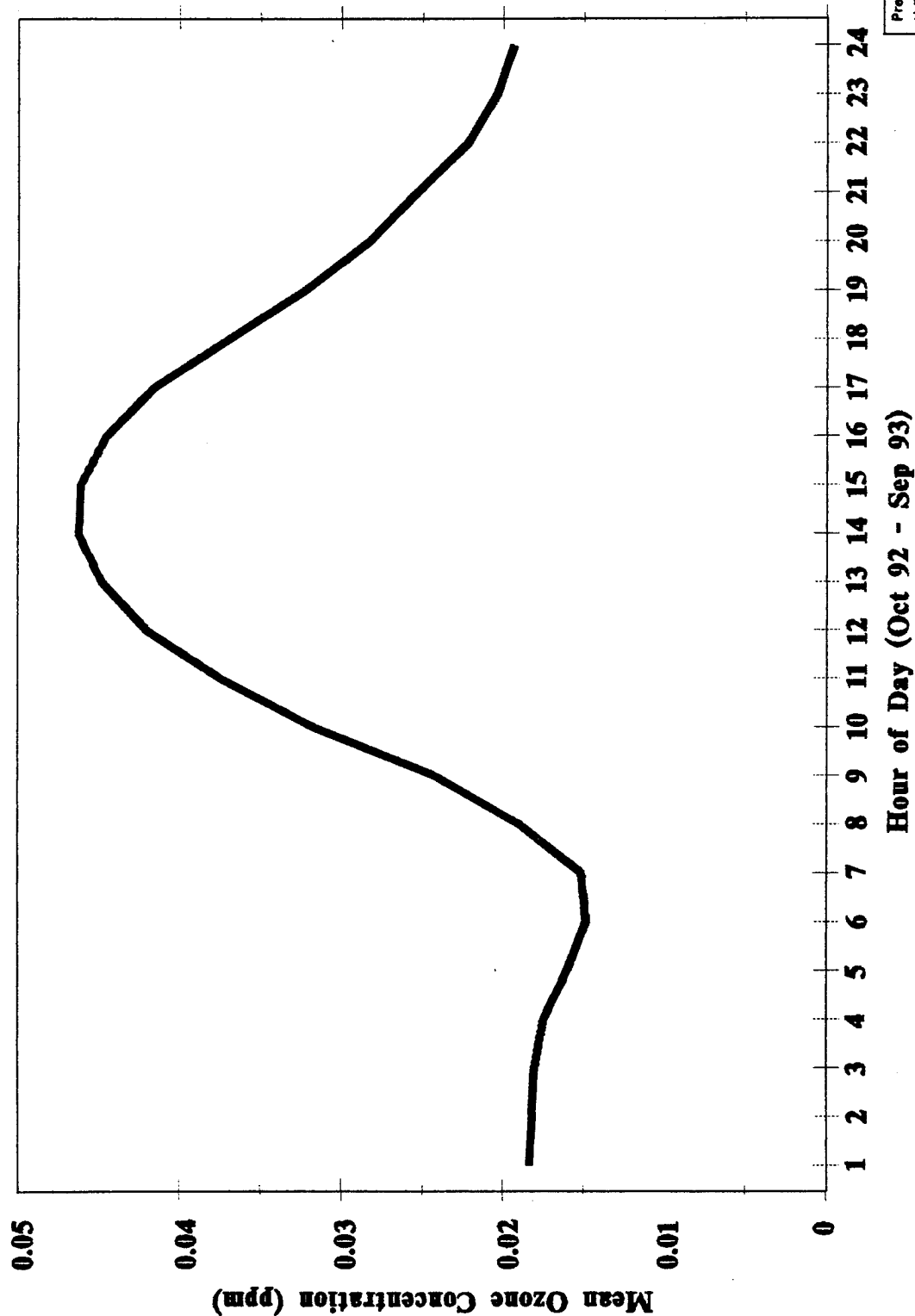
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for Rocky Mountain Arsenal

Figure 5.4-1

RMA FY93 Monthly Mean
Ozone Concentration

Rocky Mountain Arsenal
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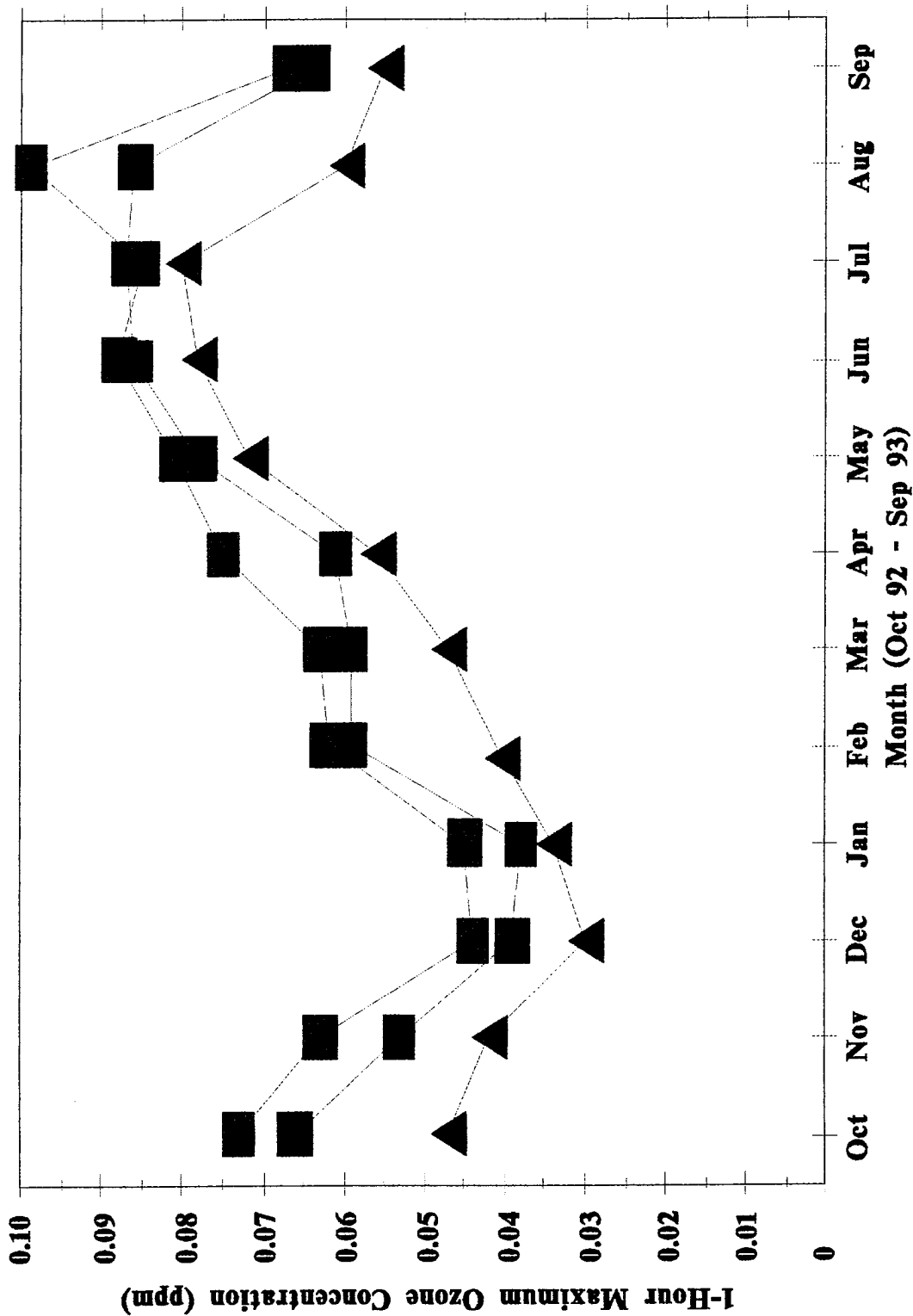
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for Rocky Mountain Arsenal

Figure 5.4-2

RMA FY93 Diurnal Cycle
for Ozone Concentration

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

— RMA



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 5.4-3

FY93 Comparison of 1-Hour
Maximum Ozone (concentration)
for RMA and CDH Sites

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

▲ CAMP ■ Welby ◆ RMA

the CDH-measured concentrations. Individual graphs and tables presenting daily maximum, minimum, and mean concentrations for each month are included in Appendices I and J.

5.5 SULFUR DIOXIDE

The daily mean and maximum sulfur dioxide concentrations at the RMA site for FY93 are presented in Appendices I and J. Tables 5.5-1, 5.5-2 and 5.5-3 present monthly summaries of 1-hour, 3-hour, and 24-hour concentrations, respectively. During this data collection period, there were numerous occasions where the daily maximum concentration was several times greater than the daily mean concentration. This most likely reflects the transport of SO₂ onto RMA from metropolitan Denver SO₂ sources. There were also instances in which the daily maximum concentration was nearly the same as the daily mean concentration, reflecting periods when pollutants from metropolitan Denver were not migrating to RMA. The monthly mean concentrations were higher in the winter months when low-level inversions over metropolitan Denver are more frequent (as shown in Figure 5.5-1).

The diurnal pattern for sulfur dioxide at RMA is illustrated in Figure 5.5-2. A sharp increase in SO₂ concentrations is evident between the hours of 0600 and 1000 MST. The increase in sulfur dioxide may be caused by an increase in power generation due to greater demand for electricity during these hours, other industrial activity, or early morning inversions.

The maximum 3-hour SO₂ concentration for FY93 was 0.039 ppm, which occurred on March 18, 1993 between the hours of 1000 and 1200 MST. This concentration represents 7.8 percent of the primary 3-hour SO₂ National Ambient Air Quality Standard of 0.50 ppm and 14.4 percent of the Colorado standard of 0.27 ppm. The maximum 24-hour SO₂ concentration during FY93 at RMA was 0.015 ppm, which occurred on January 14, 1993. This concentration is 10.7 percent of the 24-hour SO₂ National Ambient Air Quality Standard of 0.14 ppm. The annual average concentration measured at RMA during FY93 was 0.002 ppm, which is 6.7 percent of the annual SO₂ National Ambient Air Quality Standard of 0.03 ppm.

Table 5.5-1 Summary of Sulfur Dioxide 1-Hour Average Values in ppm¹: October 1, 1992 (0100 MST) through September 30, 1993 (2400 MST)

Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|------------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Mean: | 0.004 | 0.003 | 0.004 | 0.004 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 |
| Maximum | 0.067 | 0.038 | 0.054 | 0.053 | 0.026 | 0.081 | 0.031 | 0.046 | 0.047 | 0.040 | 0.060 | 0.020 |
| 2nd Highest Maximum | 0.036 | 0.038 | 0.048 | 0.052 | 0.017 | 0.034 | 0.017 | 0.030 | 0.039 | 0.030 | 0.022 | 0.018 |
| Minimum | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| Mean for Entire Period | 0.002 | | | | | | | | | | | |

¹ National (NAAQS) and Colorado Ambient Air Quality Standard for annual arithmetic mean is 0.030 ppm. (There is no NAAQS 1-hour standard for SO₂.)
MST Mountain Standard Time
ppm parts per million

Table 5.5-2 Summary of Sulfur Dioxide 3-Hour Average Values in ppm¹: October 1, 1992 (0100 MST) through September 30, 1993 (2400 MST) Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|---------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Maximum | 0.029 | 0.031 | 0.035 | 0.029 | 0.013 | 0.039 | 0.018 | 0.026 | 0.025 | 0.024 | 0.027 | 0.015 |
| 2nd Highest Maximum | 0.029 | 0.020 | 0.030 | 0.029 | 0.011 | 0.015 | 0.010 | 0.019 | 0.023 | 0.018 | 0.011 | 0.009 |

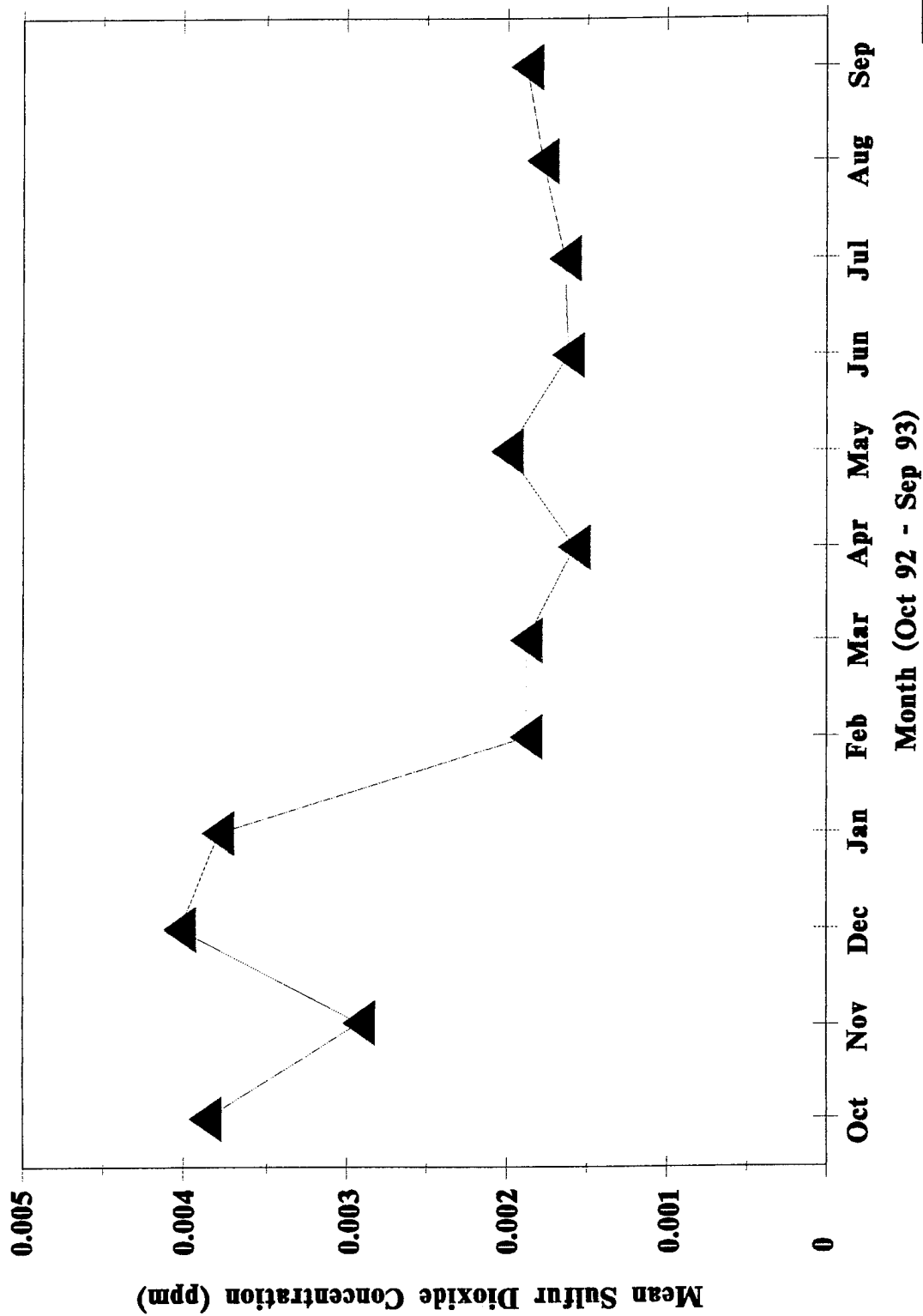
¹ National Ambient Air Quality Standard for maximum 3-hour average values is 0.500 ppm, not to be exceeded more than once per year. Colorado standard for 3-hour average is 0.270 ppm.

MST Mountain Standard Time
ppm parts per million

Table 5.5-3 Summary of Sulfur Dioxide 24-Hour Average Values in ppm¹: October 1, 1992 (0100 MST) through September 30, 1993 (2400 MST) Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|---------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Maximum | 0.011 | 0.008 | 0.015 | 0.015 | 0.004 | 0.007 | 0.004 | 0.006 | 0.005 | 0.004 | 0.004 | 0.004 |
| 2nd Highest Maximum | 0.009 | 0.005 | 0.013 | 0.013 | 0.004 | 0.004 | 0.003 | 0.005 | 0.004 | 0.004 | 0.004 | 0.004 |

¹ National Ambient Air Quality Standard for maximum 24-hour average values is 0.140 ppm, not to be exceeded more than once per year.
MST Mountain Standard Time
ppm parts per million

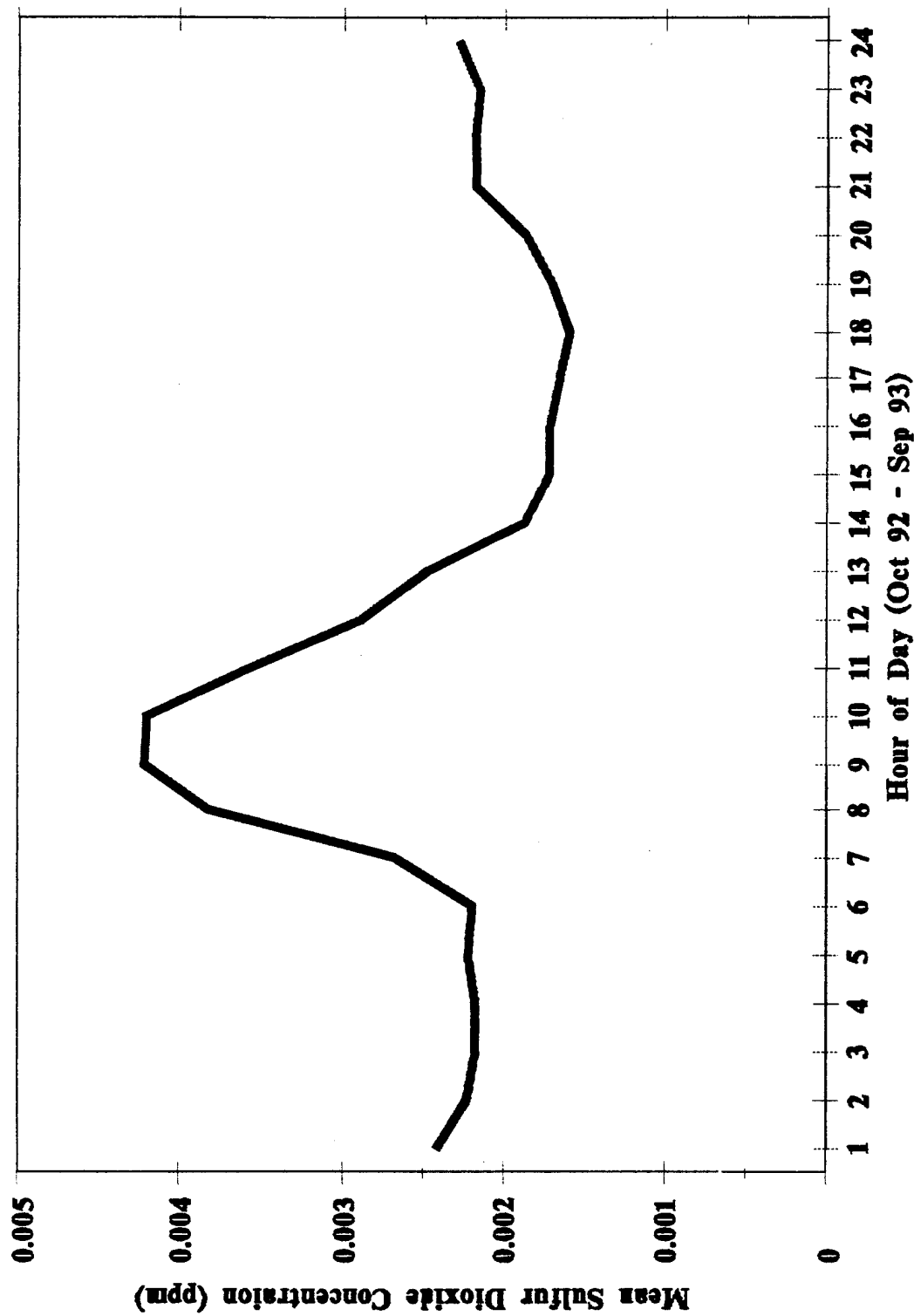


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Figure 5.5-1

RMA FY93 Monthly Mean
Sulfur Dioxide Concentration

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



— RMA

Prepared for:
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for Rocky Mountain Arsenal

Figure 5.5-2
RMA FY93 Diurnal Cycle
for Sulfur Dioxide
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

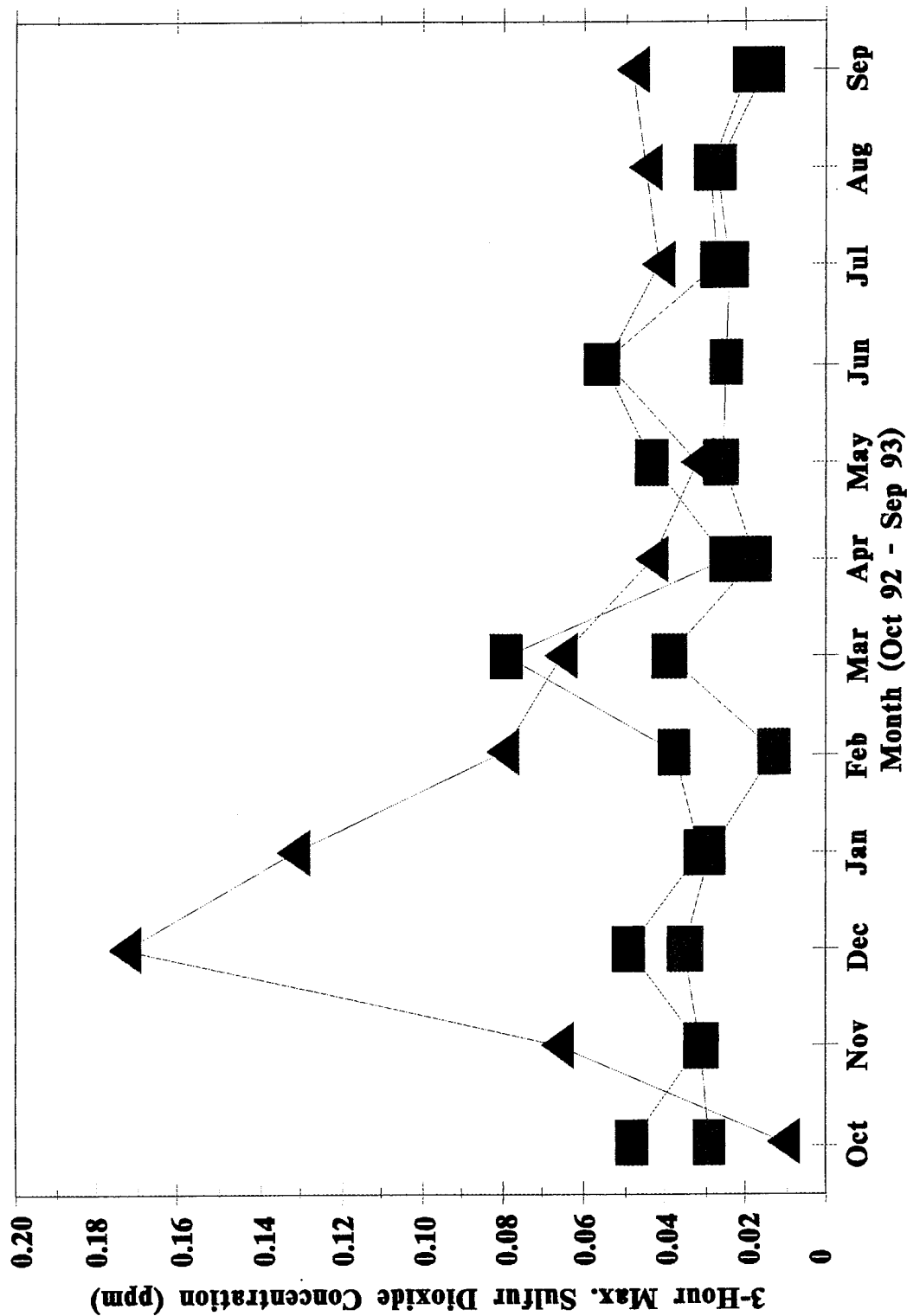
Figures 5.5-3 and 5.5-4 compare the highest 3-hour and 24-hour sulfur dioxide concentrations recorded at RMA with the two CDH sites for FY93. The RMA site recorded slightly lower values than the Welby site, and considerably lower values than the CAMP location. Individual graphs and tables presenting daily maximum, minimum, and mean concentrations for each month are provided in Appendices I and J.

5.6 NITRIC OXIDE, NITROGEN DIOXIDE, AND NITROGEN OXIDES

Daily mean concentrations for NO, NO₂, and NO_x are presented on a monthly basis in tables and graphs included in Appendices I and J. These graphs show an annual cycle for each parameter with peak concentrations most prevalent during December through February. This cycle is further illustrated in Figures 5.6-1, 5.6-2, and 5.6-3, which display mean monthly concentrations for FY93.

Monthly summaries for 1-hour average concentrations of NO, NO₂, and NO_x are given in Tables 5.6-1, 5.6-2 and 5.6-3, respectively. The National Ambient Air Quality Standard for NO₂ is 0.053 ppm and is an annual arithmetic mean. The annual mean NO₂ concentration for FY93, 0.018 ppm, represents 34 percent of the standard.

Since seasonal and diurnal trends of NO, NO₂, and NO_x are interrelated, an assessment of these three gases as a whole was made using NO_x as the indicator. The diurnal cycle for these gases illustrate a similar pattern to each other with peak concentrations between the hours of 0700 and 0900 MST, which coincided with the Denver metropolitan area morning rush-hour. For the remainder of the day, these gases exhibited their lowest concentrations, with the minimum concentration occurring at 1600 MST. NO_x concentrations then increased slightly during the late evening hours, possibly from the reformation of the surface inversion. This cycle is depicted in Figure 5.6-4. Individual graphs and tables presenting daily maximum, minimum, and mean concentrations for each day of each month are contained in Appendices I and J.

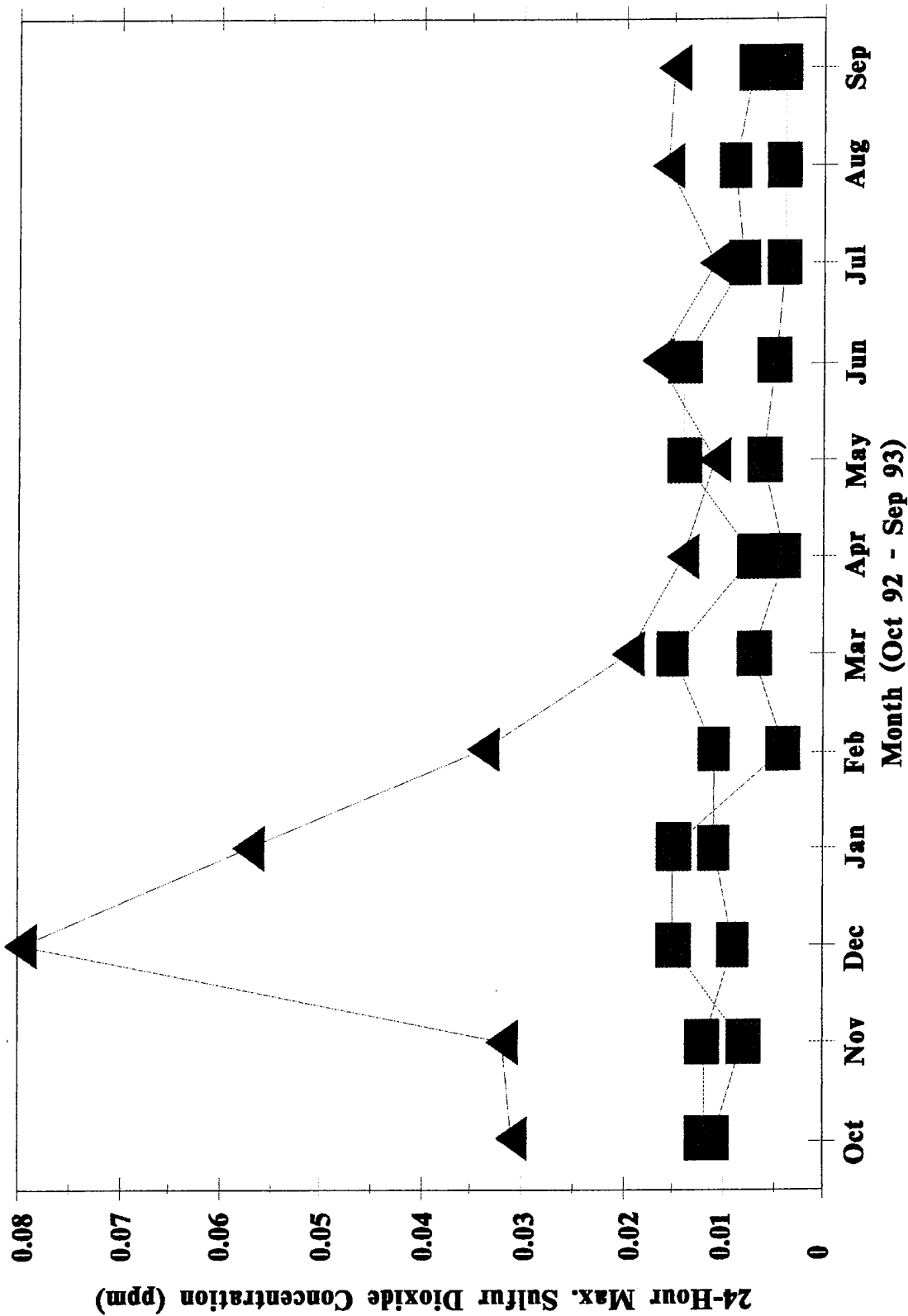


Prepared for:
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for Rocky Mountain Arsenal

Figure 5.5-3

FY93 Comparison of 3-Hour
Maximum Sulfur Dioxide for
RMA and CDH Sites

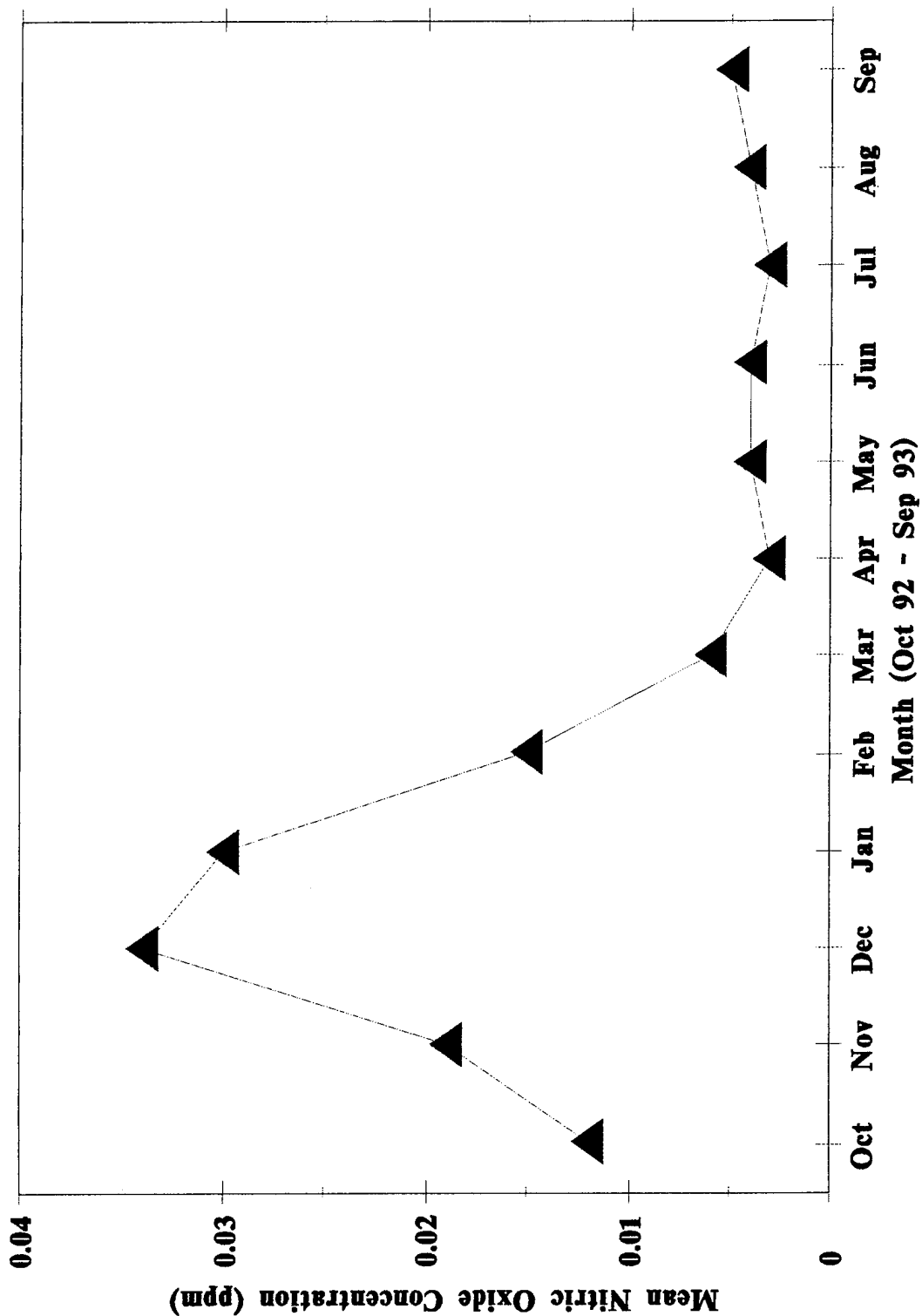
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



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Figure 5.5-4

FY93 Comparison of 24-Hour
 Maximum Sulfur Dioxide for
 RMA and CDH Sites
 Rocky Mountain Arsenal
 Prepared by: Ebasco Services Incorporated

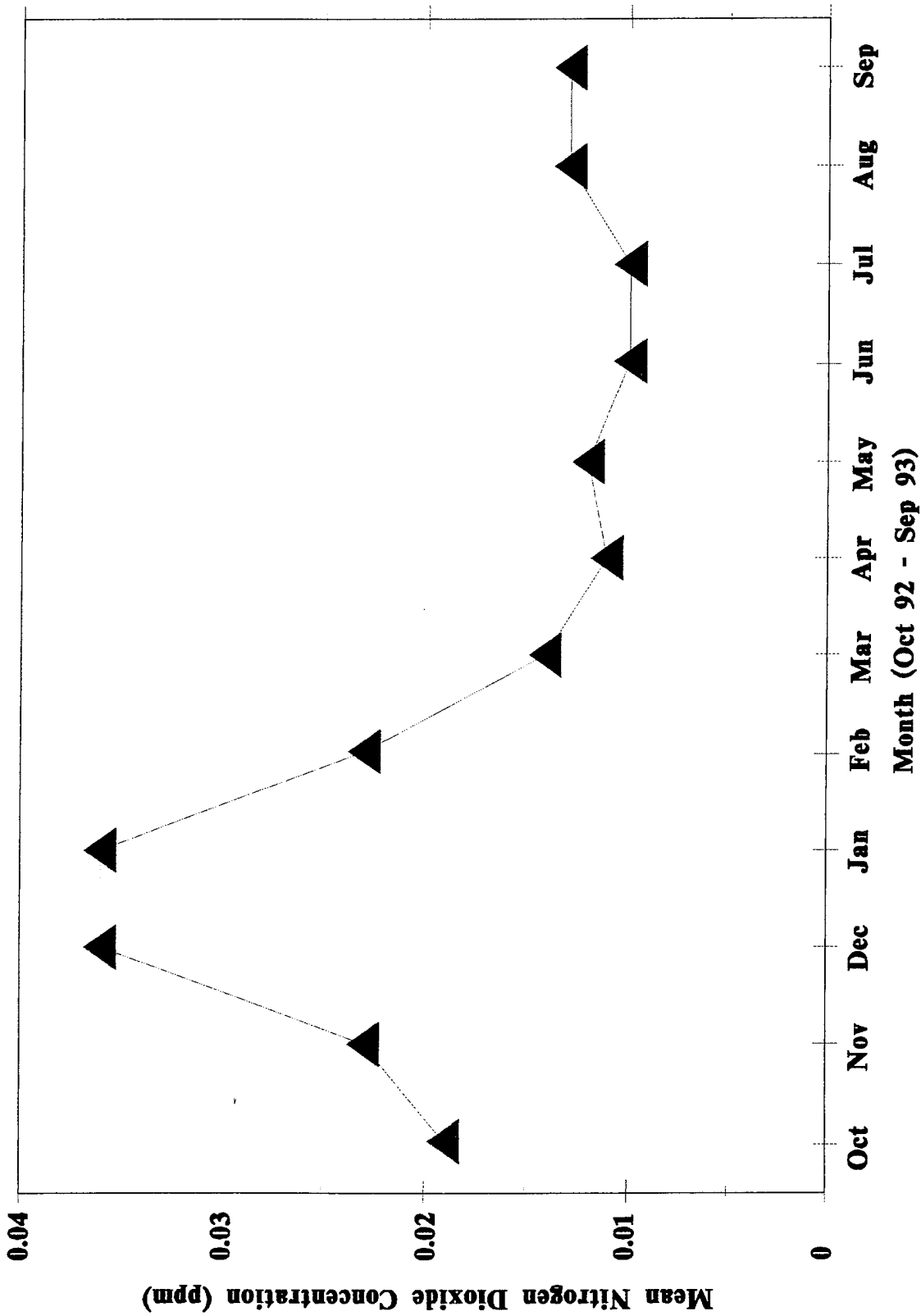


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Figure 5.6-1

RMA FY93 Monthly Mean
Nitric Oxide Concentration

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

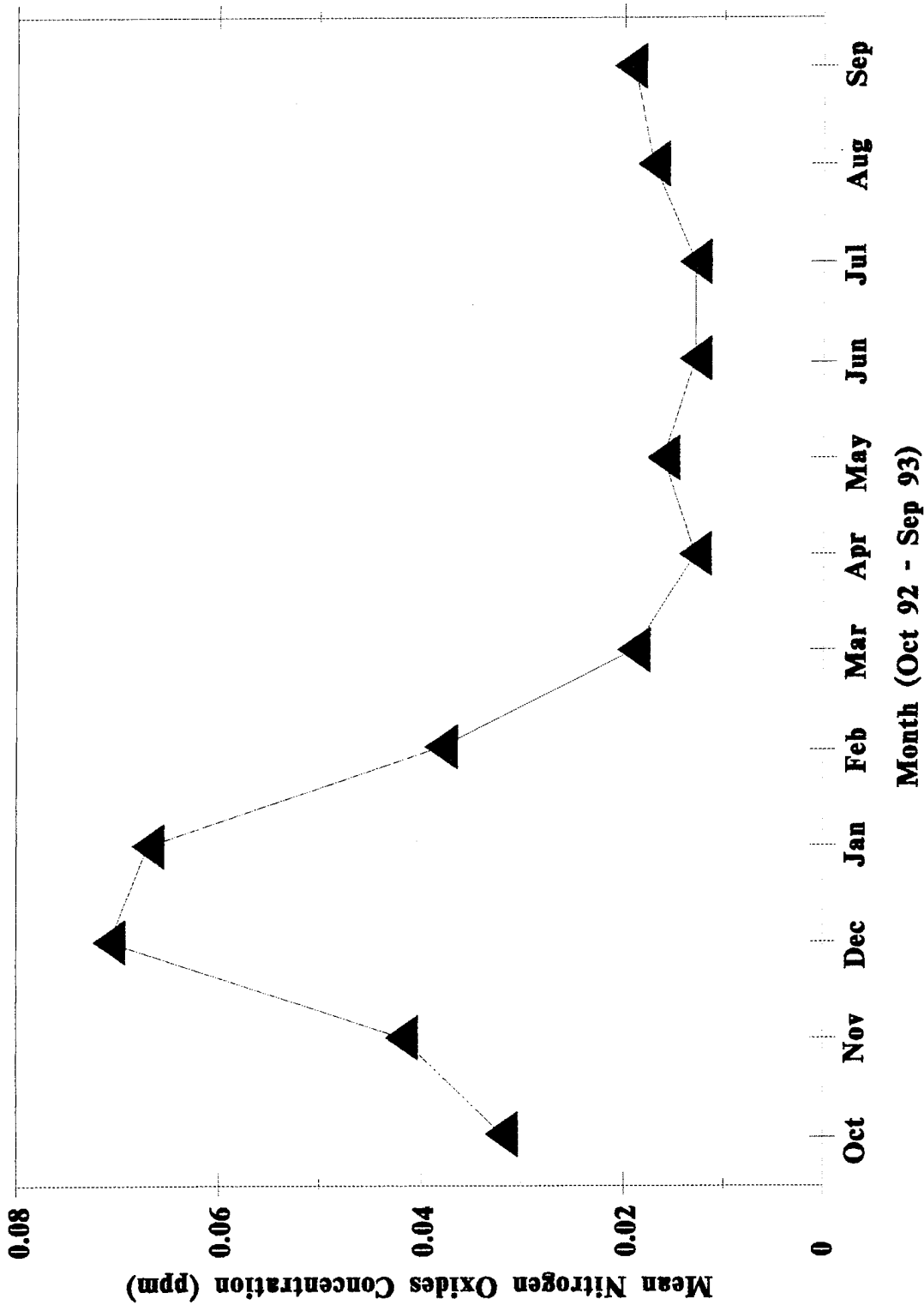


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Figure 5.6-2

RMA FY93 Monthly Mean
Nitrogen Dioxide

Rocky Mountain Arsenal
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Figure 5.6-3

RMA FY93 Monthly Mean
Nitrogen Oxides

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

Table 5.6-1 Summary of Nitric Oxide 1-Hour Average Values in ppm: October 1, 1992 (0100 MST) through September 30, 1993 (2400 MST)

Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|------------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Mean | 0.012 | 0.019 | 0.034 | 0.030 | 0.015 | 0.006 | 0.003 | 0.004 | 0.004 | 0.003 | 0.004 | 0.005 |
| Maximum | 0.184 | 0.202 | 0.317 | 0.387 | 0.271 | 0.142 | 0.074 | 0.103 | 0.084 | 0.044 | 0.048 | 0.088 |
| 2nd Highest Maximum | 0.170 | 0.171 | 0.302 | 0.327 | 0.242 | 0.113 | 0.072 | 0.085 | 0.064 | 0.037 | 0.040 | 0.077 |
| Minimum | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| Mean for Entire Period | 0.011 | | | | | | | | | | | |

MST Mountain Standard Time
ppm parts per million

RMA/0868 10/06/94 1:11 pm bpw

Table 5.6-2 Summary of Nitrogen Dioxide 1-Hour Average Values in ppm¹: October 1, 1992 (0100 MST) through September 30, 1993 (2400 MST) Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|------------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Mean | 0.019 | 0.023 | 0.036 | 0.036 | 0.023 | 0.014 | 0.011 | 0.012 | 0.010 | 0.010 | 0.013 | 0.013 |
| Maximum | 0.062 | 0.099 | 0.140 | 0.178 | 0.113 | 0.076 | 0.066 | 0.069 | 0.066 | 0.082 | 0.064 | 0.055 |
| 2nd Highest Maximum | 0.061 | 0.094 | 0.132 | 0.170 | 0.110 | 0.075 | 0.056 | 0.056 | 0.058 | 0.066 | 0.044 | 0.055 |
| Minimum | 0.001 | 0.001 | 0.002 | 0.004 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| Mean for Entire Period | 0.018 | | | | | | | | | | | |

¹ National and Colorado Ambient Air Quality Standard for annual arithmetic mean is 0.053 ppm.

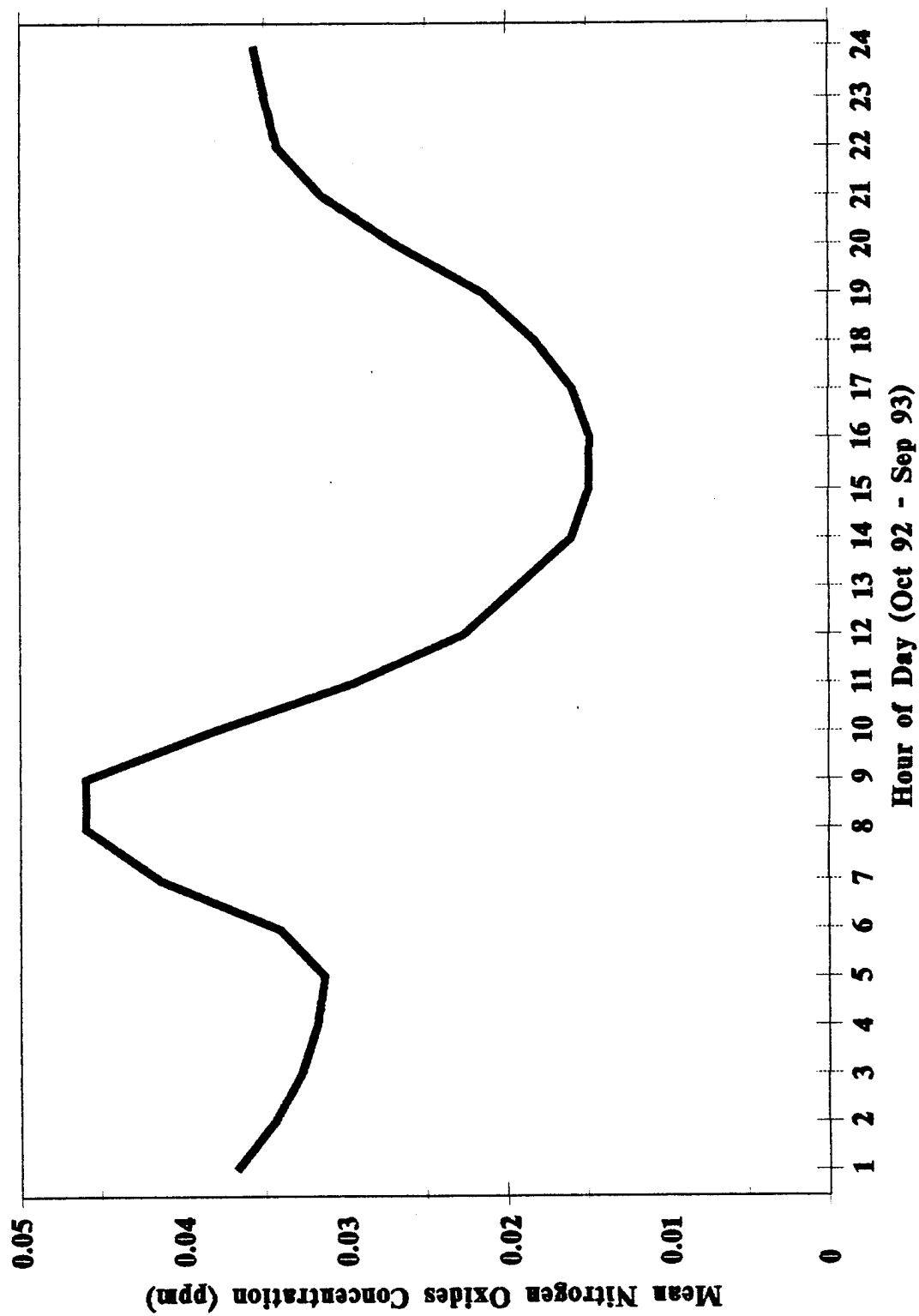
MST Mountain Standard Time
ppm parts per million

Table 5.6-3 Summary of Nitrogen Oxides 1-Hour Average Values in ppm¹: October 1, 1992 (0100 MST) through September 30, 1993 (2400 MST) Page 1 of 1

| | Oct. | Nov. | Dec. | Jan. | Feb. | Mar. | Apr. | May | June | July | Aug. | Sept. |
|------------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Mean | 0.032 | 0.042 | 0.071 | 0.067 | 0.038 | 0.019 | 0.013 | 0.016 | 0.013 | 0.013 | 0.017 | 0.019 |
| Maximum | 0.234 | 0.271 | 0.442 | 0.525 | 0.371 | 0.218 | 0.124 | 0.154 | 0.119 | 0.088 | 0.082 | 0.140 |
| 2nd Highest Maximum | 0.216 | 0.255 | 0.423 | 0.466 | 0.353 | 0.188 | 0.116 | 0.140 | 0.099 | 0.086 | 0.082 | 0.113 |
| Minimum | 0.001 | 0.001 | 0.002 | 0.005 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| Mean for Entire Period | 0.030 | | | | | | | | | | | |

MST Mountain Standard Time
ppm parts per million

RMA/0871 10/06/94 1:12 pm bpw



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 5.6-4

RMA FY93 Diurnal Cycle for
Nitrogen Oxides

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

Several case studies are presented in the next section to show the interaction between metropolitan Denver source emissions, meteorological conditions, and ambient concentrations measured at RMA.

5.7 REGIONAL EMISSION SOURCES IMPACTING RMA

Tables 5.7-1 through 5.7-3 provide listings of criteria pollutant emissions (carbon monoxide, sulfur dioxide, and nitrogen oxides) from major metropolitan Denver industrial sources. Ozone is not emitted directly from a source as are other pollutants, but forms as a secondary pollutant. Its precursors are certain reactive hydrocarbons and nitrogen oxides, which react chemically with each other in sunlight (CDH 1991). Those sources listed in Tables 5.7-1 through 5.7-3 can therefore be considered as contributors to ozone production.

In addition to stationary sources, automobile exhaust is a principal precursor of ozone. In the winter, automobile emissions produce 86 percent of the carbon monoxide (CO) measured in Denver and about 33 percent of the NO_x (CDH 1989). Both stationary and mobile emissions (primarily from vehicle traffic) jointly contribute to the ambient air quality conditions measured at RMA. Note that the RMA gaseous emission sources are quite small compared to the significant major sources in the Denver area, as shown in Tables 5.7-1 through 5.7-3.

Figures 5.7-1 through 5.7-3 show the distribution of stationary sources surrounding RMA (source locations are cross-referenced in Tables 5.7-1 through 5.7-3). Most of the sources are to the south and southwest of RMA, where the major vehicle activity also occurs. Monitoring results from the RMA station showed levels of CO, SO₂, and NO_x that were generally below metropolitan Denver monitoring values. However, when prevailing winds associated with a strong inversion blew in the direction of RMA, external sources significantly impacted RMA air quality.

A review of the air quality and meteorological data for RMA during FY93 indicates a number of probable pollution migration episodes onto RMA. Case studies for TSP, PM-10, metals, and

Table 5.7-1 Sources in the Metropolitan Denver Area Ranked by Reported Emissions of
Carbon Monoxide
[Sources with emissions greater than 1 ton per year (tpy)]

Page 1 of 2

| Facility | City | Update Year | Map No. | Estimate (tpy) | |
|---|---------------|----------------|------------|-------------------|-----|
| Public Service Co. - Cherokee Plant | Denver | 91 | a | 550.0 | (1) |
| Colorado Interstate Gas Co. - Watkins Station | Watkins | 91 | 1 | 335.0 | |
| Conoco Inc. Denver Refinery | Commerce City | 91 | b | 272.0 | (1) |
| Public Service Co. - Arapahoe Plant | Denver | 91 | c | 160.0 | (1) |
| Colorado Refining Co. | Commerce City | 91 | d | 141.0 | (1) |
| Metro Wastewater Reclamation District | Denver | 91 | 2 | 83.8 | |
| Koch Hydrocarbon Co. Third Creek Plant | Brighton | 92 | 3 | 65.2 | |
| Buckley ANG Base - Colo. Air Nat'l Guard | Aurora | 91 | 4 | 56.0 | |
| Owens-Corning Fiberglas Corp-Trumbull | Denver | 92 | 5 | 41.5 | |
| Public Service Company - Delganey | Denver | 91 | 6 | 22.7 | |
| Koch Hydrocarbon Co. Mitchell Station | Byers | 92 | 7 | 22.3 | |
| Koch Hydrocarbon Co. - Radar Plant | Commerce City | 92 | 8 | 19.4 | |
| Koch Hydrocarbon Co. - Dragoon Station | Bennett | 92 | 9 | 18.6 | |
| Explosive Fabricators Inc | Englewood | 80 | 10 | 16.7 | |
| Brannan S & G | Denver | 90 | 11 | 16.1 | |
| Colorado Interstate Gas Company Latigo | Watkins | 91 | 12 | 16.0 | |
| Bar S Foods Company | Denver | 92 | 13 | 15.2 | |
| Koch Hydrocarbon Co. - Boxelder Station | Denver | 92 | 14 | 14.6 | |
| Presbyterian/St. Lukes Healthcare | Denver | 90 | 15 | 13.6 | |
| Littleton/Englewood WWTP | Englewood | 92 | 16 | 10.7 | |
| Public Service Company - Zuni | Denver | 92 | 17 | 10.5 | |
| Koch Hydrocarbon Co. Rattlesnake Station | Bennett | 92 | 18 | 9.6 | |
| ITT Continental Baking Company 80 E 62nd | Denver | 90 | 19 | 9.3 | |
| Schafer Commercial Seating Inc | Denver | 92 | 20 | 8.6 | |
| Panhandle Eastern Pipe Line Co. Brighton | Brighton | 92 | 21 | 8.6 | |
| Lowry AFB | Denver | 90 | 22 | 8.5 | |
| Koch Hydrocarbon Co. - Antelope station | Bennett | 92 | 23 | 8.5 | |
| Irondale Gas Processing Co. Irondale | Brighton | 92 | 24 | 7.9 | |
| Sinclair Oil Denver Product Terminal | Henderson | 92 | 25 | 7.6 | |
| Univ of Colo Health Sciences Center | No City Name | 90 | 26 | 7.2 | |
| Sand Creek Chemical | Commerce City | 92 | 27 | 7.1 | |
| Rocky Mountain Arsenal - Dept of Army | Commerce City | 92 | 28 | 7.0 | |
| Ralston Purina Company Pet Food Plant | Denver | 92 | 29 | 6.9 | |
| Fitzsimons Army Hospital | Aurora | 90 | 30 | 6.8 | |
| Republic Paperboard Company | Commerce City | 92 | 31 | 6.6 | |
| AT&T Information Systems | Westminster | 90 | 32 | 6.0 | |
| The Childrens Hospital | Denver | 92 | 33 | 4.2 | |
| Robinson Brick #2 | Denver | 90 | 34 | 4.1 | |
| National By-products, Inc | Denver | 92 | 35 | 4.0 | |
| Koch Hydrocarbon Co. - Kallsen Plant | Watkins | 92 | 36 | 3.9 | |

Emissions estimates obtained from EPA Aerometric Information Retrieval System.

(1) Emissions based on February, 1993 data (as presented in FY92 Data Report)

Table 5.7-1 Sources in the Metropolitan Denver Area Ranked by Reported Emissions of
Carbon Monoxide
[Sources with emissions greater than 1 ton per year (tpy)]

Page 2 of 2

| Facility | City | Update Year | Map No. | Estimate (tpy) |
|---|---------------|----------------|------------|-------------------|
| Denver and Rio Grande Western RR | Denver | 90 | 37 | 3.7 |
| US West 12121 Grant | No City name | 90 | 38 | 3.5 |
| St Joseph Hospital | Denver | 90 | 39 | 3.5 |
| Sfuzzi VII Denver Inc. | no City Name | 90 | 40 | 3.5 |
| Koch Hydrocarbon Co. - State Plant | Watkins | 92 | 41 | 3.5 |
| PepCol Manufacturing Company | Denver | 90 | 42 | 3.4 |
| Keebler Company | Denver | 92 | 43 | 3.3 |
| Radison Hotel Denver | Denver | 92 | 44 | 3.2 |
| BFI - Tower Landfill | Commerce City | 92 | 45 | 3.1 |
| AG - Power of Colorado | Denver | 90 | 46 | 2.9 |
| Presbyterian Medical Center | Denver | 90 | 47 | 2.7 |
| Central Products Co. | Brighton | 91 | 48 | 2.7 |
| RustCo Products Co. | Denver | 91 | 49 | 2.5 |
| U.S. West | No City Name | 90 | 50 | 2.4 |
| Owens Corning | Denver | 92 | 51 | 2.2 |
| Bituminous Roadways of Co, Inc. | Commerce City | 92 | 52 | 2.1 |
| Swedish Medical Center | Englewood | 91 | 53 | 2.0 |
| Provenant Healthcare Partners | Denver | 92 | 54 | 2.0 |
| City and County of Denver - Stapleton Airport | Denver | 90 | 55 | 1.8 |
| Safeway Milk Plant | Denver | 92 | 56 | 1.7 |
| Rocky Mountain Prestress | Denver | 92 | 57 | 1.7 |
| Koch Hydrocarbon Co. Texaco Station | Bennett | 92 | 58 | 1.6 |
| Denver General Hospital | Denver | 88 | 59 | 1.6 |
| Western Paving Const.Co.-North Plant | Denver | 92 | 60 | 1.5 |
| Denver Public Schools - Kepner Jr High | No City Name | 87 | 61 | 1.5 |
| Rocky Mtn Dyeing & Finishing | Denver | 89 | 62 | 1.4 |
| Red Seal Inc. | Denver | 90 | 63 | 1.4 |
| Metrum Information Storage | Littleton | 92 | 64 | 1.4 |
| Denver Public Schools - Abraham Lincoln | No City Name | 83 | 65 | 1.4 |
| The Electron Corporation | Littleton | 91 | 66 | 1.3 |
| VA Hospital Denver | Denver | 88 | 67 | 1.2 |
| United Airlines | Denver | 89 | 68 | 1.2 |
| Humana Hospital Mt View | Thornton | 89 | 69 | 1.2 |
| Dow Chemical USA | Aurora | 92 | 70 | 1.2 |
| Packaging Corp. of America | Commerce City | 92 | 71 | 1.1 |
| Koppers Co | Adams Co | 88 | 72 | 1.1 |
| Gardner-Denver | Denver | 90 | 73 | 1.0 |

Emissions estimates obtained from EPA Aerometric Information Retrieval System.

(1) Emissions based on February, 1993 data (as presented in FY92 Data Report)

Table 5.7-2 Sources in the Metropolitan Denver Area Ranked by Reported Emissions
of Sulfur Dioxide

Page 1 of 1

[Sources with emissions greater than 1 ton per year (tpy)]

| Facility | City | Update Year | Map No. | Estimate (tpy) |
|---|---------------|----------------|------------|-------------------|
| Public Service Company - Cherokee | Denver | 91 | 1 | 13362.3 |
| Public Service Company - Arapahoe | Denver | 91 | 2 | 4567.0 |
| Conoco Inc. - Denver Refinery | Commerce City | 91 | 3 | 2336.1 |
| Colorado Refining Company / Total Petroleum | Commerce City | 91 | 4 | 631.7 |
| Metro Wastewater Reclamation District | Denver | 91 | 5 | 134.0 |
| Buckley ANG Base - Colo. Air Nat'l Guard | Aurora | 91 | 6 | 37.1 |
| Owens-Corning Fiberglas Corp | Denver | 92 | 7 | 28.7 |
| General Chemical Corporation 1271 W Bayaud | Denver | 90 | 8 | 28.0 |
| The Gates Rubber Co. | Denver | 90 | 9 | 27.9 |
| Bar S Foods Company | Denver | 92 | 10 | 25.0 |
| Fast Construction Company 64th and Steel | Commerce City | 90 | 11 | 24.6 |
| Amoco Production Co. - Wattenberg Plant | Watkins | 92 | 12 | 23.5 |
| Asarco Incorporated - Globe Plant | Denver | 91 | 13 | 17.5 |
| Kiewit Western | Portable | 90 | 14 | 15.9 |
| Brannan S & G | Denver | 90 | 15 | 15.5 |
| Western Paving Const.Co.-North Plant | Denver | 92 | 16 | 8.8 |
| Milt Adams Inc Custom Refining | Aurora | 88 | 17 | 8.7 |
| Colorado Special Chemical Inc Commerce City | Westminster | 90 | 18 | 8.0 |
| National Smelting and Refining Company | Commerce City | 90 | 19 | 7.2 |
| Northwestern Engineering 6001 Dexter | Commerce City | 90 | 20 | 6.2 |
| Rocky Mountain Arsenal - Dept of Army | Commerce City | 92 | 21 | 6.0 |
| Quikrete Colorado | Denver | 90 | 22 | 5.9 |
| OEA Inc. | Aurora | 90 | 23 | 5.2 |
| Univ of Colo Health Sciences Center | No City Name | 90 | 24 | 3.8 |
| Duwald Steel Corporation | Denver | 92 | 25 | 2.2 |
| Public Service Company - Zuni | Denver | 92 | 26 | 1.8 |
| Rocky Mountain Prestress | Denver | 92 | 27 | 1.4 |
| BFI - Tower Landfill | Commerce City | 92 | 28 | 1.2 |
| U.S. West | No City Name | 90 | 29 | 1.0 |
| U.S. West | No City Name | 90 | 30 | 1.0 |

Emission estimates obtained from EPA Aerometric Information Retrieval Information.

Table 5.7-3 Sources in the Metropolitan Denver Area Ranked by Reported Emissions of
Nitrogen Dioxide
[Sources with emissions greater than 5 tons per year (tpy)]

Page 1 of 2

| Facility | City | Update Map | | Estimate (tpy) |
|--|---------------|------------|-----|-------------------|
| | | Year | No. | |
| Public Service Company - Cherokee | Denver | 91 | 1 | 14295.2 |
| Public Service Company - Arapahoe | Denver | 91 | 2 | 5143.3 |
| Colorado Interstate Gas Co - Watkins Station | Watkins | 91 | 3 | 865.5 |
| Conoco Inc. - Denver Refinery | Commerce City | 91 | 4 | 625.8 |
| Amoco Production Co. - Wattenberg Plant | Watkins | 92 | 5 | 548.7 |
| Koch Hydrocarbon Co. Third Creek Plant | Brighton | 92 | 6 | 503.7 |
| Public Service Company - Delaney | Denver | 91 | 7 | 313.3 |
| Colorado Refining Company / Total Petroleum | Commerce City | 91 | 8 | 287.5 |
| Buckley ANG Base - Colo. Air Nat'l Guard | Aurora | 91 | 9 | 263.1 |
| The Gates Rubber Co. | Denver | 90 | 10 | 215.2 |
| Koch Hydrocarbon Co. Mitchell Station | Byers | 92 | 11 | 175.8 |
| Koch Hydrocarbon Co. - Radar Plant | Commerce City | 92 | 12 | 153.3 |
| Koch Hydrocarbon Co. - Dragoon Station | Bennett | 92 | 13 | 146.2 |
| Public Service Company - Zuni Station | Denver | 92 | 14 | 145.6 |
| Colorado Interstate Gas Company Latigo | Watkins | 91 | 15 | 125.8 |
| Koch Hydrocarbon Co. - Boxelder Station | Denver | 92 | 16 | 115.0 |
| Gulf Energy Development Chalice Plant | Aurora | 90 | 17 | 96.0 |
| Presbyterian/St. Lukes Healthcare | Denver | 90 | 18 | 91.6 |
| Littleton/Englewood WWTP | Englewood | 92 | 19 | 80.4 |
| Koch Hydrocarbon Co. Rattlesnake Station | Bennett | 92 | 20 | 75.5 |
| Panhandle Eastern Pipe Line Co. Brighton | Brighton | 92 | 21 | 68.4 |
| Koch Hydrocarbon Co. - Antelope Station | Bennett | 92 | 22 | 67.3 |
| Irondale Gas Processing Co. Irondale | Brighton | 92 | 23 | 62.9 |
| Metro Wastewater Reclamation District | Denver | 91 | 24 | 48.0 |
| ITT Continental Baking Company 80 E 62nd | Denver | 90 | 25 | 46.7 |
| Lowry AFB | Denver | 90 | 26 | 38.8 |
| Rocky Mountain Arsenal - Dept of Army | Commerce City | 92 | 27 | 32.6 |
| Koch Hydrocarbon Co. - Kallsen Plant | Watkins | 92 | 28 | 31.2 |
| Univ of Colo Health Sciences Center | No City Name | 90 | 29 | 30.6 |
| Ralston Purina Company Pet Food Plant | Denver | 92 | 30 | 30.6 |
| Owens-Corning Fiberglas Corp-Trumbull | Denver | 92 | 31 | 28.7 |
| Koch Hydrocarbon Co. - State Plant | Watkins | 92 | 32 | 27.9 |
| Fitzsimons Army Hospital | Aurora | 90 | 33 | 27.3 |
| Koch Hydrocarbon Co Denver Central Plant | Strasburg | 92 | 34 | 27.2 |
| Republic Paperboard Company | Commerce City | 92 | 35 | 26.6 |
| AT&T Info. Systems | Westminster | 90 | 36 | 24.1 |
| Western Paving Const.Co.-North Plant | Denver | 92 | 37 | 23.9 |
| Sand Creek Chemical | Commerce City | 92 | 38 | 23.5 |
| Kiewit Western | Portable | 90 | 39 | 19.0 |
| Bar S Foods Company | Denver | 92 | 40 | 18.8 |
| The Childrens Hospital | Denver | 92 | 41 | 18.5 |

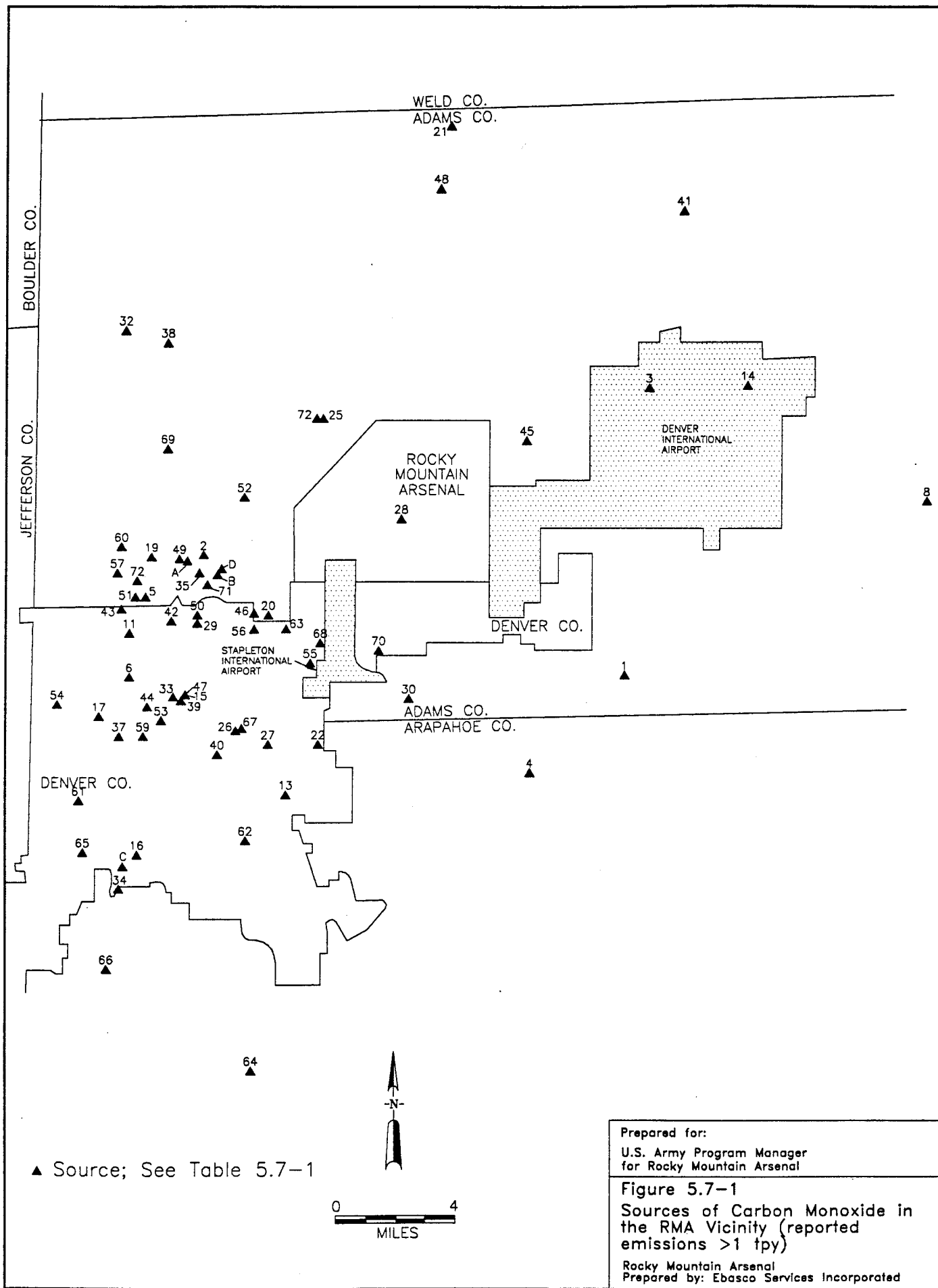
Emission estimates obtained from EPA Aerometric Information Retrieval System

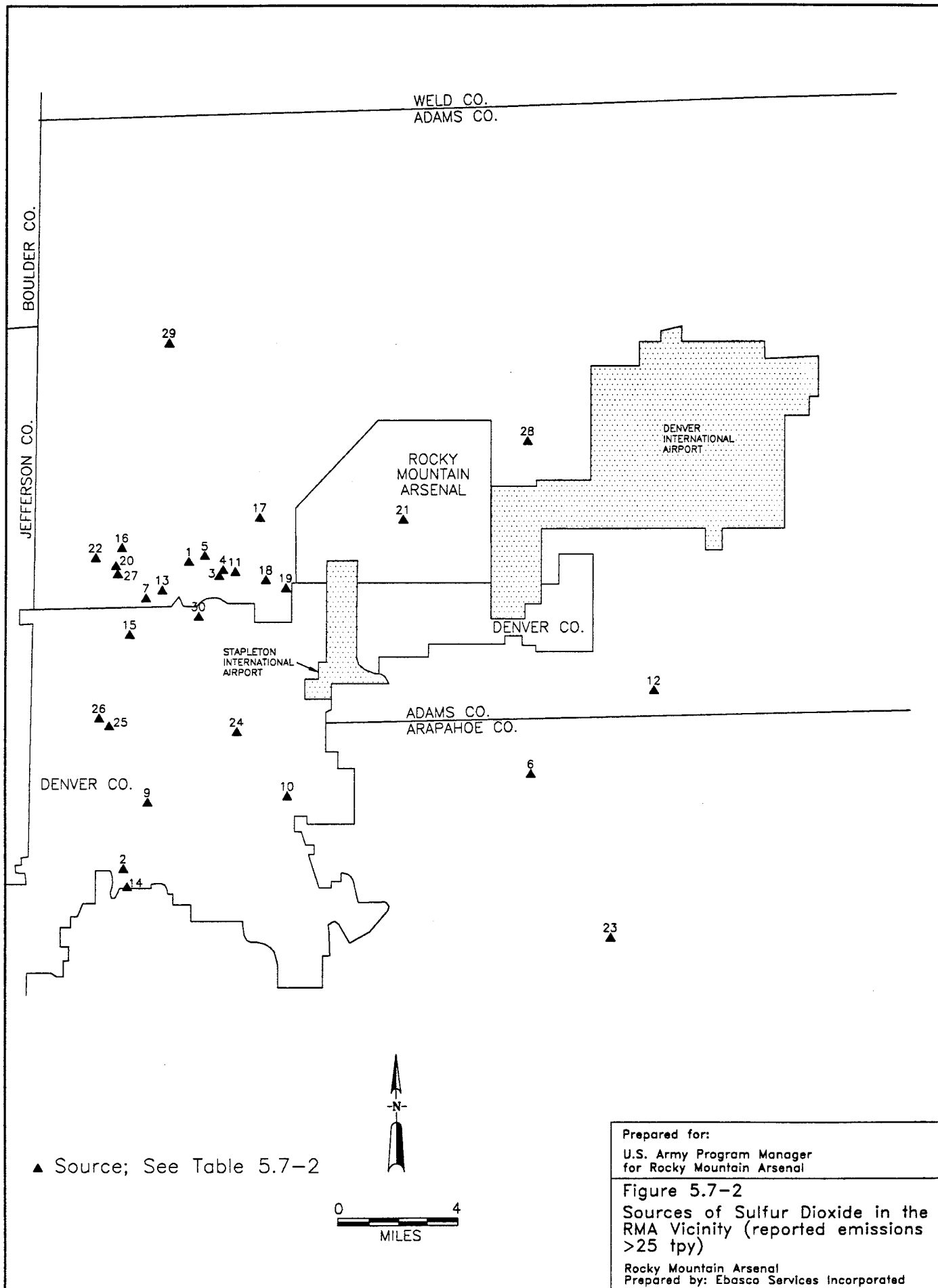
Table 5.7-3 Sources in the Metropolitan Denver Area Ranked by Reported Emissions of
Nitrogen Dioxide
[Sources with emissions greater than 5 tons per year (tpy)]

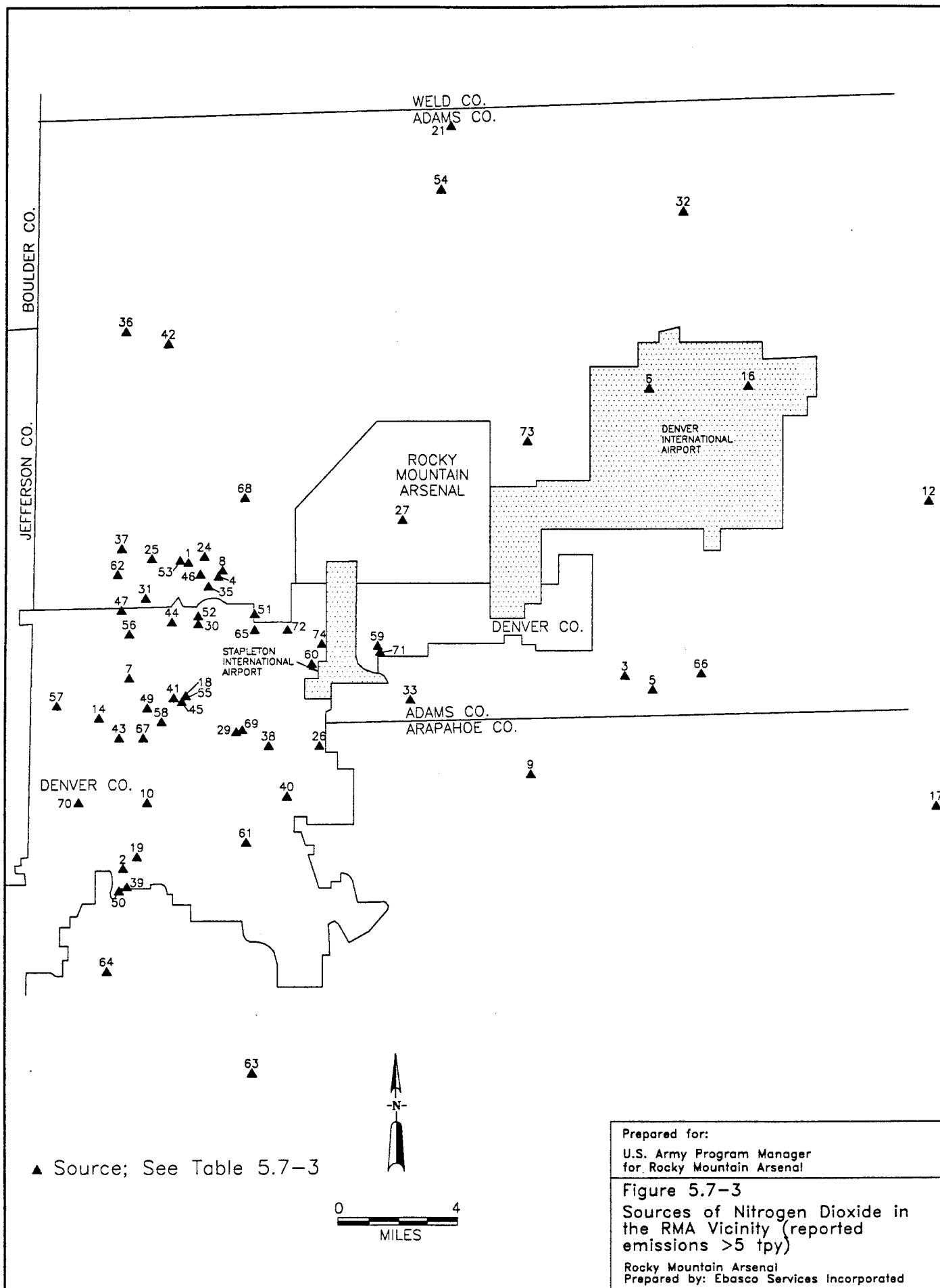
Page 2 of 2

| Facility | City | Update Map | | Estimate (tpy) |
|---|---------------|------------|-----|-------------------|
| | | Year | No. | |
| US West | No City Name | 90 | 42 | 16.4 |
| Denver and Rio Grande Western RR | Denver | 90 | 43 | 14.9 |
| Pepcol Manufacturing Company | Denver | 90 | 44 | 14.7 |
| St Joseph Hospital | Denver | 90 | 45 | 14.0 |
| National By-Products, Inc | Denver | 92 | 46 | 14.0 |
| Keebler Company | Denver | 92 | 47 | 13.5 |
| Koch Hydrocarbon Co. Texaco Station | Bennett | 92 | 48 | 13.1 |
| Radison Hotel Denver | Denver | 92 | 49 | 12.8 |
| Robinson Brick #2 | Denver | 90 | 50 | 12.3 |
| AG - Power of Colorado | Denver | 90 | 51 | 11.7 |
| U.S. West | No City Name | 90 | 52 | 11.1 |
| Rustco Products Co. | Denver | 91 | 53 | 11.0 |
| Central Products Co. | Brighton | 91 | 54 | 11.0 |
| Presbyterian Medical Center | Denver | 90 | 55 | 10.9 |
| Brannan S & G | Denver | 90 | 56 | 8.5 |
| Provenant Healthcare Partners | Denver | 92 | 57 | 8.2 |
| Swedish Medical Center | Englewood | 91 | 58 | 8.0 |
| Frito Lay Inc | Denver | 89 | 59 | 7.6 |
| City and County of Denver Stapleton Field | Denver | 90 | 60 | 7.5 |
| Rocky Mountain Dyeing & Finishing | Denver | 89 | 61 | 7.2 |
| Rocky Mountain Prestress | Denver | 92 | 62 | 7.2 |
| Metrum Information Storage | Littleton | 92 | 63 | 7.2 |
| The Electron Corporation | Littleton | 91 | 64 | 7.1 |
| Safeway Milk Plant | Denver | 92 | 65 | 7.0 |
| Western Gas Processors Ltd | Commerce City | 90 | 66 | 6.8 |
| Denver General Hospital | Denver | 88 | 67 | 6.7 |
| Bituminous Roadways of Co, Inc. | Commerce City | 92 | 68 | 6.4 |
| VA Hospital Denver | Denver | 88 | 69 | 6.3 |
| DPS - Kepner Jr | No City Name | 87 | 70 | 6.3 |
| Dow Chemical USA | Aurora | 92 | 71 | 6.3 |
| Red Seal Inc | Denver | 90 | 72 | 5.8 |
| BFI - Tower Landfill | Commerce City | 92 | 73 | 5.8 |
| United Airlines 36th & Syracuse | Denver | 89 | 74 | 5.5 |
| Vessels Oil & Gas Co. - Radar Comp. Sta. | Brighton | 92 | 75 | 5.1 |

Emission estimates obtained from EPA Aerometric Information Retrieval System







VOCs have been discussed in prior sections of this report. As noted, in the Tri-County area of Adams, Arapahoe, and Denver counties, there were a number of NO_x, CO, and SO₂ point sources and mobile sources that contributed to the RMA background air quality. As was the case with metropolitan Denver, the poorest air quality days at RMA were associated with the existence of intense ground-level inversions over the area and the subsequent development of the so-called brown cloud phenomenon. When this layer of industrial pollution drifted over RMA either directly or circuitously (after a wind shift), RMA recorded its highest levels for almost all pollutants measured. In the case of TSP, PM-10, metals, and VOCs, it was necessary to distinguish between potential local RMA sources and external sources. However, since emissions of criteria gaseous pollutants are minimal at RMA, a clear record of these incursions was provided by the meteorological and gaseous monitored data collected at RMA. Some examples are illustrated below.

5.7.1 January 14, 1993

The highest FY93 concentrations for 24-hour SO₂, 1-hour CO, and 1-hour NO_x were recorded on January 14, 1993. The maximum 24-hour SO₂ concentration was 0.015 ppm, and the maximum 1-hour CO concentration was 7.65 ppm at 2400 MST. A FY93 maximum for 8-hour CO of 4.32 ppm was recorded at 0300 MST on January 15, 1993 which included the hourly CO values from 2000 MST on January 14 through 0300 MST on January 15. Table 5.7-4 lists the air quality and meteorological data for January 14 and the first three hours of January 15.

Winds at RMA were light and blew predominantly from the south and southwest during the period as shown in the wind rose (Figure 5.7-4). Pasquill stabilities and 10m-2m temperature differences indicate that a temperature inversion was present during the morning hours of January 14, weakened during the daylight hours of the 14th, and then re-formed very strongly during the hours that the maximum 8-hour CO was recorded.

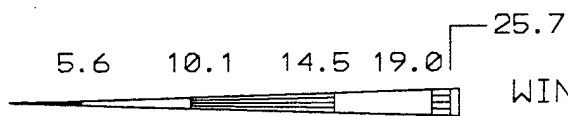
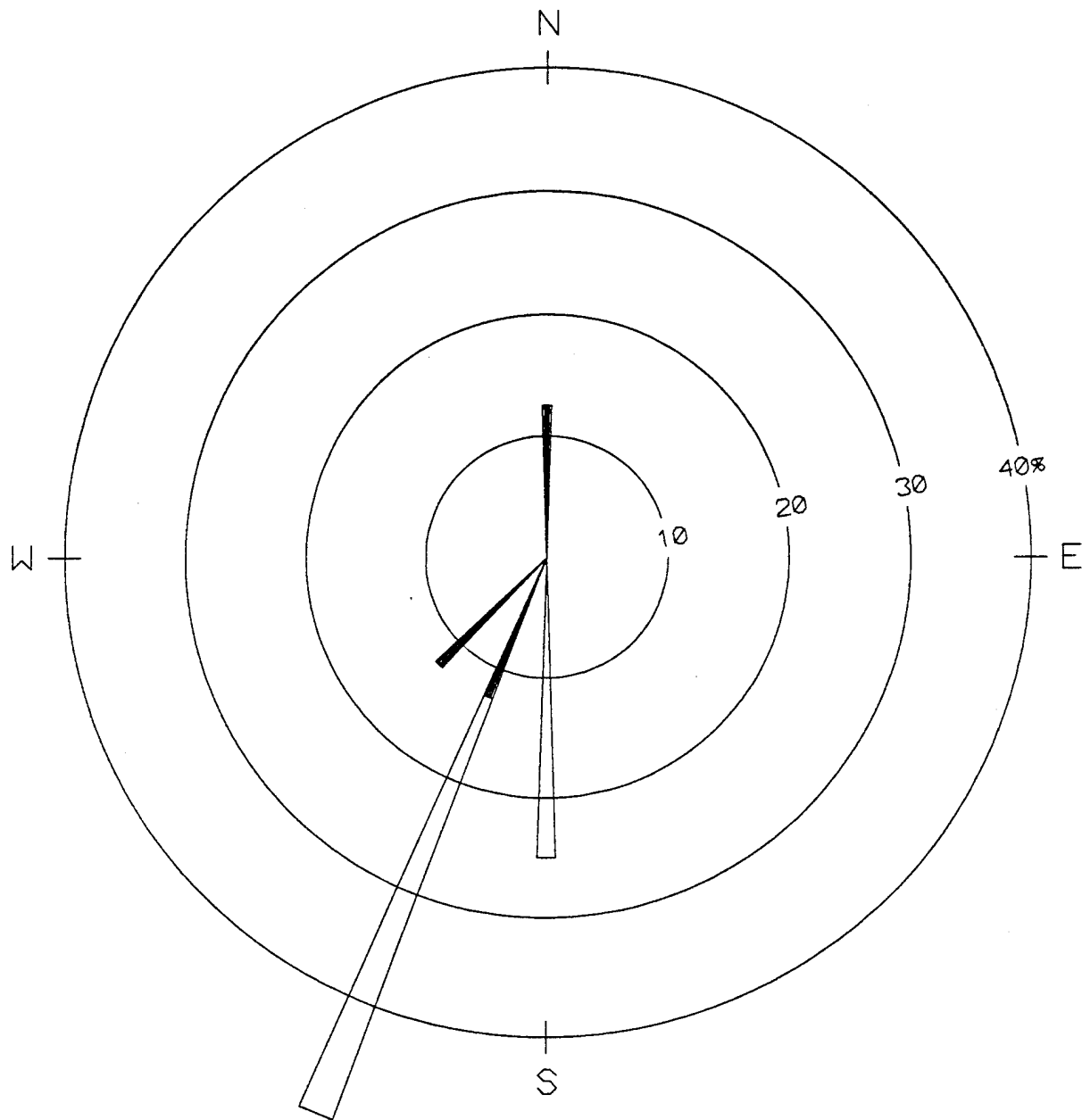
Figures 5.7-5 and 5.7-6 compare hourly CO and SO₂ recorded at RMA on January 14 with data from CDH sites and average values for the month at RMA. Although SO₂ and CO

Table 5.7-4 Air Quality and Meteorological Data for January 14-15, 1993

Page 1 of 1

| | Hour (MST) | Carbon Monoxide (ppm) | Sulfur Dioxide (ppm) | Nitrogen Oxides (ppm) | Wind Direction | Wind Speed (mph) | Temp. (°F) | Relative Humidity (%) | 10m-2m Temp. Difference (°F) | Pasquill Stability Class |
|------|---------------|-----------------------------|----------------------------|-----------------------------|-------------------|------------------------|---------------|-----------------------------|---------------------------------------|--------------------------------|
| 1/14 | 100 | 1.44 | 0.012 | 0.095 | 180.5 | 3.3 | 15.7 | 83.1 | 5.7 | D |
| | 200 | 1.59 | 0.010 | 0.107 | 106.7 | 3.5 | 15.2 | 78.3 | 4.9 | E |
| | 300 | 2.01 | 0.007 | 0.151 | 136.3 | 5.3 | 14.1 | 85.6 | 4.4 | E |
| | 400 | 1.66 | 0.012 | 0.133 | 131.8 | 4.4 | 14.9 | 84.4 | 4.2 | F |
| | 500 | 1.42 | 0.013 | 0.125 | 145.3 | 6.0 | 13.1 | 83.2 | 4.8 | E |
| | 600 | 1.73 | 0.012 | 0.149 | 230.5 | 2.8 | 14.0 | 84.0 | 4.2 | F |
| | 700 | 2.30 | 0.010 | 0.197 | 159.4 | 2.2 | 14.8 | 91.6 | 5.1 | F |
| | 800 | 2.86 | 0.010 | 0.236 | 180.1 | 4.5 | 14.1 | 91.2 | 4.6 | E |
| | 900 | 5.15 | 0.011 | 0.370 | 176.6 | 5.2 | 16.5 | 90.7 | 2.1 | D |
| | 1000 | 4.21 | 0.010 | 0.306 | 244.4 | 2.7 | 20.9 | 76.4 | 0.2 | C |
| | 1100 | 4.41 | 0.010 | 0.314 | 28.8 | 2.9 | 25.5 | 68.2 | -0.3 | B |
| | 1200 | 4.97 | 0.016 | 0.380 | 2.9 | 4.1 | 27.2 | 61.1 | -0.3 | C |
| | 1300 | 4.29 | 0.030 | 0.324 | 22.1 | 6.4 | 28.5 | 59.2 | -0.1 | D |
| | 1400 | 3.11 | 0.023 | 0.222 | 13.9 | 5.3 | 29.0 | 57.1 | -0.3 | C |
| | 1500 | 2.67 | 0.020 | 0.180 | 16.0 | 4.7 | 29.1 | 54.5 | 0.1 | B |
| | 1600 | 2.45 | 0.014 | 0.162 | 73.5 | 2.6 | 29.7 | 52.7 | 0.4 | C |
| | 1700 | 2.31 | 0.015 | 0.155 | 60.2 | 3.0 | 27.4 | 55.4 | 1.6 | D |
| | 1800 | 2.24 | 0.017 | 0.155 | 130.8 | 5.5 | 26.1 | 58.7 | 5.3 | D |
| | 1900 | 1.90 | 0.018 | 0.127 | 127.0 | 3.0 | 28.1 | 58.3 | 8.4 | E |
| | 2000 | 2.18 | 0.017 | 0.153 | 352.3 | 2.4 | 25.6 | 63.2 | 5.1 | F |
| | 2100 | 2.52 | 0.015 | 0.174 | 229.2 | 2.6 | 26.6 | 64.2 | 9.8 | F |
| | 2200 | 2.94 | 0.022 | 0.211 | 203.3 | 3.6 | 27.6 | 62.2 | 8.2 | F |
| | 2300 | 6.23 | 0.026 | 0.466 | 205.8 | 7.0 | 27.3 | 69.2 | 7.1 | E |
| | 2400 | 7.65 | 0.021 | 0.525 | 206.5 | 8.5 | 27.3 | 81.8 | 5.7 | E |
| 1/15 | 100 | 5.84 | 0.018 | 0.399 | 192.2 | 8.7 | 27.5 | 85.5 | 5.7 | E |
| | 200 | 4.88 | 0.010 | 0.333 | 172.3 | 7.9 | 27.7 | 75.2 | 6.2 | E |
| | 300 | 2.35 | 0.004 | 0.156 | 175.8 | 8.2 | 27.4 | 73.9 | 6.2 | E |

MST Mountain Standard Time
 ppm parts per million
 mph Miles per hour
 °F Degrees Fahrenheit
 % Percent
 m meters



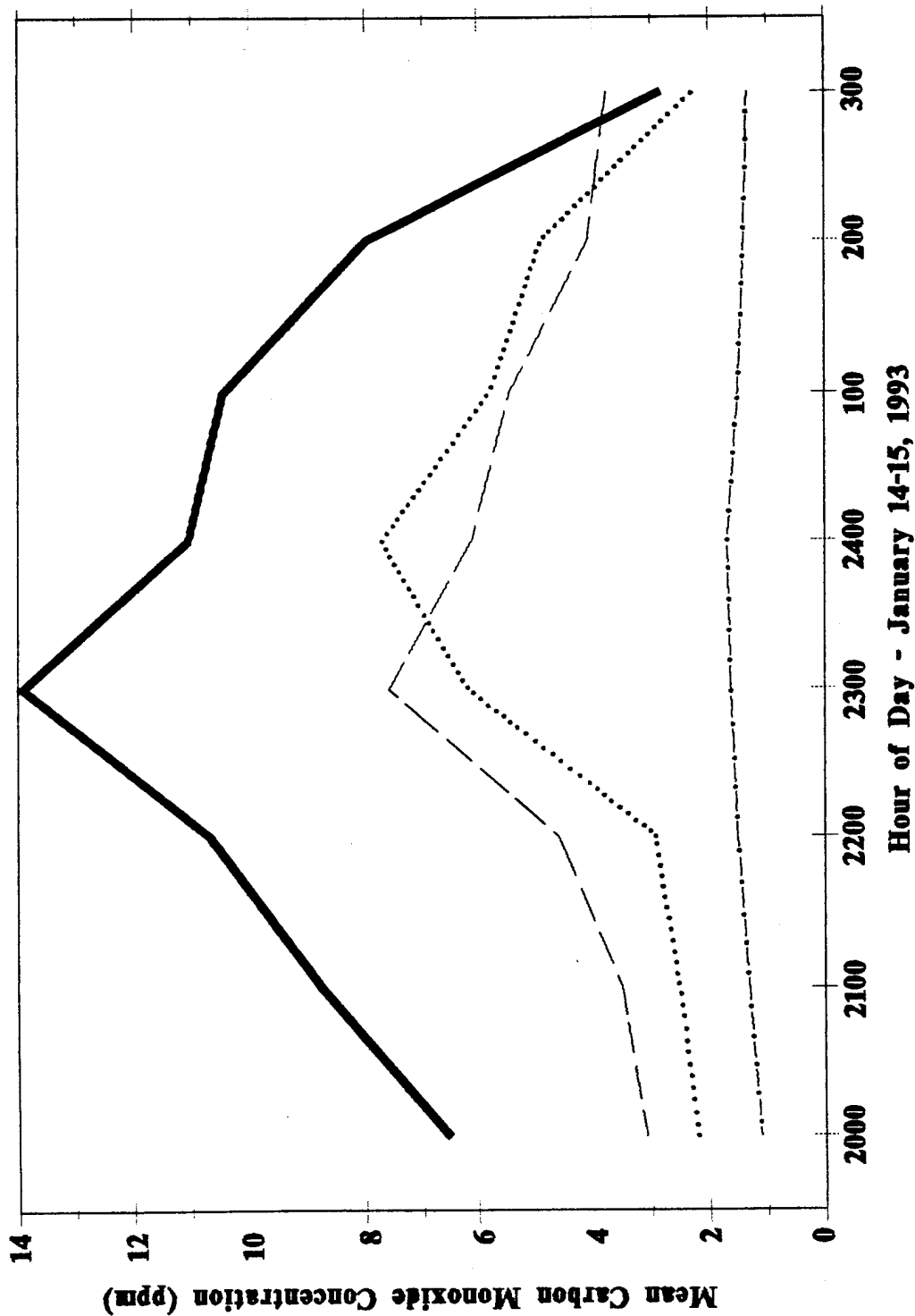
WIND SPEED CLASS BOUNDARIES
(MILES/HOUR)

Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 5.7-4

RMA Wind Rose for January
14-15, 1993 2000-0300 MST

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

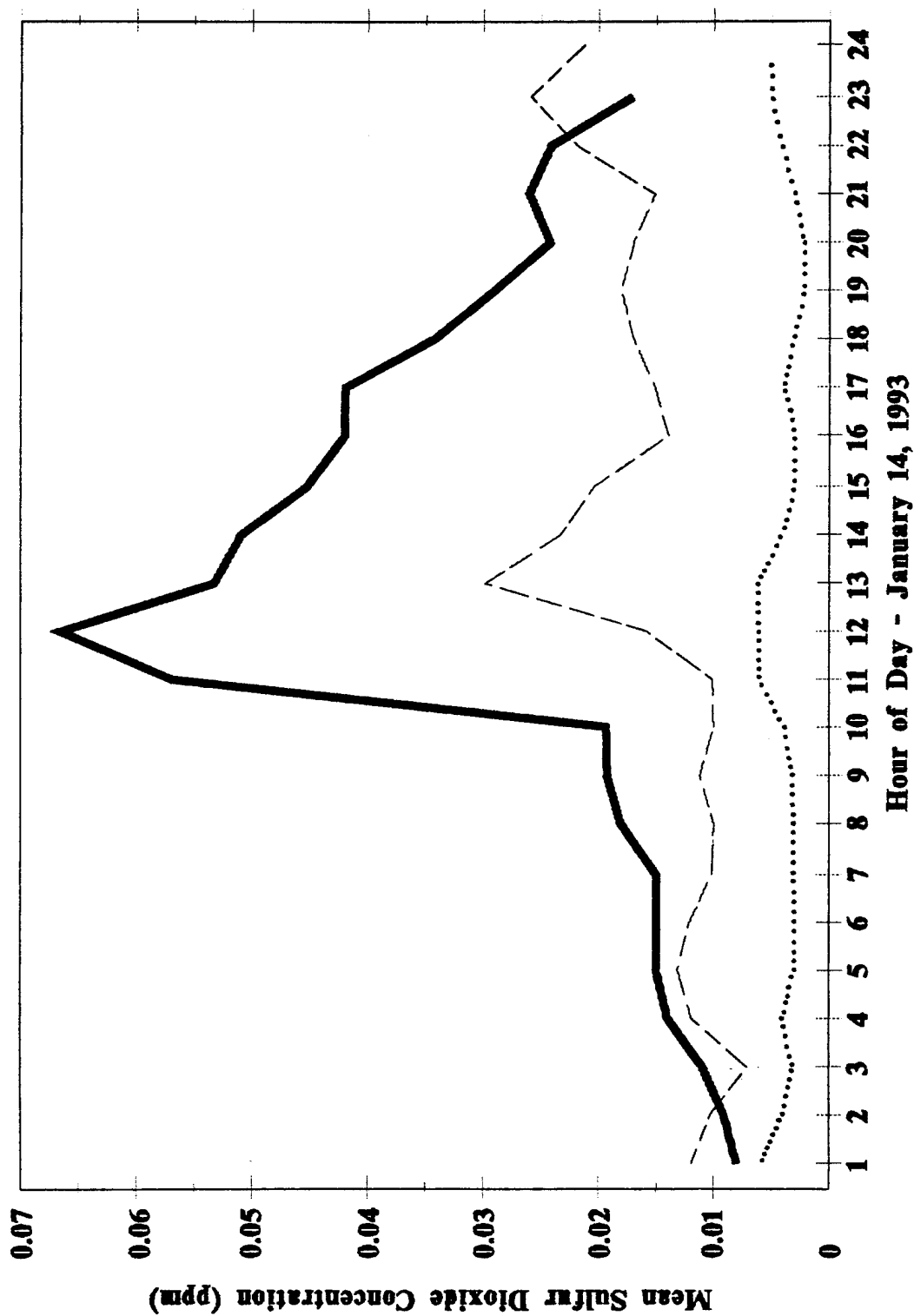


— CAMP - - - Welby RMA - . - . - RMA Jan Aves

Hour of Day - January 14-15, 1993

Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 5.7-5
Comparison of 1-Hour Carbon
Monoxide Concentrations for RMA
and CDH Sites January 14-15, 1993
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



— CAMP - - - RMA RMA Jan Aves

Prepared for:
 U.S. Army Program Manager
 for Rocky Mountain Arsenal
 Figure 5,7-6
 Comparison of 1-Hour Sulfur
 Dioxide Concentration for RMA and
 CAMP Sites January 14-15, 1993
 Rocky Mountain Arsenal
 Prepared by: Ebasco Services Incorporated

concentrations were at their highest levels of the year at RMA, the figures show that they were still much lower than the CAMP site in downtown Denver. This indicates that stagnant air and light winds were trapping pollutants over a wide area and that the likely sources of the high levels of criteria pollutants at RMA were the surrounding mobile and industrial sources.

5.7.2 December 8, 1992

High levels of CO, SO₂, and NO_x were also recorded at RMA on December 8, 1992. An hourly CO value of 6.73 ppm at 1900 MST was the second highest 1-hour value for FY93. The 24-hour SO₂ concentration of 0.015 ppm was also the second highest value for FY93. Several 1-hour values for NO_x were the highest recorded for FY93 other than on January 14. Table 5.7-5 lists the air quality and meteorological data for December 8.

Winds on this day were predominantly from the south and a persistent temperature inversion was in place (see plot of the Stapleton Airport sounding in Section 4.2). The wind rose (also in Section 4.2) shows that the predominant wind direction was due south.

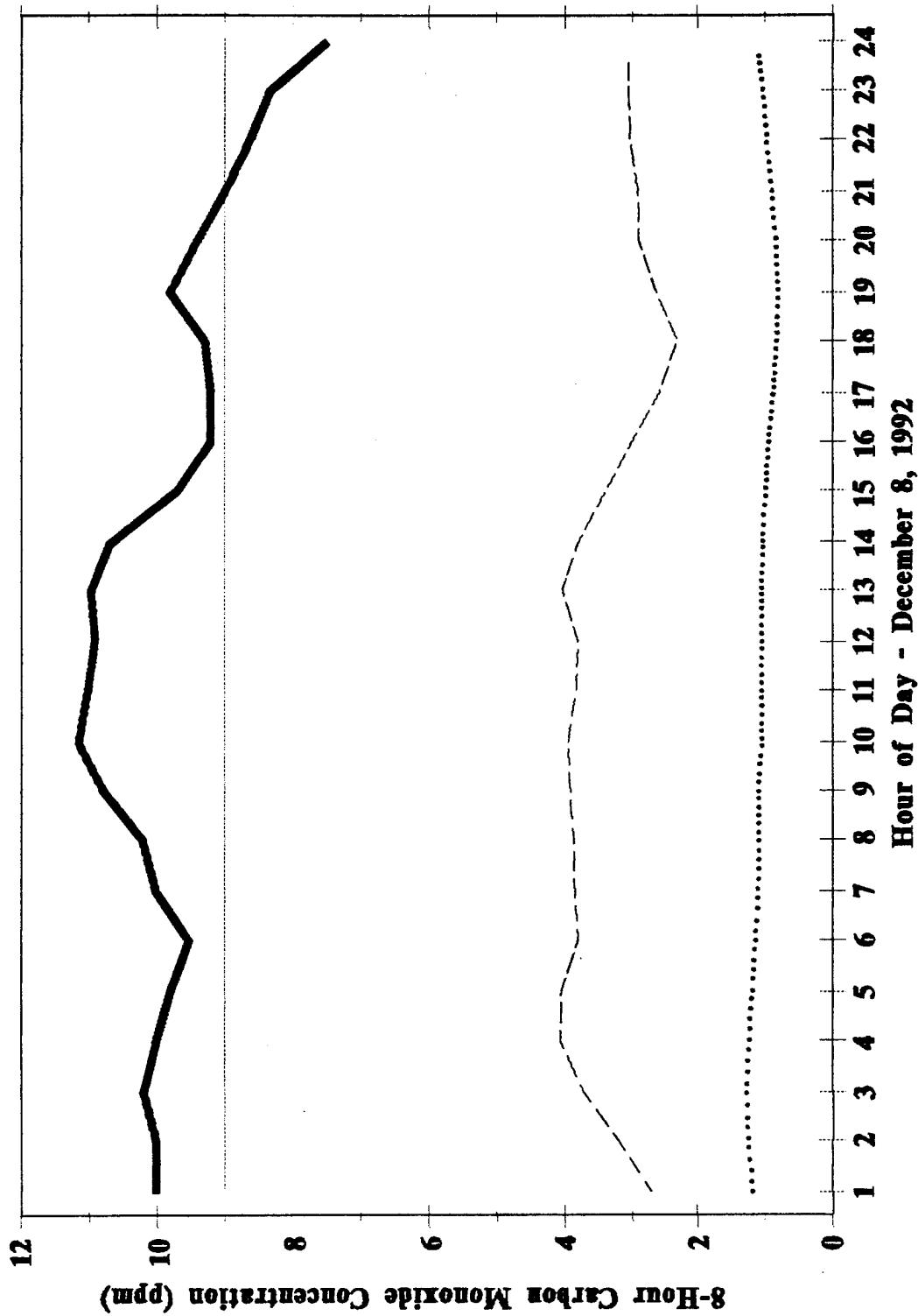
Figure 5.7-7 compares the 8-hour CO values recorded at RMA on December 8 to those recorded at the CAMP and Welby sites and average RMA concentrations for the month (Figure 5.7-8 compares the hourly SO₂ values recorded at the CAMP site to RMA). As the figures show, the CO and SO₂ levels at sites to the west and southwest of RMA were very high under the persistent inversion. The CAMP site recorded several violations of the 8-hour NAAQS for CO during the day. As with the episode of January 14, the source of the high levels of criteria pollutants at RMA on this day appears to be the mobile and industrial sources to the south and southwest.

Table 5.7-5 Air Quality and Meteorological Data for December 8, 1992

Page 1 of 1

| Hour (MST) | Carbon Monoxide (ppm) | Sulfur Dioxide (ppm) | Nitrogen Oxides (ppm) | Wind Direction | Wind Speed (mph) | Temp. (°F) | Relative Humidity (%) | 10m-2m Temp. Difference (°F) | Pasquill Stability Class |
|---------------|-----------------------------|----------------------------|-----------------------------|-------------------|------------------------|---------------|-----------------------------|---------------------------------------|--------------------------------|
| 100 | 4.65 | 0.017 | 0.338 | 179.8 | 2.5 | 25.3 | 67.4 | 9.0 | F |
| 200 | 4.95 | 0.018 | 0.407 | 315.5 | 3.5 | 23.9 | 74.8 | 5.7 | F |
| 300 | 5.26 | 0.021 | 0.423 | 313.6 | 3.2 | 23.9 | 77.4 | 6.4 | F |
| 400 | 3.97 | 0.019 | 0.314 | 17.0 | 4.4 | 22.9 | 73.2 | 5.2 | F |
| 500 | 1.75 | 0.009 | 0.126 | 320.6 | 2.6 | 23.8 | 67.4 | 3.4 | F |
| 600 | 2.48 | 0.011 | 0.184 | 246.4 | 4.0 | 23.7 | 69.8 | 6.4 | F |
| 700 | 4.03 | 0.019 | 0.359 | 185.1 | 6.8 | 23.7 | 67.6 | 6.4 | E |
| 800 | 3.85 | 0.036 | 0.298 | 182.7 | 7.0 | 25.2 | 53.2 | 6.9 | D |
| 900 | 4.74 | 0.037 | 0.303 | 175.1 | 6.7 | 26.8 | 48.9 | 4.8 | D |
| 1000 | 5.34 | 0.034 | 0.332 | 181.9 | 7.8 | 29.4 | 47.5 | 3.1 | D |
| 1100 | 4.39 | 0.022 | 0.275 | 169.7 | 5.6 | 34.4 | 43.7 | 1.7 | C |
| 1200 | 3.66 | 0.017 | 0.239 | 280.6 | 3.5 | 37.0 | 40.3 | 0.9 | B |
| 1300 | 3.58 | 0.014 | 0.259 | 8.7 | 8.7 | 31.2 | 48.6 | 0.2 | C |
| 1400 | 0.71 | 0.001 | 0.051 | 16.2 | 3.6 | 29.9 | 45.7 | 0.4 | B |
| 1500 | 0.55 | 0.001 | 0.042 | 177.8 | 3.5 | 32.8 | 36.9 | 0.5 | A |
| 1600 | 0.78 | 0.003 | 0.054 | 187.1 | 4.3 | 35.1 | 33.1 | 1.9 | B |
| 1700 | 1.57 | 0.010 | 0.132 | 197.4 | 6.5 | 36.9 | 25.5 | 5.7 | C |
| 1800 | 3.34 | 0.018 | 0.253 | 174.3 | 6.5 | 36.2 | 38.9 | 7.9 | D |
| 1900 | 6.73 | 0.019 | 0.442 | 167.4 | 5.7 | 34.3 | 50.9 | 8.4 | E |
| 2000 | 5.75 | 0.014 | 0.369 | 192.8 | 8.2 | 34.6 | 48.5 | 6.7 | E |
| 2100 | 3.44 | 0.009 | 0.237 | 187.5 | 6.9 | 32.7 | 46.5 | 6.5 | D |
| 2200 | 1.87 | 0.005 | 0.138 | 192.2 | 10.3 | 32.3 | 46.5 | 4.3 | E |
| 2300 | 0.85 | 0.001 | 0.058 | 195.2 | 13.3 | 33.8 | 46.2 | 3.4 | D |
| 2400 | 0.62 | 0.001 | 0.043 | 197.9 | 12.8 | 33.9 | 46.9 | 4.7 | D |

MST Mountain Standard Time
 ppm parts per million
 mph Miles per hour
 °F Degree Fahrenheit
 % Percent
 m Meters

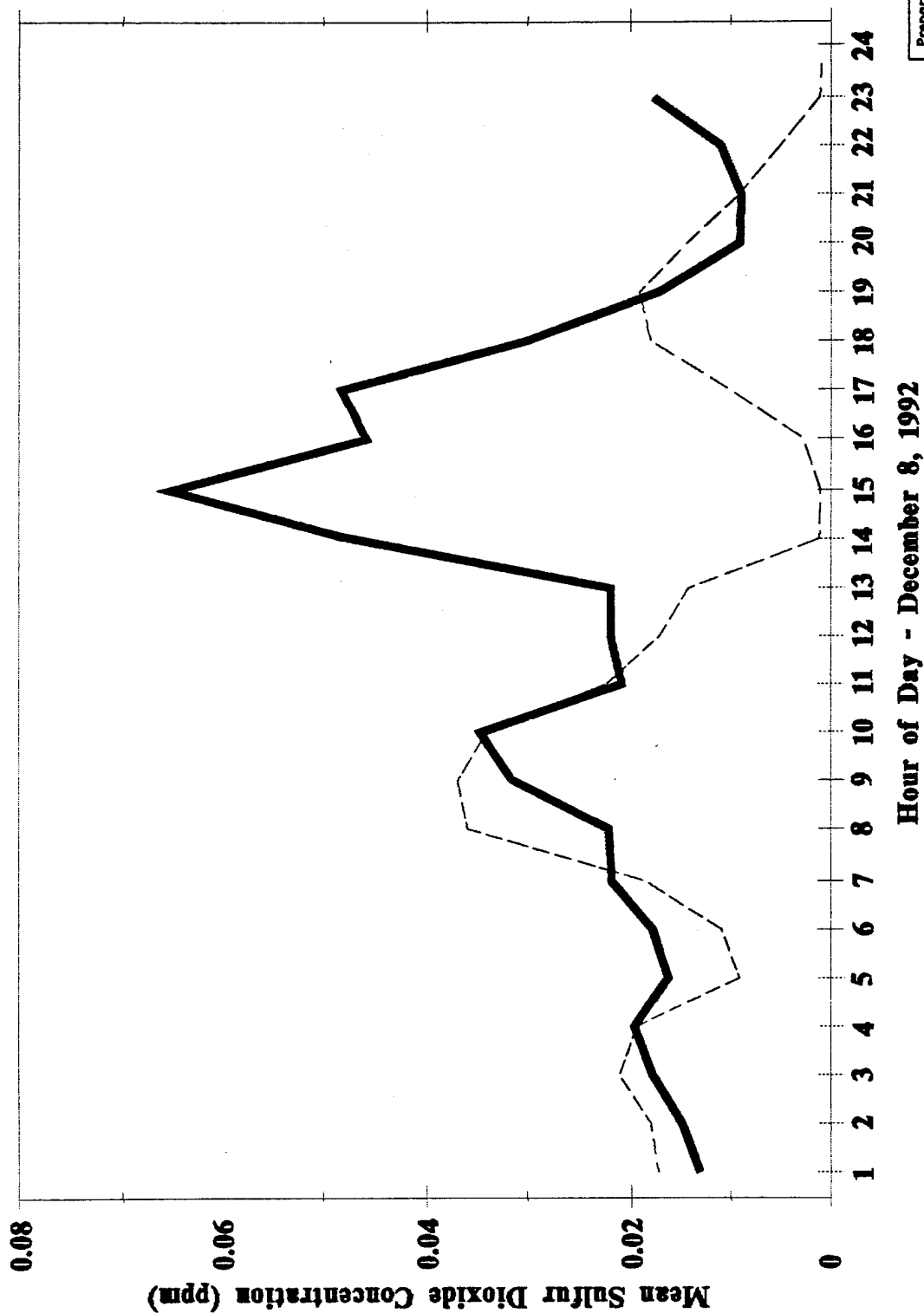


— CAMP --- RMA RMA Dec Avg — NAAQS

Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 5.7-7

Comparison of 8-Hour Carbon
Monoxide Concentration for RMA
and CAMP Site December 8, 1992
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 5.7-8

Comparison of 1-Hour Sulfur
Dioxide Concentration for RMA
and CAMP Site December 8, 1992

Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

— CAMP --- RMA

6.0 METEOROLOGICAL MONITORING AND DISPERSION MODELING PROGRAMS

6.1 METEOROLOGICAL PROGRAM OVERVIEW

The meteorological program and station locations are described in Section 3.5. The locations of meteorological monitoring sites are shown in Figure 3.5-1. Complete listings of all data collected and used in this report are provided in Appendix J. Summaries of wind speed, wind direction, temperature, relative humidity, barometric pressure, solar radiation, precipitation, and atmospheric stability are provided in the following sections.

6.1.1 Program Objectives

The meteorological assessment has several objectives. The first is to identify the atmospheric conditions associated with typical contaminant levels or any abnormally high levels that might result from existing sources and remedial activity at RMA. Prevailing wind direction, wind speed, peak wind gusts, temperature, and precipitation all influence the release and spread of atmospheric emissions. The meteorological database has been established to identify typical relationships and to examine seasonal and diurnal patterns. For example, gusty winds will often result in higher levels of total suspended particles, inhalable particles of less than 10 micrometers, metals, and possibly semivolatile organic compounds. Certain pollutant emissions may vary with the seasons, which would be an important factor for remediation planning. Also, diurnal influences (such as the formation of a drainage wind pattern during evening and early morning inversion periods) may influence the levels or spatial distribution of certain pollutants. Knowledge of these patterns is useful in assessing the potential movement of contaminants during remedial activities.

The meteorological program also supports air quality model applications that are used in pollution dispersion evaluations and predictions. These analyses are employed for assessing impacts beyond the RMA boundary, and for predicting real-time pollution levels during remedial activities.

6.1.2 Data Recovery

Details of the recovery of FY93 meteorological data for each parameter of the composite database are provided in Table 6.1-1. Recoveries are based on the total number of hours of possible (8,760) data during FY93. All observations (hourly values) were considered invalid or missing if there were less than 45 minutes of valid recorded data for that parameter.

6.1.3 Databases

A single, representative composite database was developed from data collected at the RMA meteorological sites. Meteorological Station 4 (M4) was the primary source for wind speed, wind direction, sigma theta (standard deviation of the horizontal wind direction fluctuations), temperature, precipitation, and barometric pressure. The remaining parameters in the composite database were taken from other sites. Relative humidity and the 10-meter minus 2-meter temperature difference were taken from Meteorological Station 1 (M1), and solar radiation was taken from Meteorological Station 3 (M3). In the event that any of the data from M4 were invalid, data from the other sites were substituted. These substitutions amounted to less than 0.5 percent of the FY93 database. It should be noted that temperature difference measurements at M1 were for purposes of relative indications as opposed to precise measurements. The M1 temperature sensors at each level meet PSD specifications as independent indicators, but are not intended to meet PSD requirements for delta-temperature systems.

Separate databases were also maintained for each site and were used for site comparisons within the RMA boundary to note any areal differences (see FY92 data report for comparison between RMA meteorological stations [WCC, 1993]). All summaries shown in this report were taken from the FY93 composite database. All data have been thoroughly checked for quality. Long-term climatological means for Denver Stapleton International Airport (Section 2.2) were used for comparison. In some cases, FY93 CAQMMP data were compared to FY88-FY92 CMP/CAQMMP data.

Table 6.1-1 Summary of Rocky Mountain Arsenal Meteorological Monitoring Data
Recovery for FY93 Page 1 of 1

| Parameter | Annual Recovery % |
|----------------------------------|----------------------|
| Wind Speed | 100.0 |
| Wind Direction | 100.0 |
| Sigma Theta | 100.0 |
| Temperature | 100.0 |
| Relative Humidity | 99.9 |
| Barometric Pressure | 100.0 |
| Solar Radiation | 99.9 |
| Precipitation | 100.0 |
| Maximum Gust | 100.0 |
| Temperature Difference (10m -2m) | 99.9 |

m meters
% Percent

6.2 SUMMARY OF RESULTS

A summary of monthly and annual meteorological data for FY93 at RMA is provided in Table 6.2-1. All parameters summarized in this table are based on hourly averaged values with the exception of maximum gust, which is the maximum instantaneous wind speed for the period. Maximum speed is the maximum 1-hour average wind speed value for the period. The predominant wind direction indicates the sector (of 16 possible sectors) from which the wind is blowing most frequently during the reported period. Temperature difference is the difference in temperature between the 10-meter and 2-meter sensors located on the 10-meter tower.

6.2.1 Temperature

The RMA FY93 annual mean temperature of 47.8°F was below the long-term Denver average of 50.3°F. The monthly mean temperatures were cooler than average for all but 2 of the 12 months. The months of November through February showed the greatest deviation from normal as all were 4°F to 7°F cooler than average. Only the months of October and March were warmer than average. The maximum monthly average was 70.9°F in July, and the minimum monthly average was 25.4°F in January. The maximum temperature, 94.2°F, was recorded in July, and the minimum temperature, -8.7°F, was recorded in February.

6.2.2 Relative Humidity

The annual mean relative humidity for FY93 was above normal at 58.3 percent. The monthly average relative humidities followed the typical pattern, with higher values in the winter and spring due to lower temperatures and increased amounts of precipitation, and lower values in the summer and fall. The maximum monthly average humidity was 72 percent in February, and the minimum monthly average was 44 percent in October.

6.2.3 Barometric Pressure

The annual mean station barometric pressure for FY93 was 24.75 inches of mercury (in. Hg), which was above the long-term normals. The maximum monthly average was 24.84 in. Hg in August, and the minimum monthly average was 24.66 in. Hg in April.

Table 6.2-1 Summary of Rocky Mountain Arsenal Monthly Meteorological Conditions for FY93 (October 1, 1992 Through September 30, 1993) Page 1 of 2

| | Temperature (°F) | | | | | Mean Meteorological Data | | | Precipitation (inches) | | | |
|----|------------------|------|---------|---------|------|--------------------------|---------------------------|-------------------------------|------------------------|-------------|---|------|
| | Month | Avg | Extreme | Extreme | Mean | Relative Humidity | Station Pressure (in. Hg) | Daily Solar Radiation (Ly/Hr) | Total | Max. 24-Hr. | # of days with measurable precipitation | |
| | | Max. | Min. | Max. | | | | | | | | Min. |
| 1 | Oct | 66.0 | 41.2 | 82.5 | 23.0 | 53.1 | 43.6 | 24.82 | 0.23 | 0.18 | 0.10 | 3 |
| 2 | Nov | 43.4 | 25.5 | 66.6 | 9.9 | 33.7 | 64.4 | 24.75 | 0.14 | 0.30 | 0.18 | 3 |
| 3 | Dec | 36.2 | 15.5 | 52.8 | -2.0 | 25.9 | 68.4 | 24.69 | 0.14 | 0.09 | 0.06 | 3 |
| 4 | Jan | 35.1 | 17.7 | 57.4 | -2.2 | 25.4 | 72.2 | 24.72 | 0.15 | 0.02 | 0.02 | 1 |
| 5 | Feb | 37.1 | 20.0 | 56.7 | -8.7 | 27.8 | 72.3 | 24.72 | 0.20 | 0.26 | 0.07 | 6 |
| 6 | Mar | 51.5 | 31.3 | 71.4 | 7.8 | 41.3 | 55.1 | 24.75 | 0.28 | 0.20 | 0.08 | 4 |
| 7 | Apr | 56.8 | 36.0 | 76.8 | 27.5 | 46.6 | 54.8 | 24.66 | 0.35 | 1.24 | 0.52 | 6 |
| 8 | May | 67.6 | 46.9 | 81.4 | 33.9 | 56.8 | 56.8 | 24.74 | 0.39 | 1.56 | 0.85 | 11 |
| 9 | Jun | 76.6 | 53.1 | 92.2 | 41.4 | 65.4 | 50.0 | 24.70 | 0.44 | 1.43 | 1.11 | 6 |
| 10 | July | 84.1 | 58.4 | 94.2 | 51.2 | 70.9 | 50.2 | 24.74 | 0.41 | 0.60 | 0.15 | 10 |
| 11 | Aug | 78.5 | 57.3 | 89.8 | 48.0 | 67.7 | 55.9 | 24.84 | 0.33 | 1.30 | 0.73 | 9 |
| 12 | Sept | 70.7 | 48.0 | 87.6 | 32.0 | 58.9 | 55.6 | 24.83 | 0.30 | 1.03 | 0.34 | 8 |
| | Year | 58.6 | 37.6 | 94.2 | -8.7 | 47.8 | 58.3 | 24.75 | 0.28 | 8.21 | 1.11 | 70 |

°F Degrees Fahrenheit
in. Hg Inches of mercury
Ly/Hr Langleys per hour

Table 6.2-1 Summary of Rocky Mountain Arsenal Monthly Meteorological Conditions for FY93 (October 1, 1992 through September 30, 1993) Page 2 of 2

| Wind Speed (mph) | | | | | Atmospheric Stability Category (% occurrence) | | | | | | |
|------------------|---------------|------------------|-----------------|-------------------------------|--|------|------|------|------|------|-----|
| Month | Mean Speed | Maximum Speed | Maximum Gust | Predominant Wind Direction | Average Temp Diff | A | B | C | D | E | F |
| 1 Oct | 8.0 | 35.7 | 46.7 | SSW | 2.1 | 8.1 | 9.4 | 12.4 | 38.0 | 26.9 | 5.2 |
| 2 Nov | 7.5 | 33.0 | 53.1 | SSW | 2.0 | 2.9 | 5.8 | 10.1 | 44.3 | 31.5 | 5.3 |
| 3 Dec | 7.7 | 29.4 | 39.2 | SSW | 2.8 | 1.7 | 5.1 | 9.5 | 48.5 | 26.1 | 9.0 |
| 4 Jan | 6.9 | 31.7 | 46.2 | SSW | 2.2 | 1.5 | 3.8 | 8.3 | 49.3 | 27.3 | 9.8 |
| 5 Feb | 6.9 | 27.6 | 40.4 | South | 1.4 | 3.0 | 4.9 | 10.4 | 46.3 | 30.1 | 5.4 |
| 6 Mar | 9.3 | 31.2 | 62.9 | South | 1.1 | 5.4 | 7.1 | 12.1 | 47.3 | 25.0 | 3.1 |
| 7 Apr | 9.4 | 33.5 | 50.3 | South | 0.9 | 7.4 | 11.4 | 11.3 | 47.5 | 19.4 | 3.1 |
| 8 May | 7.9 | 30.3 | 51.1 | South | 1.1 | 9.0 | 10.6 | 17.1 | 41.1 | 18.3 | 3.9 |
| 9 Jun | 8.9 | 28.7 | 47.0 | SSW | 0.8 | 10.6 | 12.6 | 16.1 | 42.8 | 15.4 | 2.5 |
| 10 July | 9.7 | 35.7 | 53.2 | South | 0.6 | 8.3 | 10.5 | 15.3 | 50.5 | 13.7 | 1.6 |
| 11 Aug | 8.0 | 25.2 | 40.6 | South | 0.8 | 9.8 | 12.0 | 16.5 | 36.4 | 23.1 | 2.2 |
| 12 Sept | 8.6 | 30.5 | 60.6 | SSW | 1.0 | 7.9 | 10.0 | 12.4 | 44.2 | 23.5 | 2.1 |
| Year | 8.2 | 35.7 | 62.9 | SSW | 1.4 | 6.3 | 8.6 | 12.6 | 44.7 | 23.3 | 4.4 |

Legend: Maximum Speed Maximum 1-hr average wind speed value for the month
Maximum Gust Maximum instantaneous wind speed for the month
Temp Diff Temperature difference (10-meter minus 2-meter) on the tower
SSW South-southwest

A = Extremely Unstable
B = Unstable
C = Slightly Unstable
D = Neutral
E = Stable
F = Extremely Stable

6.2.4 Solar Radiation

The solar radiation statistics include evening values to show the increase in solar radiation in the spring and summer due to the length of the days. The annual mean solar radiation for FY93 was 0.28 langleys per hour (ly/hr). Values were higher in the summer and lower in the winter, as expected. The maximum monthly average was 0.44 ly/hr in June, and the minimum monthly average was 0.14 ly/hr in November and December.

6.2.5 Precipitation

Precipitation for FY93 totaled 8.21 inches at RMA, compared with the Denver climatological mean of 15.31 inches. Every month during FY93 was drier than long-term Denver averages. The heaviest monthly precipitation total (1.56 inches) was measured during May, which historically receives the heaviest precipitation. The driest month was January (0.02 inches).

6.2.6 Winds

Hourly mean wind speeds for FY93 averaged 8.2 mph compared to the Denver long-term mean value of 8.7 mph. This comparison is important because pollutant levels, especially those associated with dust particles (i.e., TSP, PM-10, metals, and SVOCs), are sometimes generated by high wind speeds. The data show that FY93 mean wind speeds were below average for November through May. The maximum monthly average was 9.7 mph in July, and the minimum monthly average was 6.9 mph in January and February. The maximum gust, 62.9 mph, was recorded in March. The predominant wind direction for RMA in FY93 was south-southwest; while in Denver, winds are predominantly from due south.

The 5-year (1988-1992) wind rose for Stapleton Airport and the RMA wind rose for FY93 are shown in Figures 2.2-1 and 3.2-1, respectively. The FY93 wind rose showed predominantly southerly and south-southwesterly flows as compared to the predominant flow at Stapleton Airport of south to north. The secondary maximum wind direction at RMA was north-northeasterly while the Stapleton secondary maximum was from the north.

6.2.7 Atmospheric Stability

Atmospheric stability is an important factor in the dispersion of air pollutants. Along with wind speed and wind direction, it is a key parameter in air quality dispersion models and reflects the potential for the atmosphere to disperse pollutants horizontally and vertically. Hourly stabilities for RMA are assigned one of six possible Pasquill stability categories (A through F) with "A" being the most unstable and "F" being the most stable. The hourly categories are determined by a combination of wind speed, sigma theta, and the time of day of the observations. Table 6.2-1 shows the percent occurrence for all atmospheric stability categories for the FY93 RMA composite data set. A joint frequency distribution for FY93 is provided in Appendix L.

Unstable categories A, B, and C indicate active dispersion and occur primarily during midday and afternoon periods, and most often in the warmer seasons. These categories were measured 6.3 percent, 8.6 percent, and 12.6 percent of the time, respectively, or a total of 27.5 percent of the time. Stable categories E and F, reflect poor dispersion and occur during morning hours and more often in the colder seasons. These stability categories were measured 23.4 percent and 4.4 percent of the time, respectively, or a total of 27.8 percent of the time. The remainder of the hours, 44.7 percent, were in Category D, which reflects neutral stability. Hours with high mean wind speeds are categorized as being neutral in stability, but as mentioned earlier, strong winds also have the potential to increase dust-associated particles emissions.

Another indicator of stability is the temperature difference between the 10-meter and 2-meter levels. Greater average temperature differences were recorded in the colder seasons (more stable), compared to smaller differences in the summer (more unstable or neutral).

The stability data for FY93 appear to be typical for the Denver area. The inversion conditions associated with poor dispersion categories E and F were less frequent during the spring and summer, and occurred primarily in the evening and early morning periods. During the winter, there were many inversion periods that lasted continuously for several days and intensified the

brown cloud over the Denver area. During spring and summer, the dispersion potential was more typically bimodal, with excellent dispersion during the day and poor dispersion at night.

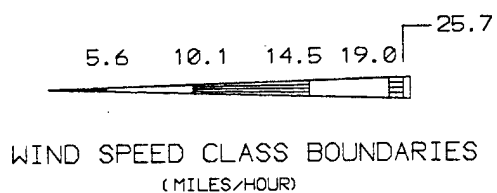
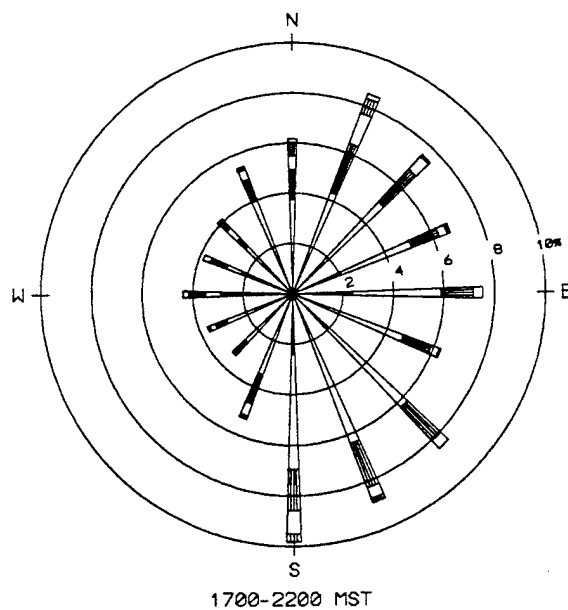
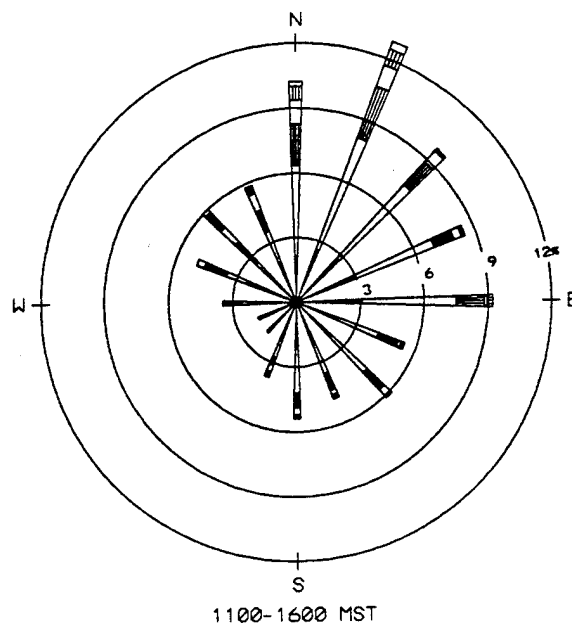
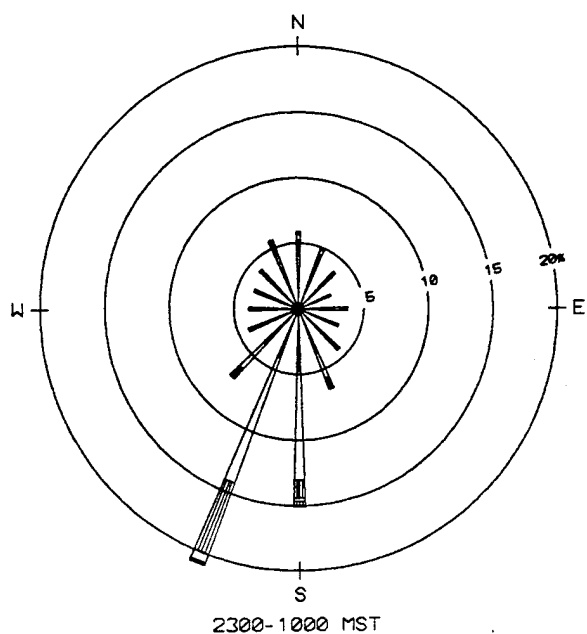
6.3 SEASONAL AND DIURNAL INFLUENCES

Dispersion characteristics are closely related to diurnal cycles in atmospheric stability and wind patterns. During the day, when dispersion conditions are good (categories A, B, and C), winds are highly variable and frequently gusty. At night, when the inversions set in and dispersion is poor (categories E and F), winds are generally light and follow a drainage pattern from the south-southwest to the north-northeast. This diurnal pattern is illustrated in Figure 6.3-1. The 2300-1000 MST wind rose shows the persistent, light winds from the south and south-southwest in the evening and morning hours, while the 1100-1600 MST wind directions are more variable in the northeast quadrant with stronger wind speeds. The 1700-2200 MST pattern is transitional from the afternoon pattern to the evening/morning pattern.

Figure 6.3-2 shows that the wind patterns in FY93 did not vary significantly from season to season. Historically, this has been typical of both RMA and Denver Stapleton Airport (see previous RMA annual reports for detailed comparisons between RMA and Stapleton seasonal and diurnal patterns). The distribution of stability categories at RMA does vary by season, however. For example, during FY93, unstable category A occurred 10.6 percent of the time in June but only 1.5 percent of the time in January, while stable category F occurred 9.8 percent of the time in January, but only 1.6 percent of the time in July.

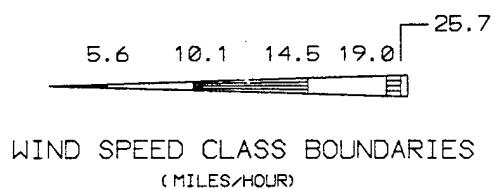
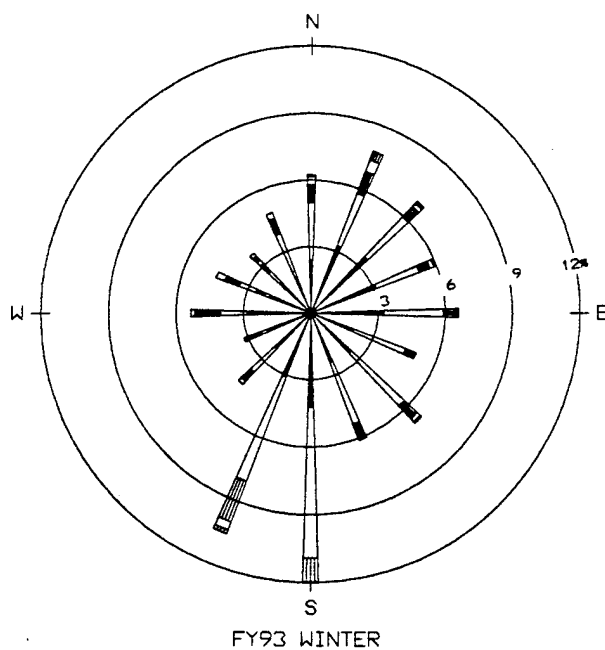
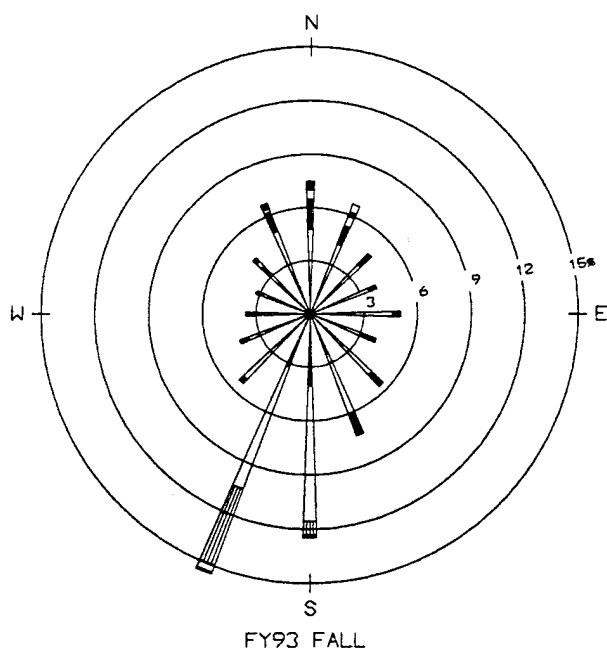
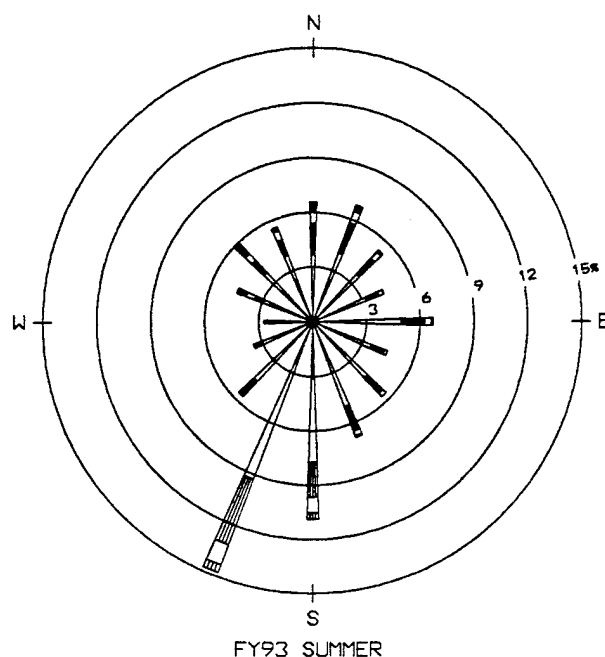
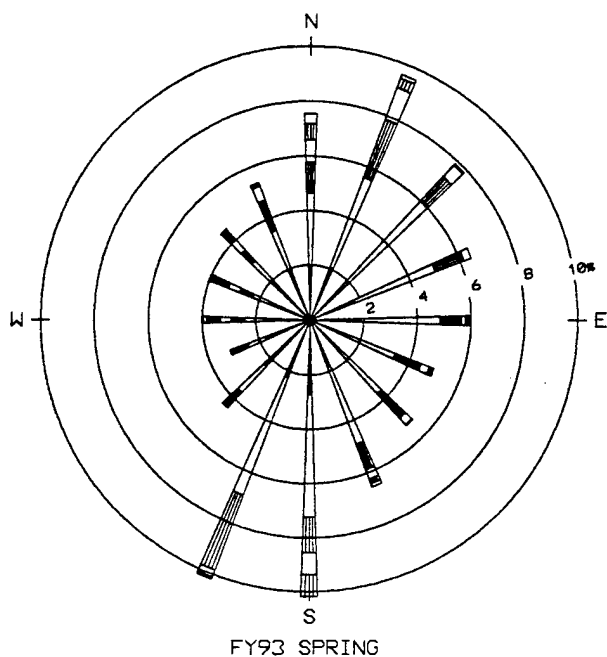
6.4 SUMMARY AND CONCLUSIONS

Although mean weather patterns may vary from year to year, it appears that the FY93 meteorological data, with minor differences, were representative of long-term data and provide a useful instrument for assessing contamination impacts during this period and for future remedial progress evaluations. Table 6.4-1 provides a comparison of FY93 meteorological parameters with those of FY89 through FY92. The only differences of note were: the predominant wind direction was southerly in FY89, FY90, FY91, and FY92, and south-southwesterly in FY93; and the solar



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 6.3-1
Diurnal Variation of RMA
Winds for FY93
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



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U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 6.3-2
Seasonal Variation of RMA
Winds for FY93
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

Table 6.4-1 FY89 - FY93 Meteorological Comparison

Page 1 of 1

| | FY93 | FY92 | FY91 | FY90 | FY89 |
|--|------|-------|-------|-------|-------|
| Average Maximum Temperature (°F) | 58.6 | 60.8 | 61.0 | 62.0 | 61.7 |
| Average Minimum Temperature (°F) | 37.6 | 39.7 | 39.1 | 39.9 | 38.6 |
| Mean Temperature (°F) | 47.8 | 50.0 | 49.5 | 50.6 | 49.9 |
| Mean Relative Humidity (%) | 58.3 | 54.3 | 49.2 | 52 | 49 |
| Mean Total Daily Solar Radiation (Ly/Hr) | 0.28 | 0.29 | 0.33 | 0.29 | 0.24 |
| Total Precipitation (in) | 8.21 | 13.09 | 16.16 | 11.56 | 12.41 |
| Total Number of Days of Precipitation | 70 | 82 | 68 | 67 | 77 |
| Mean Wind Speed (mph) | 8.2 | 8.1 | 8.7 | 8.6 | 8.5 |
| Maximum Hourly Wind Speed (mph) | 35.7 | 37.7 | 38.4 | 35.9 | 33.2 |
| Maximum Wind Gust (mph) | 62.9 | 69.4 | 62.0 | 52.8 | 72.5 |
| Predominant Wind Direction | SSW | SOUTH | SOUTH | SOUTH | SOUTH |

°F Degrees Fahrenheit
 Ly/Hr Langley's per hour
 in. inches
 mph miles per hour
 % Percent

radiation level of 0.24 ly/hr in FY89 was much lower than the values of 0.29 ly/hr in FY90 and FY92, 0.33 ly/hr in FY91, and 0.28 ly/hr in FY93. The solar radiometer was out of operation for a period during the summer of 1989, which may account for the annual differences for that parameter.

The data also suggest important guidelines for mitigation that may be appropriate during remedial activities. Emissions from daytime activities would contribute lesser impacts than evening and early morning activities because of the stronger dispersion conditions during the day. At night, inversions are prevalent and dispersion conditions are poor. Also, winds at night are channelled north and northeast of the remediation source. The distance from a remediation source to sensitive areas beyond RMA boundaries should also be taken into account.

In this section, typical impacts were inferred based on seasonal and diurnal meteorology. For real-time applications, or for short-term and long-term assessments, these influences must be incorporated into atmospheric dispersion models along with detailed source emission characteristics. Model approaches employed in the present report and recommended for future remedial progress evaluations are discussed in Section 6.5.1.

6.5 ATMOSPHERIC DISPERSION MODEL

6.5.1 Model Approaches

One of the objectives of the CAQMMP is to identify atmospheric conditions that may trigger high contamination levels and may require special precautions and mitigating actions. A related task is the development and/or application of air dispersion prediction techniques and models that use available site data and provide forecasts of potential contamination impacts. The Basin F program developed and applied such modeling techniques as an integral part of its program to provide for the health and safety of workers and the general public. The CMP/CAQMMP closely coordinated these operational activities and incorporated dispersion model results into previous reports as well as the present report. These techniques appear to be effective and are

recommended for future remedial and air quality assessment applications. Specific models used in remedial operations and in data assessments are discussed below.

The models employed were the EPA Industrial Source Complex (ISC) Model (EPA 1986a) and the PUFF Advection Model (INPUF) (EPA 1986b). These are standard and approved EPA models used for general purposes to predict air quality impacts. They are often used in environmental impact statements and air quality permit applications. For Basin F remedial activities, they were modified to support the cleanup operations, and in particular, to use real-time air quality and meteorological data.

Because precise measurements of source emissions could not be determined, a unique approach was devised to use X/Q values, in conjunction with measured ambient concentrations, to predict short-term (operational) and longer-term (assessment) impacts. The X/Q values, as noted in previous discussions relating to the CMP, CAQMMP, and Basin F monitoring results, do not indicate ambient concentrations, but rather, indicate relative strength or potential concentration levels, based on meteorological conditions and on an undetermined emission source strength. In the Basin F program, the X/Q values were calibrated with real-time monitoring data to predict ambient concentrations in support of the remedial operations. In Gaussian dispersion models, the ground-level concentrations of contaminants in a plume can be computed as follows:

$$X = \frac{Q}{\pi \sigma_y \sigma_z u} \cdot \exp \left[-1/2 \left[\frac{ESH}{\sigma_z} \right]^2 \right]$$

Where X = predicted ground-level concentration (g/m³)
 Q = source emissions (g/s)
 σ_y = horizontal dispersion coefficient (function of distance from source) (m)
 σ_z = vertical dispersion coefficient (function of distance from source) (m)
 u = mean wind speed at plume level (m/s)
 ESH = effective stack height, or plume height level assumed for Basin F
 application (m)

When the source emission term is not determined, the equation can be written as follows:

$$X/Q = \frac{1}{\pi \sigma_y \sigma_z u} \cdot \exp \left[-1/2 \left[\frac{ESH}{\sigma_z} \right]^2 \right]$$

For the recent assessment applications, assuming that Q is constant or near constant at any given time, the relationship between an observed (monitored) concentration (at a specific grid location where X/Q was calculated) and at another unmonitored location (where X/Q was also determined) was the ratio between the two values. This was a very simple approach, but was highly effective in determining the spread of a dispersion plume beyond and between the monitoring networks. This technique will allow for the prediction of future concentrations during ongoing operations, assuming that emissions do not significantly change during a particular operation. As remedial progress continues and emissions are eventually reduced, observed concentrations for a specified X/Q should also be reduced. Thus, the model has the potential to evaluate remedial progress under comparable meteorological and air dispersion conditions.

The PUFF Advection Model was used primarily for short-term predictions at Basin F in conjunction with real-time operations. Air monitoring was conducted using an OVM, an OVA, an ammonia monitor, and dust monitors at various remedial activity areas and along the Exclusion Zone perimeter. Perimeter readings were taken at four fixed locations and also downstream from Basin F, where maximum concentrations were anticipated (based on the prevailing wind). Concurrent with each monitoring reading, the PUFF Advection Model was run to determine the trajectory of the contamination plume. The X/Q values of the model were then immediately matched with actual monitored data to determine the potential distribution of the plume trajectory and the likelihood of Level B exceedances outside the protected areas. A maximum limit of 1 ppm was established for expansion of the Exclusion Zone and/or evacuation of personnel without protective clothing and equipment. The PUFF Advection Model was not employed during FY93.

Both the PUFF Advection Model and the ISC Model have the capability to provide longer-term X/Q dispersion assessments. The PUFF Advection Model operates on 15-minute trajectories. Although it takes much longer to run, the PUFF Advection Model can be used to assess longer-term episodes if more refined or sensitive analyses are required. The ISC Model was primarily employed for assessments at least 24-hours in length.

The results of the CMP, CAQMMP, and Basin F monitoring programs, discussed in Section 4, were compared with dispersion analyses obtained from the ISC Model. As noted, their evaluations provided highly useful information for associating dispersion patterns and meteorological conditions with observed and potential air contamination levels. When a distinct source was evident, such as Basin F, Basin A, or the South Plants, the model identified the spread of the dispersion plume downstream from the source. The concurrent monitored data then provided a mechanism for calibrating and projecting ambient concentrations throughout RMA and beyond its boundaries. The model identified worst-case meteorological conditions and seasonal and diurnal effects, thus providing a basis for mitigation actions if appropriate. For future assessments, it provides a basis for evaluating remedial progress by ensuring that monitoring results are compared to similar worst-case meteorological and seasonal conditions.

The strengths and weaknesses of model applications is evident from the data and comparisons given in previous RMA assessments. The impacts from known RMA sources or remedial activities have been identified. The relative strength of these impacts, taking into account distance from the source, have also been quantitatively determined. There is a considerable scattering of actual concentration values at specific site locations, which can be attributed to many factors beyond the simplified model's present capabilities. These include variations in source emissions and remedial production activity, both of which obviously result in variations in source concentration levels downwind from the source; the existence of other potential area and local sources, which present a noise factor in the database; short-term meteorological influences that are not integrated into the hourly ISC database (this suggests the special advantage

in using the short-term PUFF trajectories); and the general complexity of the dispersion process that is not entirely identified in the model.

In summary, the PUFF Advection and ISC Models, using the X/Q calibration approach, were and will continue to be workable vehicles for projecting contamination levels during remedial activities and for assessing remedial progress. It is not the purpose of the CAQMMP to perform extensive research investigations to improve model validation and applications. Nevertheless, several practical follow-up approaches may be applied in future CAQMMP evaluations that would not only improve the models but would also provide considerable insight into the general assessment of RMA contamination levels during subsequent remedial activities. These additional approaches are discussed below.

6.5.2 Additional Model Approaches and Analyses

6.5.2.1 Source Emissions Characterization

The standard dispersion model predicts ambient concentrations based upon meteorological factors and a known, or estimated, emission source. The Basin F real-time application, and the CMP/CAQMMP have, to a certain extent, effectively bypassed the emissions term and substituted a mechanism (X/Q values) for providing contamination levels based upon concurrent monitoring data. More precise data on emission releases at potential RMA and off-site contamination sources will add an additional dimension to the modeling and prediction capabilities. Data indicating the combination of emission levels, meteorological factors, and resultant ambient concentrations will also allow for further refinements of the model. For subsequent cleanup activities at RMA requiring air quality impact assessments, precise emissions data and characterizations, to the extent possible, will enhance prediction capabilities.

6.5.2.2 Remedial Activity Production Data

It is evident from the Basin F monitoring results during remedial operations, and from the Remedial Investigation Program results obtained prior to remedial operations, that the cleanup activities resulted in increases in some of the monitored pollutants, in particular: TSP, pesticides,

organic compounds, and some metals. It may be assumed that these increases were a direct consequence of and proportional to the extent of remedial efforts on a day-to-day basis. Various information pertaining to the status and intensity of cleanup operations (including the tons of soil and sludge removed and hauled each day, the type and number of vehicles and equipment employed, the type and number of storage areas and uncovered waste pile areas, the number of gallons of liquid transported, and any other emission-producing factors related to operations) would be especially useful in assessing relative emissions potential and in refining model prediction capabilities. These data were available, to a limited extent, during the Basin F operations, but not necessarily on a real-time basis; this made it difficult to interpret the cause of higher contamination events and the extent of mitigation efforts that were needed. It should be noted that many industrial real-time air quality prediction systems are directly related to production factors, and this approach would have equally valid application for waste cleanup operations.

6.5.3 Local and Regional Emissions Inventory

One of the difficulties in analyzing the CAQMMP and Basin F data was that there were obviously other off-site, local, and regional air emissions sources in the RMA vicinity that contributed to air pollution levels measured at RMA monitoring sites. As noted in the data evaluations, metropolitan Denver was a source of TSP, PM-10, some metals, gaseous criteria pollutants, and undoubtedly, certain organics. Also, various industrial sources and farming areas in north Denver, Commerce City, and Adams County may have contributed to organics and other potential contaminants. Unless these sources are properly identified, they can present an unknown factor in the dispersion model prediction evaluations of potential RMA contamination and remedial impacts. Therefore, a complete inventory of local off-site sources is needed for the interpretation of baseline data and remedial progress at RMA. Work has begun on this effort, and an inventory of known sources is included in various sections of this report. These data contribute to the overall analyses. Additional work is also required in identifying the contributions and impacts of metropolitan mobile sources on RMA monitoring results. Again, quantification of such influences would enhance model evaluations and prediction capabilities.

Much of this information is available in the CDH files, the EPA's Air Toxics Study for Denver, and other reports, including the Citizens Report on Toxic Pollution in Colorado (see Section 4.0).

6.5.4 Empirical/Statistical Adjustments

A final approach for model improvement is to apply empirical techniques and pertinent, or unique, physical relationships to improve model capabilities. For example, it is generally agreed that organics are released into the atmosphere at warmer temperatures and also under certain favorable air pressure conditions; and TSP, metals, and PM-10 are generally higher after certain threshold wind gust levels are achieved. Also, very localized, topographic, site-specific characteristics, which are not considered in the dispersion model, may be evident in the spread of potential air contaminants over the area. The database now being collected provides a mechanism for establishing some of these influences; some have already been identified in the present and previous reports. It is anticipated that this effort will continue and, if possible, validated relationships will be incorporated into model applications.

7.0 QUALITY ASSURANCE PROGRAM

7.1 OVERVIEW

The CAQMMP Quality Assurance Program (QAP) for Air Monitoring was designed to ensure that the data generated met the requirements of the project and needs of the data user. The QAP also ensures that the accuracy and precision of collected data are measurable and acceptable. The majority of monitoring and analytical techniques used were certified by PMRMA; non-certified methods had EPA approval. Each method has its own prescribed quality control and quality assurance procedures which are in accordance with the Chemical Quality Assurance Plan (PMRMA 1989). The guidelines for developing monitoring methods and procedures are described in the following documents:

- "Draft CAQMMP Standard Operating Procedures" (EBASCO 1992b)
- "Draft Quality Assurance Project Plan" (EBASCO 1992c)
- "PMRMA Certified Analytical Methods" (PMRMA 1990)
- "Ambient Monitoring Guidelines for Prevention of Significant Deterioration," EPA-450/4-87-007 (EPA 1987)
- "Ambient Air Quality Monitoring, Data Reporting, and Surveillance Provisions," 40 Code of Federal Regulations, Parts 51, 52, 53, and 58 (CFR 1987a)
- "Quality Assurance Handbook for Air Pollution Measurement Systems," Volumes I, II, and IV, EPA-600/9-76-005, EPA-600/4-77-027a, and EPA-600/4-82-060 (EPA 1984b, 1985, 1983)
- "Compendium of Methods for the Determination of Toxic Organic Compounds in Ambient Air," EPA-600/4-84-041 and updates (EPA 1984a)
- "Chemical Quality Assurance Plan," Version 1.0, July, 1989 (PMRMA 1989)

The following discussion on the definition of quality assurance is taken from the Quality Assurance Handbook for Air Pollution Measurement Systems, Volume I.

Quality assurance and quality control have been defined and interpreted in many ways. The more authoritative usages differentiate between the two terms by stating that quality control is the "system of activities to provide a quality product," while quality assurance or quality assessment is the "system of activities to provide assurance that the quality control system is adequate and effective." In other words, quality assurance is the verification of quality control.

Under PMRMA definitions, quality assurance is the total integrated program for assuring and documenting the reliability of monitoring and measurement data and for integrating quality planning, quality assessment, and quality improvement efforts to meet user requirements. Integrated into the QAP is quality control (QC), which is the routine application of procedures for obtaining prescribed performance standards in the monitoring and measurement process. Examples of QC activities are field and trip blanks, laboratory spikes, and duplicates.

For the purpose of the CAQMMP, quality control using project guidelines were those procedures that were routinely followed during the normal operation of the monitoring system. These included periodic field "spikes" and field blank sample analysis, collocated sample analysis, calibration of field and laboratory equipment, preventive maintenance, site inspection, and routine data screening and validation checks. Quality assurance (or quality assessment) were those procedures performed on a routine but less frequent basis to validate the data generation process. These assurance procedures were performed by a person not involved with the routine project activities. Quality assurance procedures included system and performance audits, standard intercomparisons, cross-checking of reported data values against original raw data records and data from other similar locations, and periodic evaluation of internal quality control data. The objective of the quality control and quality assurance procedures was to produce data that met RMA requirements measured in terms of precision, accuracy, representativeness, comparability, and completeness.

7.2 LABORATORY QUALITY ASSURANCE/QUALITY CONTROL PROGRAM

The laboratory quality assurance and quality control program was extremely rigorous and was based on PMRMA certification procedures. These procedures ensured the accuracy and integrity of the collected database through the analysis of two spiking levels and a control chart process of the resulting recoveries. The certification program consisted of the satisfactory completion of other precertification and certification procedures of the analytical methodologies, prior to sample analysis. In addition, prescribed routine quality control and quality assurance procedures were followed during the sampling program to ensure that each of the sampling methods maintained performance comparable to the level demonstrated during certification. The laboratory certification procedures consisted of determining several key parameters. Extraction or desorption efficiencies were determined by spiking the sample media with known amounts of the target analytes and performing the analyses along the analytical guidelines. A percent recovery for each of the compounds was determined from multiple spiking tests. Accuracy adjustments were determined from the mean of the percent recoveries at certification. One and two standard deviations from the mean of the recovery data were used to validate the subsequent data giving a 95 percent confidence level to the data. A range of concentrations was used to test the linearity of the laboratory instrument response and to certify an analytical range for each target analyte, with values denoted as the upper and lower certified reporting limits.

Reporting ranges were determined at the time of certification. Values found below this range were reported as "less than" the lower reporting limit. Values found above this range were reported as "greater-than" (GT), and were then estimated. These values are considered qualitative, and are used as a "best guess" of concentration ranges. Upper limits did not apply to TSP and PM-10 measurements. For SVOC, OCP and metals analysis, extract solutions were diluted (if possible) for reanalysis if the initial readings were above the certified range and holding times for the extracts had not yet expired.

The VOC laboratory analysis technique differed significantly from other techniques because there was no possibility of diluting or reanalyzing samples. The VOC Tenax and Tenax/charcoal

adsorption tubes were placed individually into a heating block connected to the inlet of the analytical instrument. This block, or thermal desorber, was quickly heated to desorb all the volatile compounds from the sorbent material. A small gas flow was passed through the tube simultaneously, which allowed these compounds to be quickly injected into the analytical instrument through a heated stainless-steel interface. Therefore, this method was a "one shot" technique, and if the sample concentration was outside certified limits, there was no way to dilute and reanalyze the sample. In the FY93 program, there were instances where a sample contained an amount above the certified range for some compounds. An assessment of the concentration was made by interpolating the concentration curve above the certified range. The results in these cases produced the best estimate of the observed concentrations. Although such GT values are not certified values, they provide essential information on the magnitude of the detected compound. The GT value supplied by the laboratory produces a partial representation of what was collected on the sample tube. Since the volatile method allows only one analysis per tube, the estimated greater than values are utilized. The estimated GT values began to appear in the Installation Restoration Data Management Information System (IRDMIS) database in August FY93. All GT values for FY93 are available in the laboratory packages and EBASCO database. For the FY93 program, additional GT results from ESE/Denver laboratory were obtained from microfilmed data packets maintained by RMA. The software employed by ESE/Denver to produce the estimated value generated the concentrations automatically. No manual integration or calculation was used.

In addition to the analyte measurements, laboratory accuracies, certified ranges, and extraction efficiencies were directly incorporated into the database processing system that was used to report air quality data. This system is the IRDMIS. In IRDMIS, the raw laboratory results are sent to EBASCO, the data are checked for errors, adjusted for sample volumes, and sent to DP Associates and held in the QC holding database. Once the Army has accepted data as having met the QA criteria, data are located into the official database. Data outside of certified ranges and rejected under the QA criteria are placed in a rejected data file for informational use only.

The resulting product is a high-quality analytical database that is available for assessment of RMA air quality.

7.3 FIELD QUALITY ASSURANCE/QUALITY CONTROL PROGRAM

7.3.1 Overview

During FY93 a comprehensive field QA/QC program was employed to ensure the collection of valid air quality and meteorological data. A Quality Assurance Project Plan was developed and implemented by an independent QA/QC group for the CAQMMP. All field staff received training in these as well as the program's SOPs. These SOPs clearly document the steps necessary to perform all field activities and serve as a resource for standardized data and calibration forms. Copies of all completed data forms were maintained in the field site files while the originals were filed by the project data management coordinator in the EBASCO Lakewood, Colorado office. All data forms were routinely reviewed and updated by the field supervisor under the guidance of the Field Technical Supervisor.

7.3.2 Ambient Air Sampling

The CAQMMP Technical plan and SOPs specify target flow rates and sample durations for each of the ambient air techniques as well air stripper and vent sampling techniques. These target values were used during the field sampling activities and were documented by periodic equipment calibrations as well as checks during each sampling event. Quality assurance limits were established using guidelines obtained from EPA QA handbooks for acceptable instrument performance. Corrective maintenance was performed whenever a flow rate or elapsed timer was outside of acceptable limits. All calibration and maintenance activities were documented in on-site logs by the field staff.

The high-volume samplers used to collect TSP, PM-10, and metals were electronically flow controlled to maintain the target flow rate. These samplers were calibrated quarterly or whenever additional maintenance was performed by using a certified orifice and following procedures prescribed in the appropriate reference method. These samplers were also equipped with Dickson

recorders that provided a 24-hour record of sample flow rate. The high volume samplers used for OCP/SVOC sampling were equipped with Magnehelic flow rate gauges. During quarterly calibrations, a linear relationship was established between the Magnehelic reading and the calibration orifice indicated flow rate.

Mercury, VOCs, and asbestos samples were collected using Volatile Organic Toxics Analyzer (VOTA) samplers or constant flow sampling pumps. These devices were adjusted to the target flow rate at sample installation by using a certified mass flow meter. Sampling equipment was also recalibrated whenever indicated flow rates exceeded target flow rates by greater than 7 percent.

7.3.2.1 VOC Quality Control Results

Collocated aerated samples and field and trip blank samples were submitted with each sample set to assess the method's precision and to identify potential sources of media contamination. Table 7.3-1 presents the FY93 VOC precision results. Average absolute percent difference for the collocated pairs range from 11.0 percent to 47.6 percent for the various target analytes. A variety of factors may account for this variability. These are discussed fully in Section 4.6.8. Table 7.3-2 summarizes the laboratory results for field and trip blanks samples. For 15 out of the 20 target analytes, all field and trip blank levels were less than the lower CRL. Trans-1,2-dichloroethene in a trip blank and benzene in a field blank registered 1 detection each at a level just above the lower CRL. 1,1,1-trichloroethene was detected on the field blank in 5 instances, most notably on 7/21/93 when 0.640 microgram was reported on the field blank. This amount is comparable to that measured on the field, aerated samples. Thus, 1,1,1-Trichloroethene results for this day should be viewed as suspect. Methylene chloride and toluene appear on both field and trip blank samples with some frequency at low to moderate levels. In most cases, the amount detected on the field and/or trip blank sample is small (i.e., < 10 percent) relative to the mass collected on the aerated samples. Nevertheless, data from the affected sample sets should be used with some caution. A complete listing of all field and trip blank results is provided in Appendix K.

Table 7.3-1 Summary of FY93 Volatile Organic Compound Precision Results

| Analyte | Total # Samples | # Samples > Lower CRL | # Pairs > 30% Diff. | Avg. Abs. % Diff. | Avg. % Diff. | Std. Dev. | Min. Abs. % Diff. | Max. Abs. % Diff. | Lower 95% Conf. Limit | Upper 95% Conf. Limit |
|---------|--------------------|--------------------------|------------------------|----------------------|-----------------|--------------|----------------------|----------------------|--------------------------|--------------------------|
| 111TCE | 56 | 54 | 13 | 22.3 | 4.9 | 32.1 | 0.3 | 117.1 | -39.5 | +49.4 |
| 112TCE | 59 | 0 | - | - | - | - | - | - | - | - |
| 111DCE | 59 | 0 | - | - | - | - | - | - | - | - |
| 12DCLE | 59 | 2 | 0 | 11.0 | -1.9 | 15.6 | 9.1 | 13.0 | -23.6 | +19.7 |
| 12DCE | 59 | 0 | - | - | - | - | - | - | - | - |
| BCHPD | 59 | 6 | 2 | 26.6 | -23.7 | 31.1 | 5.0 | 78.8 | -66.8 | 19.4 |
| C6H6 | 56 | 54 | 15 | 27.1 | 2.6 | 44.2 | 0.3 | 192.1 | -58.6 | 63.9 |
| CCL4 | 56 | 53 | 16 | 23.9 | 6.2 | 33.7 | 0 | 116.9 | -40.5 | 52.9 |
| CH2CL2 | 56 | 54 | 26 | 42.8 | 3.5 | 61.4 | 0 | 168.7 | -81.6 | 88.6 |
| CHCL3 | 57 | 53 | 16 | 33.6 | -2.9 | 54.5 | 0 | 165.8 | -78.4 | 72.7 |
| CLC6H5 | 59 | 0 | - | - | - | - | - | - | - | - |
| DBCP | 59 | 1 | - | - | - | - | - | - | - | - |
| DCPD | 59 | 0 | - | - | - | - | - | - | - | - |
| DMDS | 59 | 0 | - | - | - | - | - | - | - | - |
| ETC6H5 | 57 | 35 | 23 | 41.7 | -0.7 | 50.2 | 0 | 109.4 | -70.4 | 68.9 |
| MEC6H5 | 56 | 54 | 30 | 47.4 | -1.0 | 62.5 | 1.4 | 174.7 | -87.6 | 85.6 |
| MIBK | 59 | 0 | - | - | - | - | - | - | - | - |
| TCLEE | 56 | 44 | 20 | 34.2 | -4.0 | 44.5 | 0 | 112.0 | -65.6 | 57.7 |
| TRCLE | 59 | 5 | 0 | 14.4 | -5.5 | 17.6 | 3.5 | 22.9 | -29.9 | 18.8 |
| XYLENES | 57 | 36 | 24 | 47.6 | 0.7 | 56.7 | 3.4 | 120.0 | -77.9 | 79.3 |

CRL - Certified Reporting Limit
 ABS. - Absolute
 AVG. - Average
 Dev. - Deviation
 Diff. - Difference
 Conf. - Confidence
 Std. - Standard

CLC6H5 - Chlorobenzene
 DBCP - Dibromochloropropane
 DCPD - Dicyclopentadiene
 DMDS - Dimethylsulfide
 ETC6H5 - Ethylbenzene
 MEC6H5 - Toluene
 MIBK - Methyl Isobutyl Ketone
 TCLEE - Tetrachloroethene
 TRCLE - Trichloroethene
 XYLENES - Total Xylenes

Table 7.3-2 Summary of FY93 Volatile Organic Compound Field
and Trip Blank Results

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| Analyte | Sample Type | Total # Samples | Total # Detections | Average Weight (µg) | Maximum Weight (µg) |
|--------------------------|-------------|-----------------|--------------------|---------------------|---------------------|
| 1,1,1-Trichloroethane | FBLK | 73 | 5 | <0.023 | 0.640 |
| | TRIP | 75 | 0 | <0.023 | <0.023 |
| 1,1,2-Trichloroethane | FBLK | 73 | 0 | <0.037 | <0.037 |
| | TRIP | 75 | 0 | <0.037 | <0.037 |
| 1,1-Dichloroethane | FBLK | 73 | 0 | <0.012 | <0.012 |
| | TRIP | 75 | 0 | <0.012 | <0.012 |
| 1,2-Dichloroethane | FBLK | 73 | 0 | <0.012 | <0.012 |
| | TRIP | 75 | 0 | <0.012 | <0.012 |
| trans-1,2-Dichloroethene | FBLK | 73 | 0 | <0.012 | <0.012 |
| | TRIP | 75 | 1 | <0.012 | 0.017 |
| Bicycloheptadiene | FBLK | 73 | 0 | <0.012 | <0.012 |
| | TRIP | 75 | 0 | <0.012 | <0.012 |
| Benzene | FBLK | 73 | 1 | <0.016 | 0.021 |
| | TRIP | 75 | 0 | <0.016 | <0.016 |
| Carbon Tetrachloride | FBLK | 73 | 0 | <0.018 | <0.018 |
| | TRIP | 75 | 0 | <0.018 | <0.018 |
| Methylene Chloride | FBLK | 73 | 43 | 0.285 | 2.710 |
| | TRIP | 75 | 13 | <0.097 | 0.799 |
| Chloroform | FBLK | 73 | 0 | <0.012 | <0.012 |
| | TRIP | 75 | 0 | <0.012 | <0.012 |
| Chlorobenzene | FBLK | 73 | 0 | <0.013 | <0.013 |
| | TRIP | 75 | 0 | <0.013 | <0.013 |
| Dibromochloropropane | FBLK | 73 | 0 | <0.100 | <0.100 |
| | TRIP | 75 | 0 | <0.100 | <0.100 |
| Dicyclopentadiene | FBLK | 73 | 0 | <0.089 | <0.089 |
| | TRIP | 75 | 0 | <0.089 | <0.089 |
| Dimethyldisulfide | FBLK | 73 | 0 | <0.048 | <0.048 |
| | TRIP | 75 | 0 | <0.048 | <0.048 |
| Ethylbenzene | FBLK | 73 | 0 | <0.032 | <0.032 |
| | TRIP | 75 | 0 | <0.032 | <0.032 |
| Toluene | FBLK | 73 | 17 | 0.020 | 0.280 |
| | TRIP | 75 | 7 | <0.014 | 0.210 |
| Methyl Isobutyl Ketone | FBLK | 73 | 0 | <0.160 | <0.160 |
| | TRIP | 75 | 0 | <0.160 | <0.160 |
| Tetrachloroethene | FBLK | 73 | 0 | <0.013 | <0.013 |
| | TRIP | 75 | 0 | <0.013 | <0.013 |
| Trichloroethene | FBLK | 73 | 0 | <0.013 | <0.013 |
| | TRIP | 75 | 0 | <0.013 | <0.013 |
| Xylenes | FBLK | 73 | 0 | <0.149 | <0.149 |
| | TRIP | 75 | 0 | <0.149 | <0.149 |

< ### Indicates value less than lower certified reporting limit
µg Micrograms

7.3.2.2 OCP/SVOC Quality Control Results

Table 7.3-3 presents the FY93 OCP precision results. Average absolute percent differences for the collocated pairs range from 6.3 percent to 15.9 percent for the target analytes. These results indicate that the OCP results are generally reliable. A complete listing of all OCP and SVOC field blank samples is provided in Appendix K. There were no detections of the target OCP or SVOC compounds in any of the blank samples. This indicates that the OCP/SVOC contamination is not present at detectable levels in the media or gained through the shipping and handling process.

In an effort to quantify the recovery efficiency of the SVOC/OCP sampling/analytical methods, three field spiking events were performed during FY93. On 12/22/92 and 4/29/93, two spiked samples were taken concurrently with one control (unspiked) sample, and the third event took place on 8/3/93 and included two aerated spiked samples, one non-aerated sample, and one control sample. A solution spiked with all of the SVOC/OCP target compounds was prepared by the PMRMA laboratory and employed by the project field staff for the spiking events. Samples were spiked by applying the solution directly onto a felt filter at the beginning of the sample period. The felt filter was placed over the regular quartz filter in the sampling module. Twenty microliters of the spiking mixture was applied to each spiked sample taken during the three spiking events. The results of the three spiking events are listed in Table 7.3-4. The recoveries for the nonaerated spiked sample (event #3) exhibit acceptable concentrations for all SVOC/OCP spiking compounds. This indicates a loss of target compounds during the aeration process. Consequently, analytical results for aldrin, CPMS, DBCP, vapona, DIMP, dithiane, DMMP, isodrin, and oxathiane must be interpreted carefully in view of their low or zero recoveries. The recoveries for the other target analytes showed the methods are efficient at collecting and retaining these contaminants of concern.

Table 7.3-3 Summary of FY93 Organochlorine Pesticide Precision Results

| Analyte | Total # Samples | # Samples w/i CRL | # Pairs > 30% Diff. | Avg. Abs. % Diff. | Avg. % Diff. | Std. Dev. | Min. Abs. % Diff. | Max. Abs. % Diff. | Lower 95% Conf. Limit | Upper 95% Conf. Limit |
|-----------|--------------------|----------------------|------------------------|----------------------|-----------------|--------------|----------------------|----------------------|--------------------------|--------------------------|
| Aldrin | 60 | 6 | 1 | 8.8 | -8.8 | 17.8 | 0 | 44.4 | -33.4 | 15.8 |
| Chlordane | 60 | 10 | 0 | 6.3 | -6.3 | 6.1 | 0 | 15.4 | -14.8 | 2.2 |
| Dieldrin | 60 | 38 | 2 | 7.2 | -2.7 | 11.3 | 0 | 40.0 | -18.4 | 13.0 |
| Endrin | 60 | 6 | 1 | 15.9 | -3.1 | 22.3 | 0 | 40.0 | -34.0 | 27.8 |
| Isodrin | 60 | 0 | - | - | - | - | - | - | - | - |
| PPDDE | 60 | 0 | - | - | - | - | - | - | - | - |
| PPDDT | 60 | 1 | - | - | - | - | - | - | - | - |

| | | | | | | | |
|-------|---|------------|------|---|-------|---|------------|
| w/i | - | within | Avg. | - | Conf. | - | Confidence |
| Diff. | - | Difference | Abs. | - | Std. | - | Standard |
| | | | | - | Dev. | - | Deviation |

Table 7.3-4 Summary of Semivolatile Organic Compound/Organochlorine Pesticide Field Spike Percent Recoveries

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| Target Analyte | Event #1 12/22/92 | | Event #2 4/29/93 | | Event #3 8/3/93 | | |
|----------------|-------------------|-----------|------------------|-----------|-----------------|-----------|------------|
| | Sample #1 | Sample #2 | Sample #1 | Sample #2 | Sample #1 | Sample #2 | Sample #3* |
| Aldrin | 64.7 | 88 | 26 | 28 | 22 | 0 | 89 |
| Atrazine | 92.4 | 140 | 40 | 51 | 71 | 78 | 84 |
| Chlordane | GT | GT | 110 | 110 | 94 | 90 | 92 |
| CPMS | 0 | 0 | 0 | 0 | 0 | 0 | 115 |
| CPMSO | 107 | 154 | 106 | 116 | 49 | 50 | 104 |
| CPMSO2 | 102 | 148 | 100 | 110 | 88 | 89 | 99 |
| DBCP | 0 | 0 | 0 | 0 | 0 | 0 | 123 |
| Vapona | 0 | 22 | 0 | 0 | 0 | 0 | 55 |
| DIMP | 0 | 0 | 0 | 0 | 0 | 0 | 121 |
| Dithiane | 0 | 0 | 0 | 0 | 0 | 0 | 88 |
| Dieldrin | 72 | 118 | 87 | 95 | 81 | 80 | 82 |
| DMMP | 0 | 0 | 0 | 0 | 0 | 0 | 87 |
| Endrin | 96 | 152 | 115 | 121 | 86 | 87 | 87 |
| Isodrin | 56 | 78 | 26 | 33 | 24 | 21 | 88 |
| Malathion | 147 | 220 | 107 | 122 | 114 | 112 | 114 |
| Oxathiane | 0 | 0 | 0 | 0 | 0 | 0 | 51 |
| ppDDE | 60 | 89 | 91 | 97 | 86 | 84 | 90 |
| ppDDT | 81 | 129 | 93 | 95 | 87 | 88 | 82 |
| Parathion | 91 | 128 | 113 | 131 | 93 | 94 | 99 |
| Supona | 92 | 137 | 108 | 124 | 98 | 95 | 99 |

* Spike sample was not aerated

GT Above certified reporting limit (no percent recovery calculated)

7.3.2.3 Metals, Arsenic, and Mercury Quality Control Results

The results of collocated sampling for metals, arsenic, and mercury are presented in Table 7.3-5. These precision calculations demonstrate that the sampling method for these elements yields generally stable results. Copper proves to be a notable exception, with an average absolute percent difference of 43.6 percent. It is important to consider this variability when analyzing the data for spatial or temporal trends. As Table 7.3-6 shows, there was only one detection of a target metal analyte on a field or trip blank sample during FY93. Copper was measured on one field blank sample (out of a possible 81) at a level just above the lower certified reporting limit (12.8 micrograms). All arsenic field and trip blank samples were non-detects. Mercury was measured on 3 of the 84 field blank samples submitted at levels comparable to those typically measured on the aerated tubes. On two of those sample days, April 10, 1993 and July 9, 1993, no mercury was detected on any of the aerated samples. On June 10, 1993, mercury was detected at site AQ3. This value appears suspect as there was a greater mass of mercury measured on the field blank sample for this day. All field and trip blank data are presented in Appendix K.

7.3.2.4 TSP and PM-10 Quality Control Results

As discussed in Section 4.2, the laboratory experienced many problems in its gravimetric analysis of TSP and PM-10 filters during FY93. Quality control samples played a key role in identifying these problems and determining the effectiveness of corrective actions. The majority of these problems occurred during the period of October 1992 through January 1993. During the entire fiscal year sampling program, there were 13 field blank and 2 trip blank samples that showed appreciable weight gains or losses (> 5.0 milligrams net gain or loss). In addition, the laboratory reported negative net weights (i.e., < 0 milligrams) for 10 aerated TSP and 14 PM-10 samples. Furthermore, there were 69 pairs of samples at the same site where the reported PM-10 concentration was greater than the corresponding TSP value. All these data have been flagged in the data listings. Precision calculations were performed on the remaining valid collocated data sets and these results are documented in Table 7.3-7. All field and trip blank data are reported in Appendix K.

Table 7.3-5 Summary of Metals, Arsenic, and Mercury Precision Results

| Analytes | Total # of Samples | # Samples within CRL | # Pairs > 30% | Average | | Standard Deviation | Minimum | | Maximum Absolute % Difference | Lower 95% Confidence Limit | | Upper 95% Confidence Limit | |
|----------|-----------------------|-------------------------|------------------|--------------------------|-------------------------|-----------------------|--------------------------|-------------------------|-------------------------------------|----------------------------------|---------------------|----------------------------------|---------------------|
| | | | | Absolute % Difference | Average % Difference | | Absolute % Difference | Average % Difference | | Confidence Limit | Confidence Limit | Confidence Limit | Confidence Limit |
| Cadmium | 66 | 1 | 0 | 6.7 | 6.7 | -- | 6.7 | 6.7 | 6.7 | -- | -- | -- | -- |
| Chromium | 66 | 0 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| Copper | 66 | 66 | 38 | 43.6 | -12.7 | 51.5 | 0.3 | 5.5 | 114.3 | -84.1 | 58.7 | 58.7 | 58.7 |
| Lead | 66 | 3 | 0 | 9.5 | 9.5 | 5.2 | 5.5 | 9.5 | 15.3 | 2.3 | 16.6 | 16.6 | 16.6 |
| Zinc | 66 | 57 | 0 | 7.2 | 2.9 | 9.4 | 0.1 | 2.9 | 26.2 | -10.1 | 16.0 | 16.0 | 16.0 |
| Arsenic | 66 | 4 | 0 | 11.4 | 11.4 | 9.4 | 0 | 11.4 | 20 | -1.6 | 24.4 | 24.4 | 24.4 |
| Mercury | 67 | 0 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |

CRL Certified Reporting Limits

%

>

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Table 7.3-6 Summary of FY93 Metals, Arsenic, and Mercury Field and Trip
Blank Results

Page 1 of 1

| Analyte | Sample Type | Total # Samples | Total # Detections | Average Weight (µg) | Maximum Weight (µg) |
|----------|-------------|-----------------|--------------------|---------------------|---------------------|
| Cadmium | FBLK | 81 | 0 | <4.00000 | <4.00000 |
| | TRIP | 18 | 0 | <4.00000 | <4.00000 |
| Chromium | FBLK | 81 | 0 | <20.0000 | <20.0000 |
| | TRIP | 18 | 0 | <20.0000 | <20.0000 |
| Copper | FBLK | 81 | 1 | <10.0000 | 12.80000 |
| | TRIP | 18 | 0 | <10.0000 | <10.0000 |
| Lead | FBLK | 81 | 0 | <40.0000 | <40.0000 |
| | TRIP | 18 | 0 | <40.0000 | <40.0000 |
| Zinc | FBLK | 81 | 0 | <20.0000 | <20.0000 |
| | TRIP | 18 | 0 | <20.0000 | <20.0000 |
| Arsenic | FBLK | 81 | 0 | <1.4100 | <1.4100 |
| | TRIP | 18 | 0 | <1.4100 | <1.4100 |
| Mercury | FBLK | 84 | 3 | < 0.1000 | 0.2180 |
| | TRIP | 0 | 0 | NA | NA |

< ##### Indicates result less than the lower certified reporting limit

µg micrograms

NA Not applicable

Table 7.3-7 Summary of TSP and PM-10 Precision Results

Page 1 of 1

| Analyte | Total # Samples | # Samples Valid | # Pairs > 15% Diff. | Avg. Abs. % Diff. | Avg. % Diff. | Std. Dev. | Min. Abs. % Diff. | Max. Abs. % Diff. | Lower 95% Conf. Limit | Upper 95% Conf. Limit |
|---------|--------------------|--------------------|------------------------|----------------------|-----------------|--------------|----------------------|----------------------|--------------------------|--------------------------|
| | | | | | | | | | | |
| TSP | 69 | 42 | 2 | 6 | 1 | 9 | 0 | 33 | -12 | +13 |
| PM-10 | 60 | 35 | 9 | 11 | 0 | 21 | 0 | 86 | -29 | +28 |

| | | | | | |
|-------|---|------------|-----------|---|--------------------|
| Diff. | - | Difference | Std. Dev. | - | Standard Deviation |
| Avg. | - | Average | Conf. | - | Confidence |
| Abs. | - | Absolute | > | - | Greater than |
| % | - | Percent | | | |

7.3.3 Continuous Gaseous Monitoring Quality Control Results

Manual zero, span, and precision checks were performed biweekly at the continuous air quality monitoring site on all parameters. Tables 7.3-8 to 7.3-11 and Figures 7.3-1 to 7.3-4 present the results of these checks. The carbon monoxide, sulfur dioxide, ozone, and oxides of nitrogen results were generally stable with average absolute percent difference of 5.0 percent, 3.7 percent, 1.7 percent, and 5.8 percent, respectively. The NO_x analyzer underwent corrective maintenance in several instances during FY93 when daily and manual zero, span, or precision checks indicated out of tolerance conditions.

7.3.4 Audit Results

The CAQMMP quality assurance procedures included system audits, performance audits of sampling equipment, and participation in the EPA administered National Performance Audit Program (NPAP).

The project quality assurance manager conducted periodic inspections and reviews of the monitoring operation and documentation procedures. Audited operations included sample preparation, installation, removal, reclamation, shipping, and documentation. Observed procedures were compared to those detailed in the project SOPs. Results were reported to project management with recommendations of corrective actions when appropriate. These reports are contained in project files. Also, an independent audit of field and laboratory procedures was performed by Martin Marietta Energy Systems during FY93. Their findings were reported to PMRMA.

Performance audits were performed quarterly on all field samples, including the continuous gaseous analyzer; and semi-annually on the 4 meteorological towers by an independent auditor. All sampler, analyzer, and tower calibrations were tested against standards other than those used to assess operations on a routine basis. For each sampler, the auditor-determined flow rate was compared to the operator-determined flow rate. During FY93, 10 out of 214 samplers audited

Table 7.3-8 Carbon Monoxide Precision Calculations CAQMMP FY1993

Page 1 of 1

| Date | Analyzer Response (ppm) | Calibrator Output (ppm) | % Difference | Absolute % Difference |
|-----------------------------|-------------------------------|-------------------------------|-----------------|--------------------------|
| 10-14-92 | 9.0 | 9.1 | -1.1 | 1.1 |
| 10-28-92 | 9.01 | 9.68 | -6.9 | 6.9 |
| 11-11-92 | 9.3 | 9.7 | -4.1 | 4.1 |
| 11-11-92 | 9.04 | 9.71 | -6.9 | 6.9 |
| 11-25-92 | 8.7 | 9.7 | -10.3 | 10.3 |
| 12-09-92 | 8.6 | 9.7 | -11.3 | 11.3 |
| 12-23-92 | 9.2 | 9.7 | -5.2 | 5.2 |
| 01-06-93 | 9.13 | 9.7 | -5.9 | 5.9 |
| 01-20-93 | 9.4 | 9.7 | -3.1 | 3.1 |
| 02-03-93 | 9.7 | 9.7 | 0.0 | 0.0 |
| 02-27-93 | 9.2 | 9.7 | -5.2 | 5.2 |
| 03-10-93 | 9.4 | 9.7 | -3.1 | 3.1 |
| 03-24-93 | 9.5 | 9.7 | -2.1 | 2.1 |
| 04-07-93 | 9.5 | 9.7 | -2.1 | 2.1 |
| 04-21-93 | 10.5 | 10.1 | 4.0 | 4.0 |
| 05-12-93 | 9.2 | 10.1 | -8.9 | 8.9 |
| 05-12-93 | 9.5 | 10.1 | -5.9 | 5.9 |
| 05-26-93 | 9.2 | 10.1 | -8.9 | 8.9 |
| 06-07-93 | 9.5 | 10.1 | -5.9 | 5.9 |
| 06-07-93 | 9.6 | 10.1 | -5.0 | 5.0 |
| 06-23-93 | 9.5 | 10.1 | -5.9 | 5.9 |
| 07-07-93 | 9.3 | 10.1 | -7.9 | 7.9 |
| 07-21-93 | 9.9 | 10.1 | -2.0 | 2.0 |
| 08-04-93 | 9.6 | 10.1 | -5.0 | 5.0 |
| 08-18-93 | 9.7 | 10.1 | -4.0 | 4.0 |
| 09-02-93 | 9.7 | 10.1 | -4.0 | 4.0 |
| 09-23-93 | 10.2 | 10.1 | 1.0 | 1.0 |
| Average % Difference | | | -4.7 | 5.0 |
| Standard Deviation | | | 3.4 | 3.4 |
| Upper 95% Probability Limit | | | 2.1 | |
| Lower 95% Probability Limit | | | -11.4 | |

ppm - parts per million
 % - Percent

Table 7.3-9 Sulfur Dioxide Precision Calculations CAQMMP FY93

Page 1 of 1

| Date | Analyzer Response (ppb) | Calibrator Output (ppb) | % Difference | Absolute % Difference |
|-----------------------------|-------------------------------|-------------------------------|-----------------|--------------------------|
| 10-14-92 | 98.1 | 93.0 | 5.5 | 5.5 |
| 10-28-92 | 92.0 | 99.4 | -7.4 | 7.4 |
| 11-11-92 | 94.4 | 99.4 | -5.0 | 5.0 |
| 11-16-92 | 103.5 | 103.8 | -0.3 | 0.3 |
| 11-16-92 | 92.7 | 99.4 | -6.7 | 6.7 |
| 11-25-92 | 102.8 | 103.8 | -1.0 | 1.0 |
| 12-09-92 | 97.04 | 103.8 | -6.5 | 6.5 |
| 12-23-92 | 100.4 | 103.8 | -3.3 | 3.3 |
| 01-06-93 | 97.3 | 103.8 | -6.3 | 6.3 |
| 01-19-93 | 98.1 | 103.8 | -5.5 | 5.5 |
| 01-19-93 | 99.74 | 105.7 | -5.6 | 5.6 |
| 01-20-93 | 102.0 | 105.7 | -3.5 | 3.5 |
| 02-04-93 | 104.7 | 105.7 | -0.9 | 0.9 |
| 02-27-93 | 99.4 | 105.7 | -6.0 | 6.0 |
| 03-10-93 | 99.4 | 105.7 | -6.0 | 6.0 |
| 03-24-93 | 103.7 | 105.7 | -1.9 | 1.9 |
| 04-07-93 | 100.4 | 105.7 | -5.0 | 5.0 |
| 04-21-93 | 114.1 | 115.3 | -1.0 | 1.0 |
| 05-12-93 | 77.5 | 78.7 | -1.5 | 1.5 |
| 05-12-93 | 75.2 | 78.7 | -4.4 | 4.4 |
| 05-26-93 | 104.7 | 101.5 | 3.2 | 3.2 |
| 06-07-93 | 100.0 | 101.5 | -1.5 | 1.5 |
| 06-23-93 | 108.8 | 108.8 | 0.0 | 0.0 |
| 07-07-93 | 112.1 | 108.8 | 3.0 | 3.0 |
| 07-21-93 | 115.1 | 108.8 | 5.8 | 5.8 |
| 08-04-93 | 110.4 | 108.8 | 1.5 | 1.5 |
| 08-18-93 | 112.4 | 108.8 | 3.3 | 3.3 |
| 09-02-93 | 114.2 | 108.8 | 5.0 | 5.0 |
| 09-23-93 | 111.0 | 108.8 | 2.0 | 2.0 |
| Average % difference | | | -1.7 | 3.7 |
| Standard deviation | | | 4.0 | 4.0 |
| Upper 95% probability limit | | | 6.2 | |
| Lower 95% probability limit | | | -9.7 | |

ppb - parts per billion
 % - Percent

Table 7.3-10 Ozone Precision Calculations CAQMMP FY93

Page 1 of 1

| Date | Analyzer Response (ppb) | Calibrator Output (ppb) | % Difference | Absolute % Difference |
|-----------------------------|-------------------------------|-------------------------------|-----------------|--------------------------|
| 10-14-92 | 88.7 | 89.8 | -1.2 | 1.2 |
| 10-28-92 | 88.5 | 89.0 | -0.6 | 0.6 |
| 11-11-92 | 89.7 | 88.8 | 1.0 | 1.0 |
| 11-25-92 | 89.3 | 88.7 | 0.7 | 0.7 |
| 12-09-92 | 88.0 | 88.3 | -0.3 | 0.3 |
| 12-23-92 | 88.5 | 89.5 | -1.1 | 1.1 |
| 01-06-93 | 89.0 | 89.0 | 0.0 | 0.0 |
| 01-20-93 | 88.8 | 89.7 | -1.0 | 1.0 |
| 02-03-93 | 86.6 | 89.0 | -2.7 | 2.7 |
| 02-27-93 | 92.0 | 89.0 | 3.4 | 3.4 |
| 03-01-93 | 75.8 | 73.7 | 2.8 | 2.8 |
| 03-10-93 | 91.9 | 89.2 | 3.0 | 3.0 |
| 03-24-93 | 91.0 | 89.0 | 2.2 | 2.2 |
| 04-07-93 | 89.9 | 88.7 | 1.4 | 1.4 |
| 04-21-93 | 90.3 | 89.2 | 1.2 | 1.2 |
| 05-12-93 | 91.2 | 89.0 | 2.5 | 2.5 |
| 05-26-93 | 90.3 | 89.5 | 0.9 | 0.9 |
| 06-07-93 | 91.3 | 89.7 | 1.8 | 1.8 |
| 06-23-93 | 90.8 | 89.7 | 1.2 | 1.2 |
| 07-07-93 | 90.2 | 89.0 | 1.3 | 1.3 |
| 07-21-93 | 90.1 | 89.5 | 0.7 | 0.7 |
| 08-04-93 | 89.6 | 89.2 | 0.4 | 0.4 |
| 08-18-93 | 94.2 | 89.0 | 5.8 | 5.8 |
| 09-02-93 | 92.5 | 89.3 | 3.6 | 3.6 |
| 09-02-93 | 91.2 | 90.2 | 1.1 | 1.1 |
| 09-23-93 | 91.3 | 89.5 | 2.0 | 2.0 |
| Average % Difference | | | 1.2 | 1.7 |
| Standard Deviation | | | 1.8 | 1.8 |
| Upper 95% Probability Limit | | | 4.7 | |
| Lower 95% Probability Limit | | | -2.4 | |

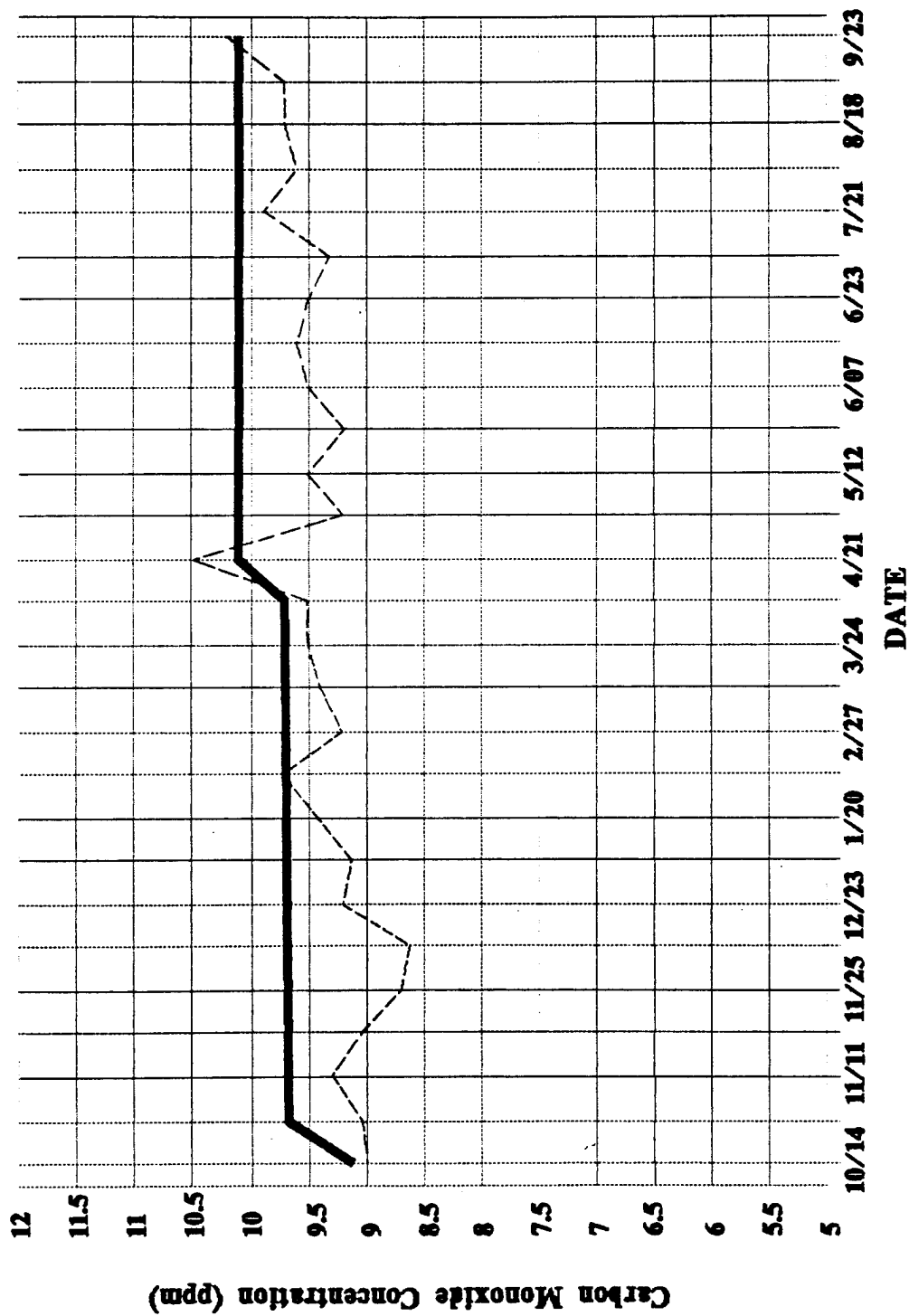
ppb - parts per billion
 % - Percent

Table 7.3-11 Oxides of Nitrogen Precision Calculations CAQMMP FY1993

Page 1 of 1

| Date | Analyzer Response (ppb) | Calibrator Output (ppb) | % Difference | Absolute % Difference |
|-----------------------------|-------------------------------|-------------------------------|-----------------|--------------------------|
| 10-14-92 | 102.5 | 104 | -1.4 | 1.4 |
| 10-28-92 | 94.7 | 109.3 | -13.4 | 13.4 |
| 11-11-92 | 102.1 | 109.3 | -6.6 | 6.6 |
| 11-16-92 | 106.2 | 102.8 | 3.3 | 3.3 |
| 11-16-92 | 97.4 | 109.3 | -10.9 | 10.9 |
| 11-25-92 | 101.8 | 102.8 | -1.0 | 1.0 |
| 12-09-92 | 104.8 | 102.8 | 1.9 | 1.9 |
| 12-23-92 | 92.7 | 102.8 | -9.8 | 9.8 |
| 01-06-93 | 88.3 | 102.8 | -14.1 | 14.1 |
| 01-19-93 | 101.0 | 102.8 | -1.8 | 1.8 |
| 01-19-93 | 95.4 | 105.0 | -9.2 | 9.2 |
| 01-20-93 | 102.4 | 105.0 | -2.5 | 2.5 |
| 02-03-93 | 107.4 | 105.0 | 2.3 | 2.3 |
| 02-04-93 | 116.2 | 105.0 | 10.7 | 10.7 |
| 02-27-93 | 100.1 | 105.0 | -4.7 | 4.7 |
| 02-27-93 | 101.1 | 105.0 | -3.7 | 3.7 |
| 03-10-93 | 99.7 | 105.0 | -5.0 | 5.0 |
| 03-24-93 | 94.0 | 105.0 | -10.5 | 10.5 |
| 03-31-93 | 74.9 | 105.0 | -28.7 | 28.7 |
| 03-31-93 | 93.4 | 105.0 | -11.0 | 11.0 |
| 04-07-93 | 106.4 | 105.0 | 1.3 | 1.3 |
| 04-21-93 | 118.8 | 114.5 | 3.8 | 3.8 |
| 05-12-93 | 76.9 | 78.2 | -1.7 | 1.7 |
| 05-26-93 | 101.0 | 100.8 | 0.2 | 0.2 |
| 06-07-93 | 98.7 | 100.8 | -2.1 | 2.1 |
| 06-07-93 | 99.4 | 100.8 | -1.4 | 1.4 |
| 06-23-93 | 113.5 | 108.1 | 5.0 | 5.0 |
| 07-07-93 | 116.5 | 108.1 | 7.8 | 7.8 |
| 07-21-93 | 107.4 | 108.1 | -0.6 | 0.6 |
| 08-04-93 | 107.1 | 108.1 | -0.9 | 0.9 |
| 08-18-93 | 111.4 | 108.1 | 3.1 | 3.1 |
| 09-02-93 | 113.5 | 108.1 | 5.0 | 5.0 |
| 09-23-93 | 114.0 | 108.1 | 5.5 | 5.5 |
| Average % Difference | | | -3.7 | 6.3 |
| Standard Deviation | | | 8.0 | 7.8 |
| Upper 95% Probability Limit | | | 11.9 | |
| Lower 95% Probability Limit | | | -19.3 | |

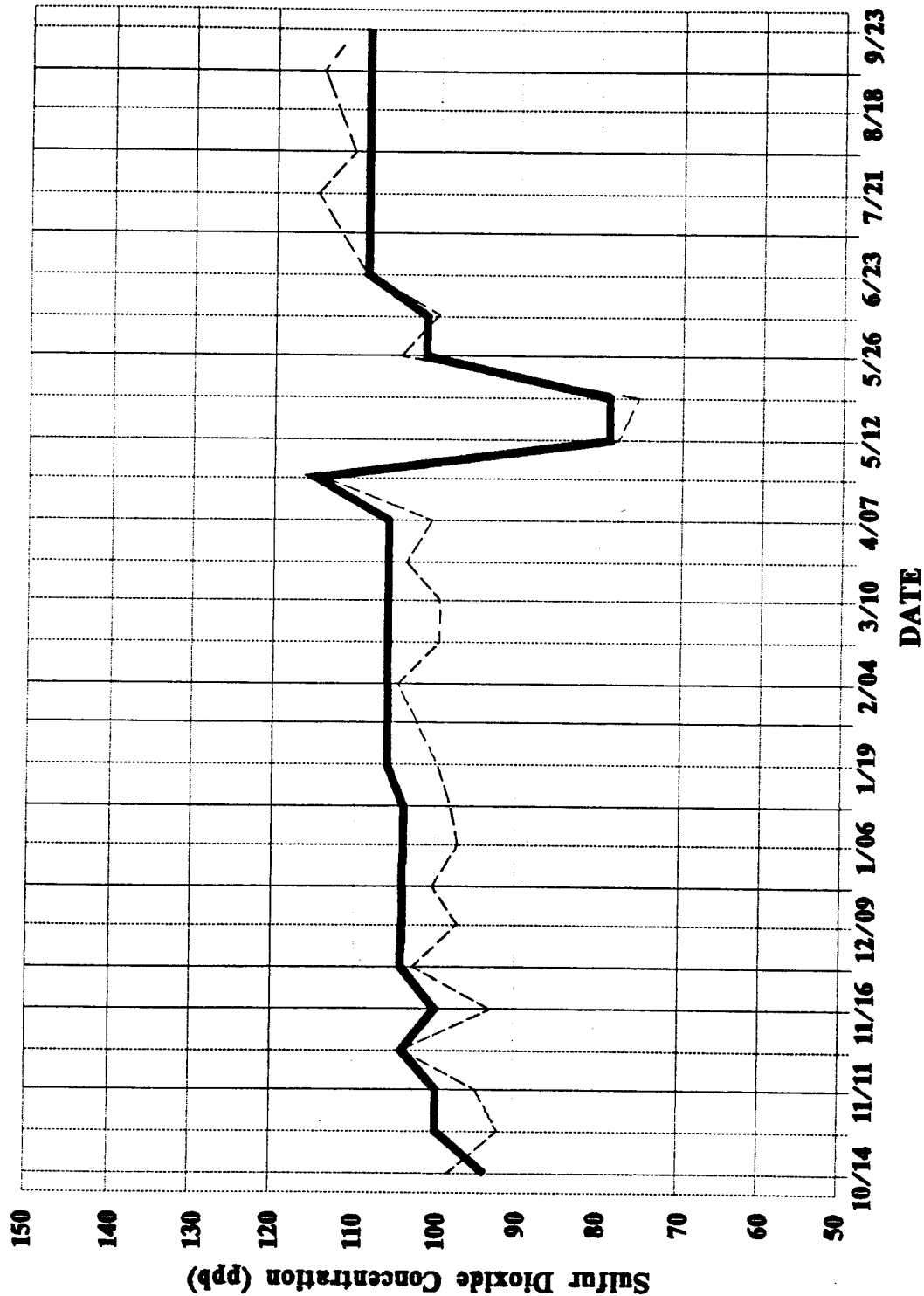
ppb - parts per billion
 % - Percent



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 7.3-1

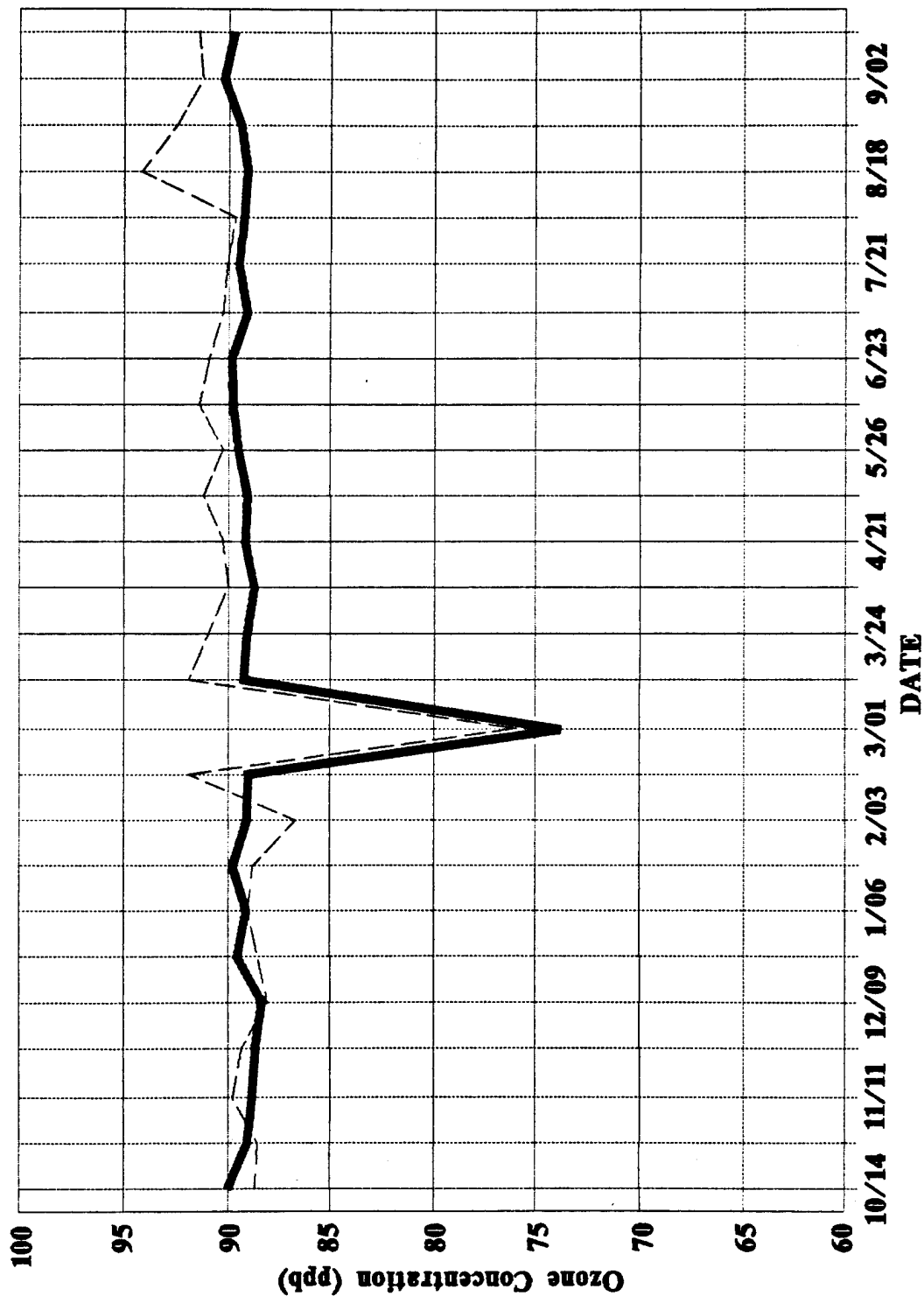
FY93 Carbon Monoxide
Concentration Precision Checks
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated



Prepared for:
 U.S. Army Program Manager
 for Rocky Mountain Arsenal

Figure 7.3-2

FY93 Sulfur Dioxide
 Concentration Precision Checks
 Rocky Mountain Arsenal
 Prepared by: Ebasco Services Incorporated



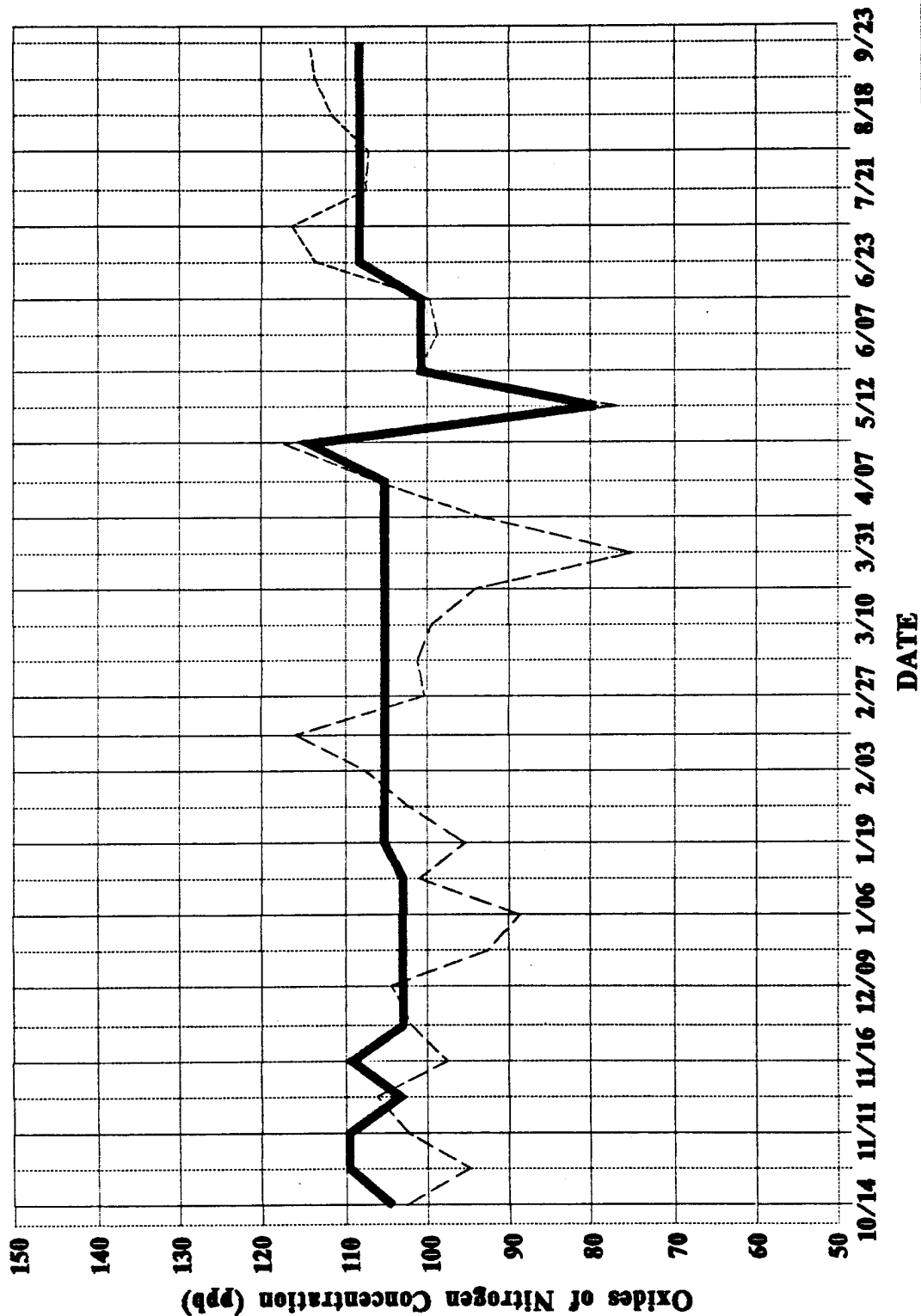
Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 7.3-3

FY93 Ozone Concentration
Precision Checks
Rocky Mountain Arsenal

Prepared by: Ebasco Services Incorporated

— CALIBRATOR OUTPUT --- ANALYZER RESPONSE



Prepared for:
U.S. Army Program Manager
for Rocky Mountain Arsenal

Figure 7.3-4

FY93 Oxides of Nitrogen
Concentration Precision Checks
Rocky Mountain Arsenal
Prepared by: Ebasco Services Incorporated

showed a percent difference greater than ± 7 percent. Field staff were informed of these discrepancies and they immediately performed the required corrective action. These samplers were then rechecked by the auditor as appropriate. Detailed results of all samplers, analyzers, and meteorological towers are provided in the CAQMMP Quarterly Audit Reports. In addition, these reports contain copies of all relevant certification paperwork that documents the traceability of the standards utilized in the CAQMMP.

The CAQMMP voluntarily participated in the NPAP during FY93. The accuracy of reported flow rates for 4 TSP samplers and 4 PM-10 samplers was assessed using EPA-provided test equipment. The CO, SO₂, and NO_x analyzers were checked once and the contract laboratory's lead analysis method was evaluated four times as well. Table 7.3-12 provides the results of these audits. All results were well within EPA prescribed tolerances.

7.3.5 Data Processing

A series of formal steps was implemented to ensure the quality of data generated under the CAQMMP. All field data sheets, calibration forms, and chain of custody forms were reviewed and double-checked by the field team. The project data management coordinator reviewed these files and provided PMRMA's contract data processing group (D.P. Associates) with calculated sample volumes. For each sample, the laboratory provided a raw weight for each target analyte. These values were entered into the IRDMIS system where method accuracy corrections were applied and sample concentrations calculated. Project chemists reviewed the associated laboratory quality control data; including surrogate and spiked recoveries and general compliance with PMRMA quality control methodology. The acceptability of each lot was addressed by the laboratory and reviewed along with control charts by the project QA team and PMRMA. Periodic checks of the IRDMIS air quality database were made by the project data management coordinator to verify sample identifications and concentration calculations.

Table 7.3-12 FY93 National Performance Audit Program

Page 1 of 1

| Type | Date | CAQMMP Reported Result | EPA Reported Result | % Difference | Comment |
|---------------------------------|----------|---------------------------|------------------------|-----------------|--------------|
| Lead (µg) | 1/15/93 | 700.00 | 703.00 | -0.4 | |
| | | 194.00 | 200.00 | -3.0 | |
| | 4/5/93 | 972.00 | 954.00 | 1.9 | |
| | | 96.80 | 100.00 | -3.2 | |
| | 7/9/93 | 847.00 | 869.00 | -2.5 | |
| | | 236.00 | 251.00 | -6.0 | |
| | 10/25/93 | 645.00 | 615.00 | 4.9 | |
| | | 153.00 | 148.00 | 3.4 | |
| Hi Vols (µg/m ³) | 2/5/93 | 1.126 | 1.129 | -0.3 | AQ1 PM-10 |
| | | 1.150 | 1.129 | 1.9 | AQ4 PM-10 |
| | | 1.130 | 1.148 | -1.6 | AQ1 TSP |
| | | 1.131 | 1.160 | -2.5 | AQ4 TSP |
| | 7/27/93 | 1.108 | 1.113 | -0.4 | SQ1 PM-10 |
| | | 1.172 | 1.185 | -1.1 | SQ2 PM-10 |
| | | 1.084 | 1.153 | -6.0 | SQ1 TSP |
| | | 1.170 | 1.180 | -0.8 | SQ2 TSP |
| Analyzers | 7/12/93 | | | | |
| CO | | | | 3.75 | Mean % Diff. |
| SO ₂ | | | | 0.55 | Mean % Diff. |
| NO ₂ | | | | 7.92 | Mean % Diff. |

PM-10 - Particulate matter less than 10 micrometers
 TSP - Total Suspended Particulates
 Diff. - Difference
 CO - Carbon monoxide
 SO₂ - Sulfur dioxide
 NO₂ - Nitrogen oxides
 % - Percent
 µg - Micrograms
 µg/m³ - Micrograms per cubic meter
 Diff. - Difference

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